

Environmental Science and Pollution Research

Potentially toxic elements in sediments near mines – a comprehensive approach for the assessment of pollution status and associated risk for the surface water environment

--Manuscript Draft--

Manuscript Number:	ESPR-D-23-21450	
Full Title:	Potentially toxic elements in sediments near mines – a comprehensive approach for the assessment of pollution status and associated risk for the surface water environment	
Article Type:	Research Article	
Keywords:	mining activities, toxic elements, magnetic susceptibility, contamination indices, Monte Carlo simulation, river sediments	
Corresponding Author:	Sanja Sakan, Ph.D Institute of Chemistry Technology and Metallurgy: Institut za hemiju tehnologiju i metalurgiju Belgrade, SERBIA	
Corresponding Author Secondary Information:		
Corresponding Author's Institution:	Institute of Chemistry Technology and Metallurgy: Institut za hemiju tehnologiju i metalurgiju	
Corresponding Author's Secondary Institution:		
First Author:	Sanja Sakan, Ph.D	
First Author Secondary Information:		
Order of Authors:	Sanja Sakan, Ph.D Aleksandra Mihajlidi-Zelić Nenad Sakan Stanislav Frančičković-Bilinski Igor Kodranov Dragana Đorđević	
Order of Authors Secondary Information:		
Funding Information:	Ministarstvo Prosvete, Nauke i Tehnološkog Razvoja (451-03-47/2023-01/200026)	Dr Sanja Sakan
Abstract:	<p>This research is focused on the assessment of the pollution status of river and lake sediments near Pb, Zn, and Cu mines and tailings in the most southeastern part of Serbia - Krajište area. High potentially toxic elements (PTEs) contents are detected in studied river sediments (up to 7892 mgkg⁻¹ for Zn, 3224 mgkg⁻¹ for Cu, 36790 mgkg⁻¹ for Pb, 64.2 mgkg⁻¹ for Cd, and 1444 mgkg⁻¹ for As). Given that the contents of the studied elements in most of the river sediments exceeded the background values, values prescribed by regulations of the Republic of Serbia, as well as probable effect concentration (PEL), it is possible to conclude that sediments were heavily polluted and that detrimental effects can be expected. Contamination indices including EF, CF, Igeo, Eri, RI, PLI and ATI were used to assess the degree of pollution by PTEs. The ecological risk assessment revealed that there is a significant risk observed for toxic elements (primarily Pb, Cu, Cd and As) at this moment. The highest contamination indices (EF, Igeo, CF, PLI, ATI) are mainly associated with historical and current mining activities. The Monte Carlo analysis based on the risk assessment indices was used to evaluate the uncertainty. The most pronounced toxic risk is found for the Pb, Cu, Cd and As which assessment was in the range of high and extremely high-risk</p>	

	probabilities. The obtained results suggest that levels of toxic elements pose a significant ecological risk to the surface water environment near Pb, Zn, and Cu mines in the Krajište area. The methodology applied in this paper could be very useful for other researchers dealing with the problem of environmental pollution by toxic elements.
Suggested Reviewers:	<p>Vladimir Dauvalter Kola Science Centre RAS: Federal'nyj issledovatel'skij centr Kol'skij naucnyj centr Rossijskoj akademii nauk vladimir@inep.ksc.ru</p> <p>Ahmad Jamshidi-Zanjani Tarbiat Modares University Faculty of Basic Sciences ajamshidi@modares.ac.ir</p> <p>Tomas Matys Grygar Institute of Inorganic Chemistry Czech Academy of Sciences: Ustav anorganicke chemie Akademie ved Ceske republiky grygar@iic.cas.cz</p>
Opposed Reviewers:	
Additional Information:	
Question	Response
§Are you submitting to a Special Issue?	No

06. October 2023.

Dear Editor,

We are pleased to submit an original research article entitled "**Potentially toxic elements in sediments near mines – a comprehensive approach for the assessment of pollution status and associated risk for the surface water environment**" by Sanja Sakan, Aleksandra Mihajlidi-Zelić, Nenad Sakan, Stanislav Frančišković-Bilinski, Igor Kodranov and Dragana Đorđević for consideration for publication in **Environmental Science and Pollution Research**.

We show the first results on the influence of mining activities on river sediments in Eastern Serbia, offering insight into the problem of Pb, Zn, As and Cu contamination in river systems. The impact of the mining activities in the Karamanica and Blagodot ore fields on the pollution status of river and lake sediment was estimated in this research. The main goal of this research was to evaluate the level and distribution of toxic elements (i.e., Pb, Cd, Cu, Zn, As, Mn, Ni and Cr), assess the contamination degree of sediments using contamination indices, (EF, CF, Igeo, Eri, RI, PLI and ATI) and evaluate the uncertainty associated with the risk assessment indices using Monte Carlo analysis. The methodology applied here, namely the combined use of pollution indices, geochemical background values, magnetic susceptibility, statistical analysis, and Monte Carlo analysis, gives a suitable information about contamination levels, sources of PTEs, as well as the current and future potential risks of mining activities to the downstream environments. Those methods have been used separately earlier, but their combined use is proven to be a more valuable tool for the analysis of the situation since if the several independent methods point to same result, the certainty of the obtained data is more pronounced.

We hope that our manuscript meets criteria to be published in **Environmental Science and Pollution Research**, since it deals with the ecological risk assessment and evaluation of the pollution status of surface waters, especially taking into account that the problem of surface water pollution is growing globally and that the lack of clean drinking water has been observed all over the world.

All co-authors agree with the contents of the manuscript, everyone meriting authorship in this work has been named above, and there is no conflict of interest involving this study. This manuscript has not been published before and it is not under consideration for publication anywhere else. Its publication has been approved by all co-authors. I have not submitted my manuscript to a preprint server before submitting it to **Environmental Science and Pollution Research**.

Sincerely,

Dr Sanja M. Sakan (ssakan@chem.bg.ac.rs, sanja.sakan@ihtm.bg.ac.rs)

Centre of Excellence in Environmental Chemistry and Engineering – Institute of Chemistry, Technology and Metallurgy, National Institute of the Republic of Serbia, University of Belgrade, Njegoševa 12, 11158 Belgrade, Serbia

Click here to view linked References

Potentially toxic elements in sediments near mines – a comprehensive approach for the assessment of pollution status and associated risk for the surface water environment

Sanja Sakan^{1*}, Aleksandra Mihajlidi-Zelić¹, Nenad Sakan², Stanislav Frančičković-Bilinski³, Igor Kodranov⁴, Dragana Đorđević¹

¹ Centre of Excellence in Environmental Chemistry and Engineering – Institute of Chemistry, Technology and Metallurgy, National Institute of the Republic of Serbia, University of Belgrade, Njegoševa 12, 11158 Belgrade, Serbia

² Institute of Physics, National Institute of the Republic of Serbia, University of Belgrade, Pregrevica 118, 11000 Belgrade, Serbia

³ Ruđer Bošković Institute, Division for marine and environmental research, Bijenička cesta 54, 10000 Zagreb, Croatia

⁴ Faculty of Chemistry, University of Belgrade, Studentski trg 12-16, 11000 Belgrade, Serbia

*Corresponding author

ssakan@chem.bg.ac.rs; sanja.sakan@ihtm.bg.ac.rs

Telephone: + 381 11 3336 801

Fax: + 381 11 2636 061

Abstract

This research is focused on the assessment of the pollution status of river and lake sediments near Pb, Zn, and Cu mines and tailings in the most southeastern part of Serbia - Krajište area. High potentially toxic elements (PTEs) contents are detected in studied river sediments (up to 7892 mgkg⁻¹ for Zn, 3224 mgkg⁻¹ for Cu, 36790 mgkg⁻¹ for Pb, 64.2 mgkg⁻¹ for Cd, and 1444 mgkg⁻¹ for As). Given that the contents of the studied elements in most of the river sediments exceeded the background values, values prescribed by regulations of the Republic of Serbia, as well as probable effect concentration (PEL), it is possible to conclude that sediments were heavily polluted and that detrimental effects can be expected. Contamination indices including EF, CF, Igeo, Eri, RI, PLI and ATI were used to assess the degree of pollution by PTEs. The ecological risk assessment revealed that there is a significant risk observed for toxic elements (primarily Pb, Cu, Cd and As) at this moment. The highest contamination indices (EF, Igeo, CF, PLI, ATI) are mainly associated with historical and current mining activities. The Monte Carlo analysis based on the risk assessment indices was used to evaluate the uncertainty. The most pronounced toxic risk is found for the Pb, Cu, Cd and As which assessment was in the range of high and extremely high-risk probabilities. The obtained results suggest that levels of toxic elements pose a significant ecological risk to the surface water environment near Pb, Zn, and Cu mines in the Krajište area. The methodology applied in this paper could be very useful for other researchers dealing with the problem of environmental pollution by toxic elements.

Keywords: mining activities, toxic elements, magnetic susceptibility, contamination indices, Monte Carlo simulation, river sediments

Introduction

The dynamics of the behaviour of PTEs in freshwater environments have been one of the main focuses of environmental monitoring in the last years. High toxicity and nonbiodegradability of PTEs in the environment have raised extensive concern (Chen et al., 2019), and the presence of these elements in river systems is considered an indicator of anthropogenic influence. PTEs represent significant hazard to aquatic environment because anthropogenic sources are globally widespread and there is a fast increase in their number (Ustaoğlu et al., 2020). Natural processes, such as rock weathering can also emit toxic elements, but their contribution is small (Haris et al., 2017). One of the significant sources of environmental pollution by toxic elements are mining activities. Mine tailings, the waste material leftover after the extraction of target minerals from the ore material, frequently contain high contents of PTEs (García-Giménez and Jiménez-Ballesta, 2017). Various researchers have studied the impact of active and closed mines on aquatic ecosystems (Ayari et al., 2023; Barago et al., 2023; Lidman et al., 2023; Maftai et al., 2014; Shen et al., 2019; Shikazono et al., 2008; Yucel et al., 2018). The results of these studies indicate that all processes related to mining activities during exploitation, but also after their closure caused significant water contamination.

58 PTEs in the environment can have deleterious effects on the health of living biota through the food chain and
59 water supply (Ayari et al., 2023). Sediments have a high retention capacity for PTEs which can be subsequently
60 released into water, therefore many studies have investigated PTE levels in sediments and surface water (Haris et
61 al., 2017). Given that PTEs are very toxic, a mere comparing their sediment content with the geochemical
62 background values may not give the appropriate information on PTEs' adverse impacts on the ecosystems
63 (Ustaoğlu et al., 2020). For that reason, many other additional methods are used to assess the degree of pollution
64 in river and lake systems. According to Kowalska et al. (2018), pollution indices are crucial for the proper
65 assessment of toxic elements contamination of soil and sediment. The most widely used indices are Contamination
66 factor (CF), Enrichment factor (EF), Index of geoaccumulation (Igeo), Potential ecological risk index (Eri),
67 Ecological risk index (RI), and Pollution load index (PLI) (Ayari et al., 2021; Gantayat et al., 2023; Haris et al.
68 2017; Pujiwati et al., 2022). One of the relatively new indices, known as the Aggregative toxicity index (ATI),
69 proved to be very significant in assessing the degree of contamination. Jamshidi-Zanjani and Saeedi (2017)
70 developed this index and ATI was also applied in the Sakan et al. (2023). Furthermore, more accurate information
71 on contamination levels and sources can be provided when pollution indices are applied along with geochemical
72 background values (Maftei et al., 2014).

73 In addition to chemical analyses to determine levels of toxic elements and subsequent processing of data to assess
74 the degree of pollution in soil and sediment, magnetic susceptibility (MS) is also used as a proxy for chemical
75 methods for the assessment of soil and sediment pollution. This method is rapid, simple and non-destructive.
76 Pollutants and magnetic particles are usually associated, enabling the use of magnetic susceptibility as an indicator
77 of pollution; however, the intricacy of these relationships makes it hard to construct a single function to compute
78 pollutant concentrations from MS (Paradelo et al., 2009). Combined use of magnetic susceptibility and pollution
79 indices (Duodu et al., 2016) is a valuable tool for the assessment of sediment pollution status (Chaparro et al.,
80 2020).

81 The importance of more precise determining of the uncertainties has led us to apply a Monte Carlo simulation, as
82 it was seen in a number of manuscripts (Fakhri et al., 2018). Based on the input, as it was done in manuscript by
83 Fakhri et al. (2018), data health risks has been defined as a mean of variability in health risk assessments and
84 quantifying uncertainty, while the Monte Carlo simulation is applied to perform the probabilistic risk assessment.
85 Mine tailings represent a significant risk to the environment even after mining activities have terminated because
86 of the presence of toxic elements (García-Giménez and Jiménez-Ballesta, 2017). The environment in Serbia, due
87 to the long history of mining on its territory, is also affected by this threat. Taking that into account, the content
88 of several toxic elements and their geochemical impact on the river and lake pollution status in the Krajište area
89 (southeastern Serbia), near mines and their tailings were studied in this research. Namely, the impact of the mine
90 tailings located in the Karamanica ore field with deposits of Pb, Zn and Cu ore, and the Blagodot ore field
91 containing Pb and Zn ore was investigated.

92 Although negative effects of the flotation tailings in the Blagodot ore field, near the village of Kriva Feja, on the
93 streams and soil in the vicinity of flotation tailings have been reported (Đokić et al., 2013; Đokić et al., 2012;
94 Đorđević et al., 2012), data regarding the influence of the Musulj-Kekerinci dump in the Blagodot ore field on the
95 surrounding rivers are extremely scarce, and the environmental impact of mining activities in the Karamanica ore
96 field has not been investigated yet. For that reason, the content of several toxic elements in river sediments near
97 the Musulj-Kekerinci dump (Blagodot ore field) and the flotation tailings pond in the Karamanica ore field and
98 their geochemical impact on the river and lake pollution status were investigated.

99 The specific objectives of this study were: (1) to evaluate the level and distribution of toxic elements (i.e., Pb, Cd,
100 Cu, Zn, As, Mn, Ni and Cr) content in sediment samples taken from the rivers and lakes near the area of mines
101 and their tailings; (2) to assess the sediment contamination level using contamination indices, including EF, CF,
102 Igeo, Eri, RI, PLI and ATI; 3) application of magnetic susceptibility for assessing the PTE pollution levels; (4)
103 evaluating the uncertainty associated with the risk assessment indices using Monte Carlo analysis; and (5) to
104 differentiate between natural and anthropogenic sources of contamination. This kind of research has not
105 previously been conducted in this area. For detailed research, in addition to chemical methods, magnetic
106 measurements are also included, as well as the application of statistical methods, including Monte Carlo
107 simulation. We expect that the obtained results will be of great importance for the assessment of the ecochemical
108 status of the examined locality and in showing the influence of (old) tailings and pilot plants on pollution with
109 toxic elements, which is also of great importance due to the proximity of the border with Bulgaria and the
110 possibility of transboundary pollution. We believe that the combination of applied methods will be extremely
111 valuable for other researchers dealing with the issue of pollution with toxic elements not exclusively related to
112 metal mining.

114 **Materials and methods**

116 **Study area and sampling**

117 In the period from August 13 to 14, 2020. samples of sediments were collected in surface waters in the Krajište
118 area in southeastern Serbia (Fig. 1). GPS coordinates and the names of the localities are shown in Table 1. Eight
119 samples of river sediment and two samples of lake sediment were taken during sampling. Sediment samples were
120 taken with a plastic spatula.

121 The Karamanica ore field is located in the southeasternmost part of Serbia, on the southern slopes of the Bele
122 Vode mountain (1829 m a.s.l.), southwest of the town of Bosilegrad and close to the border with Bulgaria and
123 North Macedonia. The mine exploits Podvirovi and Popovica deposits of lead, zinc and copper ore located near
124 Karamanica village. Popovska and Karamanička rivers, tributaries of the Golema river, and the Nameless stream
125 that flows into the Karamanička river, stretch through the exploitation field. Additionally, occasional torrential
126 tributaries of Popovska and Karamanička rivers flow through the exploitation field. The Golema river is a tributary
127 of the Brankovačka river, which empties into the transboundary Dragovištica river. Dragovištica river in Bulgaria
128 joins the Struma river, eventually draining into the Aegean Sea. The location of planned mines of the Karamanica
129 ore field includes a part of the ecologically important area of the Ecological Network of the Republic of Serbia -
130 Prime Butterfly Area – PBA. To test flotation technology for processing lead, zinc and copper ore from the
131 Podvirovi and Popovica deposits, a pilot plant was built (Request for EIAS, 2021). The pilot facility, which
132 includes the flotation tailings pond, is located less than 5km from the borders of Serbia with Bulgaria and North
133 Macedonia, and was operating at the time of sampling. The Nameless stream that flows into the Karamanička
134 river flows alongside the flotation tailings pond. The sediments analyzed in this manuscript were taken from The
135 Nameless stream that flows into the Karamanička river (two samples, upstream and downstream of the flotation
136 tailing dump of the Podvirovi ore deposit pilot plant), as well as from Golema and Brankovačka rivers.

137 Grot mine (EX Blagodat mine) is a lead-zinc mine located in the Blagodat ore field, on the southeast side of the
138 mountain Besna Kobila, in the southeastern part of Serbia, near the village of Kriva Feja and around 20 km from
139 the state borders with Bulgaria and North Macedonia. From the western slopes, the waters flow into the
140 Barisavička and Jelašnička rivers towards the South Morava river and the Danube, which empties into the Black
141 Sea. From the eastern slopes, the Musuljska and Crna rivers, tributaries of the Ljubatska river, together with
142 Ljubatska river flow into the transboundary river Dragovištica, a tributary of the Struma river, which flows into
143 the Aegean Sea. A part of the Ljubatska river is directed into the Lisina lake, a reservoir built on the confluence
144 of the Lisinska and Božička rivers, and from there the water is pumped through the Božica channel into the Vlasina
145 lake, a reservoir formed on the Vlasina river. The reservoirs of the Vlasina and Lisina lakes together with the
146 system of channels and tunnels and four hydropower plants represent the hydropower system Vlasina. Vlasina
147 lake is a protected natural asset (category I) of exceptional importance. In addition, the area of Vlasina is included
148 in the internationally significant areas for birds - IBA (Important Bird Areas), for plant species - IPA (Important
149 Plant Areas), and, as a wetland of international importance, in the list of Ramsar areas.

150 The Blagodat ore field includes Blagodat, Đavolja Vodenica, Vučkovo II, and Kula deposits
151 (<https://www.mindat.org/loc-62233.html>). The first commercial research and ore exploitation in these areas was
152 organised by the Italian Mining Company "Societe Comercial d'Oriente" in 1903 (Stojmenović et al., 2017). More
153 recently, geological, geochemical and geophysical investigation of the Blagodat ore field in the 60s and 70s
154 resulted in a mine opening, which is still active. The tailings that accompanied mining activities in the Blagodat
155 ore field are located in two places. The oldest one, occupying 6,628 m², in the vicinity of which river sediment
156 samples were taken, is located in Musulj - Kekerinci Mahala near the confluence of Krasnodolska river and Crna
157 river, where tailing material deposited more than a century ago, together with recent mine waste material,
158 contaminates Crna river and still prevents the development of plants in this location (Đokić, 2012; Đokić et al.,
159 2012; Đokić and Jovanović, 2008). In this manuscript, samples were taken from Krasnodolska river, Crna river
160 (two samples, upriver and downriver from the confluence with Krasnodolska river) and Ljubatska river (before
161 the water intake for Lisina lake), as well as samples from Lisina and Vlasina lake. Sediment samples of the
162 Krasnodolska and Crna rivers were taken in the vicinity of their confluence, near the Musulj - Kekerinci dump.

163

164 **Research methodology**

165

166 **Sample analysis**

167

168 After taking sediment samples in the field and arriving at the laboratory, samples were air-dried for 8 days before
169 analysis (Arain et al., 2008). Analysis of dried sediments was performed using BCR sequential extraction (Sakan
170 et al., 2016; Sutherland 2010). The procedure is described in more detail in the Supplementary material. After
171 obtaining the extracts, the content of studied elements was determined using Thermo Scientific ICP-OES iCap
172 6500 Duo and Thermo Scientific ICP-MS iCap Q.

173 In this research, the contents of the following elements were determined in each sample: Pb, Cd, Cu, Zn, As, Mn,
174 Ni, Cr, Al, B, Ba, Ca, Co, Fe, K, Li, Mg, Na, P, S, Sr, and V. The results are expressed in mg kg⁻¹ dry sediment.

175 Within this manuscript, the presented results refer to the total contents of elements and the calculation of
176 contamination factors and the assessment of the degree of pollution was performed using methods related to the

177 total content of elements. The total content is represented by the sum of the extracted content of each element
178 individually during the three steps of BCR sequential extraction and digestion using aqua regia, which represents
179 an additional fourth extraction step that is not part of the standard BCR extraction procedure.

180

181 **Magnetic susceptibility (MS)**

182

183 The magnetic susceptibility was measured in the dried sample using a magnetic susceptibility meter SM30 (Sakan
184 et al. 2022) and the mean value of three consecutive measurements was taken as a final result.

185

186 **Toxic element pollution indices**

187

188 For the quantification of pollution level of studied sediments the following contamination indices were calculated:
189 Enrichment factor (EF), Contamination factor (CF), Index of geoaccumulation (Igeo), Ecological risk factor (Eri),
190 Ecological risk index (RI), Pollution load index (PLI) and Aggregative toxicity index (ATI). Definitions and
191 detailed explanation of the calculation of pollution indices are presented in the Supplementary material. As
192 geochemical background contents the 75th percentiles of the frequency distribution of the data were used (Dos
193 Santos and Alleoni, 2013).

194

195 **Statistical analyses**

196

197 Descriptive statistical parameters, as well as correlation analysis, were applied to investigate the distribution of
198 studied elements. For statistical analyses, IBM SPSS 21 software was used.

199

200 **Monte Carlo simulation**

201

202 A Monte Carlo analysis based on mathematical statistics and probability theory was used to assess model
203 uncertainty by random sampling of a probability distribution for each variable (Qu et al., 2018). In this manuscript,
204 the Monte Carlo simulation was applied to estimate the distribution of the toxic element content (or Eri and HRI)
205 in river sediments in order to determine a probability distribution (i.e., uncertainty) for the assessment metrics
206 with a purpose of evaluation of uncertainty associated with the risk of toxic elements in the studied area.

207

208 **Results and discussion**

209

210 **Performance of the analytical procedure**

211

212 The accuracy of the obtained results was checked by analysing sediment reference material (BCR-701) for three-
213 step sequential extraction. The average recovery values for heavy metals in the standard reference material were
214 in the range of 77.2–121.2%. The relative standard deviations of the means of duplicate measurement were less
215 than 10%. Expanded uncertainty of the measurement with coverage factor 2 was max. 6%.

216

217 **Element content in studied river sediments**

218

219 The descriptive statistics of 22 elements are shown in Table 2. The spatial distribution of the content of selected
220 elements, some of which are very significant due to their toxicity, is shown in Figs. 2a and 2b.

221 Distribution of the element's contents by locality indicates that in locality 4 (Brankovačka river), increased content
222 of barium was observed compared to other localities, and that the contents of Cr, V, and Sr in sediment are also
223 increased at this locality. More depicted in Fig. 2b, sediment collected at locality 6 (Crna river) contains a much
224 higher content of Pb than sediments in other localities.

225 For the preliminary risk assessment, the total content of the studied elements was compared with the: 1) limit and
226 remediation values prescribed by the Regulation of the Government of the Republic of Serbia (2012); 2) threshold
227 effect concentration (TEC) and probable effect level (PEL); 3) element concentrations in the continental crust
228 (Wedepohl, 1995) and background values, and 4) as well as with data for river sediments in the world. The TEC
229 and PEL are defined in MacDonald et al. (2000).

230 The obtained results show that the studied river sediments contain very high levels of PTEs. The total content of
231 elements were up to 7892 mgkg⁻¹ for Zn, 3224 mgkg⁻¹ for Cu, 36790 mgkg⁻¹ for Pb, 64.2 mgkg⁻¹ for Cd, and 1444
232 mgkg⁻¹ for As. The mean values of Zn, Cu, Pb, Cd, and As contents are higher than the maximum allowable values
233 (MAV) and remediation value (RV) (Government of the Republic of Serbia, 2012). The Ni content is similar to
234 the MAV value, while the Cr content is lower, which indicates the geochemical origin of these elements. The
235 contents of As, Cd, Cu, and Pb in river sediment were characterised by high coefficient of variation values (CV ≥
236 100 %) (Table 2; Fig. 2). The highest value of the coefficient of variation was observed for Pb, followed by As

237 and Cu, while the CV for Cd has a value of 100 %. The mean values of these PTEs, as well as Zn were much
238 higher than the geochemical background and element content in the Continental Crust (Tables 2 and 3). The
239 content of Ni and Cr is slightly higher compared to the estimated background value (Table 3), which may be a
240 consequence of the specific geochemistry of the investigated localities. The levels of As, Cd, Cu, Pb, Zn, and Ni
241 exceeded PEL, which indicates that these elements pose a potential risk to aquatic life in studied rivers. Ni
242 exceeded the TEC levels and the Cr content is similar to the TEC level, so it is possible to conclude that
243 investigated river systems are not polluted with these elements.

244 The results of comparisons of the content of other elements with the crust values show that the content of Al, Ba,
245 Ca, Co, Li, Mg, Na, Sr, and V (mean value) is lower, while the content of B, Fe, K and Mn larger than the crust
246 values. Slightly higher content of these elements in relation to crust values is associated with specific geology in
247 certain localities. It is known that manganese sulfides are present to a greater or lesser extent along with sulphides
248 of lead, copper, and cadmium. The increased content of Ca and Co indicates an increased content of carbonates
249 in the mentioned localities, i.e. specific rocks with a higher cobalt content.

250 Results of comparison with other studies have shown that the mean content of Pb, Cd, Zn, Cu, and As in studied
251 river sediments were similar to data for the areas worldwide affected by the mining activities (Table 3). The mean
252 values of Cu, Pb, Cd, and As contents in the surveyed river sediment on the territory of Krajište are higher
253 compared to the values listed in Table 3 for the river sediment located near the mine area in Tunisia (Ayari et al.,
254 2023), Romania (Maftei et al. 2014), Japan (Shikazono et al. 2008), UK (Wolfenden and Lewin, 1978), Spain
255 (Oyarzun et al. 2011) and China (Shen et al., 2019). The results for stream sediment, in terms of Zn and Cd
256 contents, in Guizhou province, China (abandoned Zn–Pb smelter, Hezhang Country), which were presented in the
257 paper by Yang et al. (2010), are higher than the values obtained in our research, but the content of Pb is higher
258 than in Guizhou province, even though these sediments are affected by the abandoned Zn–Pb smelter. It is possible
259 to conclude that water originating from the area near the flotation tailings and pilot plant contains high
260 concentrations of toxic elements.

261

262 **Magnetic susceptibility (MS)**

263

264 The values of magnetic susceptibility and results of the determination of anomalies (extremes and outliers) by the
265 Box plot method are shown in Table S1. Correlations between MS and elements contents are presented in Table
266 S2.

267 MS shows a very high positive correlation with Pb and P. Correlations between magnetic susceptibility and Pb
268 are also shown in the literature (Karimi et al., 2011). Desai et al. (1989) reported some aspects of the magnetic
269 susceptibility of ferroelectric magnesium hydrogen phosphate crystals, which may explain the existence of a
270 relationship between phosphate and magnetic susceptibility in the investigated area.

271 Despite the following correlations are not statistically significant, correlations with Co, Fe and K should also be
272 mentioned. The existence of a positive, but weak correlation of MS with Fe and K indicates that magnetic
273 susceptibility is influenced by the high contents of iron-bearing ferromagnetic minerals, but also clays minerals
274 (Putra et al., 2019).

275 Anomalies of studied elements and magnetic susceptibility (MS) have been determined, using a boxplot statistical
276 approach. Elements with anomalies are: Cr, Pb, Al, Ba, Co, P, Sr, as well as MS.

277 When looking at cases (sampling locations), it was found that the most anomalous location is 6 (Crna river, upriver
278 from the confluence with Krasnodolska river), with four extremes: Cr, Pb, P and MS. As according to correlation
279 analysis, MS is in very good correlation with Pb and P, also in a weak correlation with Cr, it could be assumed
280 that they form an association, probably associated with some ore occurrence.

281 The second anomalous location is 4 (Brankovačka river), with extremes of Cr, Ba and Sr and an outlier of Al.
282 This could be due to the local geological composition of surrounding rocks.

283 Two more locations show one anomaly: 5 (Krasnodolska river) an extreme of P, and 1 (The Nameless stream)
284 one outlier of Co, which is the result of the specific geological composition of the rocks in the aforementioned
285 localities.

286 Classification values of topsoil magnetic susceptibility (Magiera, 2022) are presented in Table S3. The obtained
287 MS values (Table S1) of the examined sediment samples were compared with the MS classification values for the
288 soil (Table S3). The value of MS in sample 6 (Crna river) is greater than 1×10^{-3} SI units, which indicates that at
289 least one element concentration is above the threshold value. At this site, the Pb content is extremely high and
290 many of the contamination factors calculated for this site are extremely high.

291

292 **Contamination indices**

293

294 The grades of contamination and ecological risk based on the calculated factors (i.e. EF, CF, Igeo, Eri, RI, PLI,
295 ATI) are shown in Tables S4 and S5.

296

297 **Enrichment factor (EF)**

298

299 Calculated enrichment factors (EF) for analyzed sediments are presented in Fig. 3a. The EF values were
300 interpreted according to Acevedo-Figueroa (2006). EF values are calculated using Al as a normalizer. Being the
301 main component of fine-grained aluminosilicate minerals, and as such, a good surrogate for clay particles (Donkor
302 et al., 2005), Al has been the most frequently used tracer element in normalization calculations/models. As
303 background values, 75 percentile of the frequency distribution (Dos Santos and Alleoni, 2013) of the element
304 content in sediments from the Vlasina river watershed are used (Table 3). The reason for this choice of values for
305 the background content is that the sediments of the rivers in the Vlasina region have been shown to be
306 uncontaminated by toxic elements (Sakan et al., 2023; Sakan et al., 2022).

307 As shown in Figure 3a and Table S4, the sediment samples were extremely severely enriched with Zn, Cu, Pb,
308 Cd and As, moderately severely enriched in Mn and minor enrichment was observed for Ni and Cr. The highest
309 EF values were calculated in sediments at site 6 (Crna river) which was sampled upriver from the confluence with
310 Krasnodolska river, close to the tailing dump.

311

312 **Contamination factor (CF)**

313

314 The calculated results for the contamination factors (CF) are given in Fig. 3b. In accordance with the grades of
315 contamination factor (Table S5), CF is classified into four groups in Pekey et al. (2004) and Hakanson (1980).
316 This investigation showed that CF values for studied elements ranged from 0.87 to 1774 (low to very high
317 contamination). According to the analysis of the data, the order of mean CF values in sediment samples, indicating
318 the degree of sediment contamination by the individual element, is the following: Pb >> Cd > Cu > As > Mn > Cr
319 > Ni. Again, the highest CF values were calculated for sediments at site 6 (Crna river).

320

321 **Index of geoaccumulation (Igeo)**

322

323 The calculated Igeo values are shown in Figure 4a. Müller (1979) proposed seven grades or classes of pollution
324 level (Table S4). The calculated results of Igeo indicated that investigated sediments were uncontaminated to
325 strongly contaminated by Pb; uncontaminated to moderately contaminated with respect to Zn, Cu, Cd and As, and
326 that the degree of contamination of sediments by Ni, Cr and Mn was in the range of uncontaminated to
327 uncontaminated to moderately contaminated.

328 According to the mean values of Igeo, the pollution of studied sediments by elements is decreasing in the following
329 order: Pb > Zn > Mn > Cd > Cu > As > Cr > Ni, which is consistent with the calculated EF values.

330

331 **Ecological risk factor (Eri)**

332

333 The results of the evaluation of potential risk factor (Eri) are summarized in Figure 4b. The Eri for studied
334 elements and the associated ecological risk were as follows: Zn (1 - 106) – low to considerable; Ni (4.30- 11.13)
335 – low; Cu (4.34- 661) - low to very high; Cr (2.98 – 9.21) – low; Pb (3.62 - 8837) - low to very high; Cd (3.91-
336 4189) - low to very high; and As (3.78 - 1301) - low to very high ecological risk. It was observed that the values
337 of potential risk factor correspond to low to very high ecological risk (Cu, Pb, Cd, and As), considerable (Zn), and
338 low ecological risk (for Ni and Cr). The mean Eri values for studied elements in descending order were: Cd > Pb
339 > As > Cu > Zn > Ni > Cr. It should be emphasized that the high mean value of Eri for Cd compared to Pb and
340 other elements is a consequence of the fact that the potential ecological risk factor of a single toxic element is
341 calculated from the perspective of the biological toxicity of elements.

342

343 **Ecological risk index (RI)**

344

345 The RI represent the sum of individual elements risk factors. The calculated values of RI (Tables S5 and S6, Fig.
346 S1) are in the range from 24.40 to 9637, indicating low to very high ecological risk in the investigated area (mean
347 value is 3971). The results showed that there was low potential ecological risk for the sediment sample collected
348 from Vlasina lake, and higher ecological risk at other localities. Fig. S2 shows the contribution of different PTEs
349 to the average potential ecological risk index (RI). The contribution of each element to the average RI is the
350 following: Cd (42%) = Pb (42%) > As (9%) > Cu (6%) > Zn (1%), while the contribution of Ni and Cr is 0%. As
351 well as for EF and CF, the highest RI values were calculated for sediments at site 6 (Crna river).

352

353 **Pollution load index (PLI)**

354

355 The results of the PLI are presented in Table S6 and Fig. S1. The value PLI > 1 indicates the existence of pollution;
356 otherwise, if the PLI < 1, there is no metal pollution (Tomlinson et al., 1980). In this study, the average PLI value

357 for sediment samples was 13.78, ranging from 0.65 to 27.81, which indicated no pollution for Vlasina lake, and
358 elevated pollution level for all other localities: high ecological risk (1,2 - the Nameless stream emptying into
359 Karamanička river, 3 - Golema river, 5 -Krasnodolska, 6, 7 - Crna and 8 - Ljubatska rivers), considerable (9 -
360 Lisina lake) and moderate ecological risk (4 - Brankovačka river).

361

362 **Aggregative toxicity index (ATI)**

363

364 In Tables S6 and Fig. S1 the results of ATI values for studied localities are shown. The parameters used to calculate
365 the aggregative toxicity index are presented in Table S7. The proposed classification of ATI is as follows: $ATI < 0.1$
366 indicates no toxic level; $0.1 \leq ATI < 0.3$ reveals a low toxic degree; $0.3 \leq ATI < 0.5$ signifies a medium toxic
367 degree; $0.5 \leq ATI < 1$ points to a high toxic degree, and $ATI > 1$ indicates an extremely high toxic degree
368 (Jamshidi-Zanjani and Saeedi, 2017).

369 The highest values of the ATI parameter were observed for Crna river - site 6 (20.384), followed by the Nameless
370 stream emptying into Karamanička river – sites 1 and 2, and Golema river – site 3 (about 10), and sites 7 and 5
371 (Krasnodolska and Crna rivers). These values of the ATI factor indicate an extremely high toxic degree. A high
372 toxic degree ($0.5 \leq ATI < 1$) was observed at localities 8 and 9 (Ljubatska river and Lisina lake), a medium toxic
373 degree was registered for Brankovačka river (site 4), and for Vlasina lake (site 10) calculated ATI value indicated
374 a low toxic degree.

375

376 **Conclusion for contamination factors**

377

378 Taking into account the individual contamination factors, Pb is the biggest contributor to pollution in most
379 localities, followed by Cu, Cd and As, whose values are usually extremely severe, and sometimes moderately
380 severe. Also, there is significant pollution by Zn, followed by Mn, and low contamination with respect to Ni and
381 Cr was observed. Based on this result, it can be concluded that there is a high probability that Pb could cause harm
382 to the environment in the area. Of all calculated contamination factors, only the mean value of factor Eri is higher
383 for Cd compared to Pb. Unlike Igeo, EF and CF, which are based on environmental geochemistry, and their values
384 show the degree of heavy metals pollution by anthropogenic activities, the calculation of the potential ecological
385 risk factor of a single toxic element takes into account the toxicity of elements. Cd toxicity coefficient is 3 to 30
386 times that of other elements, and even low concentration of Cd are capable of causing severe health problems.
387 Contamination factors for As are higher in sites 1, 2, and 3, which are related to the Karamanica ore field, due to
388 the higher content of arsenic in ores from this ore field (Milošević et al., 2015). Based on the obtained results, it
389 can be concluded that the highest values of the contamination factors were observed in the river sediment at site
390 6, i.e. Crna river, upriver from the confluence with Krasnodolska river (in the vicinity of Musulj – Kekerinci dump
391 located in Blagodot ore field).

392

393 **Correlation analysis**

394

395 Pearson's correlation analysis was performed for the contents of the examined elements (Table S8). In our study,
396 a significant positive correlation is observed between most of the studied elements. Positive correlations were
397 found between Zn and Cd, Ca, S, Mn; Cu and Mn, S, As; Cr and Fe, Al, V; Cd and Zn, Ca, S, Mn; Ca and Zn, S,
398 Cd; Fe and Cr, Al; Mn and Zn, Cu, S, Cd; and As and Cu, S; also a negative correlation exists between As and Li.
399 Obtained results confirm that hydroxides of iron play a role in the distribution of Cr and that manganese
400 hydroxides are significant for Zn, Cu, and Cd bounding. Positive correlation of Mn with Zn, Cu, S and Cd can
401 also point to the common origin of these elements, i.e. the source of these elements that is associated with mining
402 activities. Sulfides of cadmium, manganese, arsenic, antimony, nickel and cobalt are present in the composition
403 of lead-zinc or lead-zinc copper ores of the sulphide type to a greater or lesser extent (Stanojević and Filipović-
404 Petrović, 2014). Since iron is abundant in the Earth's crust, and aluminium mainly originates from aluminosilicate
405 minerals, positive correlations of Cr with these two elements indicate its natural, i.e. geochemical origin. Positive
406 correlations of Zn and Cd with Ca may indicate the existence of carbonate-hosted Zn ores. The metamorphic
407 (schist) complex which dominates both Blagodot and Karamanica ore fields includes carbonate horizons
408 composed of marbles and calcschists, metallogenically favourable for deposition of metasomatic mineralization
409 type. Ore deposits were formed by metasomatic reactions of marbles and calcschists induced by hydrothermal
410 solutions of Pb, Zn and Cu (Simić, 2001; Simić, 2002; Đokić, 2012).

411 The existence of a positive correlation between phosphorus and lead was shown in this paper. This correlation
412 may indicate the presence of some Pb phosphates in the area. Phosphorus (P) in the river sediment can influence
413 the fate and transport of heavy metals (Shen et al., 2020). In soils phosphorus can be associated with Al, Fe, or
414 Ca (Werner and Prietzel, 2015). It was already reported that calcium-rich pyromorphite $[Pb_5(PO_4)_3Cl]$ can act as
415 a main lead-bearing phase in mine-waste soils (Cotter-Howells et al., 1994.)

416 The absence of a greater number of positive correlations between the elements may be a consequence of a
417 relatively small number of samples covered by this research.

418

419 **Monte Carlo simulation**

420

421 The assessment of the ecological risk of Ni, Zn, Pb, Cu, Cr, Cd, and As, was conducted by the Monte Carlo
422 simulation (Fig. 5, Table S9). Ecological Risk Analysis results, arranged for each toxic element in order from
423 highest to lowest probability, showed: Low risk for Ni and Cr; Lower > Low > Median > High risk for Zn; High
424 > Extremely High > Median > Lower ≈ Low risk for Pb, Cu, Cd and As. It is possible to conclude that there is a
425 high risk of Pb, Cu, Cd and As pollution in the investigated area. The most of the investigated sites possess a high
426 and extremely high-risk probabilities for Pb, Cu, Cd and As, as it was shown in result analysis. In comparison
427 with the assessment of the ecological risk that have been carried out so far in localities in Serbia (Sakan et al.,
428 2021), those results are the one with the highest estimated risk values so far. If the form of the HRI graph in the
429 Fig. 5 is considered it could be expected that the most pronounced toxic, elements, Pb, Cu, Cd and As could also
430 have a majority of their concentration in high mobility fractions, see (Sakan et al., 2021).

431 Based on the assessment of the total HRI (Table S10), the highest probability in the examined area is for low risk,
432 then lower and finally median risk. However, given that the individual risks of Pb, Cu, Cd, and As pollution are
433 extremely high, there is a need for improvements in tailings management strategies and to prevent any further
434 exploitation of ore deposits.

435

436 **Conclusion**

437

438 The impact of the mining activities in the Karamanica and Blagodot ore fields on the pollution status of river and
439 lake sediment was estimated in this research. The geochemical signature of the mining activity in surface
440 sediments in the most southeastern part of Serbia (Krajište area) was evidenced by notable contents of the PTEs
441 associated with the Pb, Zn, and Cu mines and tailings.

442 Obtained results confirm that hydroxides of iron play a role in the distribution of Cr and that manganese
443 hydroxides are significant for Zn, Cu, and Cd bounding. Positive correlation of Mn with Zn, Cu, S and Cd can
444 also point to the common origin of these elements, i.e. the source of these elements that is associated with mining
445 activities. The origin of Pb, Cd, As, Zn, and Cu is anthropogenic, rather than a simple crustal, whereas Ni and Cr
446 are dominantly of geochemical origin. It is possible to conclude that water originating from the area near the
447 flotation tailings and pilot plant contains high concentrations of toxic elements, causing the high content of these
448 elements in the examined river sediments. The elevated levels of PTEs in sediments indicate the existence of a
449 potential risk to the environment and health of the local population.

450 Taking into account the individual contamination factors, Pb is the biggest contributor to pollution in most
451 localities, followed by Cu, Cd and As. The values of contamination factors usually indicated very high levels of
452 contamination. Also, there is significant pollution by Zn, followed by Mn, and low contamination with respect to
453 Ni and Cr was observed.

454 The highest values of the contamination factors were registered for the river sediment from Crna river, upriver
455 from the confluence with Krasnodolska river, in the vicinity of Musulj-Kekerinci dump (Blagodot ore field). The
456 results of MS were high for that locality and the Pb content was extremely high, indicating the existence of a high
457 probability that Pb could cause harm to the environment. In this manuscript, the use of MS as a contamination
458 indicator has been shown to be efficient for contamination by Pb.

459 Based on the Monte Carlo simulation, the individual risks of Pb, Cu, Cd, and As pollution are extremely high,
460 and the recommendation is that additional protection measures should be taken regarding tailings and other mining
461 activities in the investigated area, and that any further exploitation of ore deposits should be prevented. The evident
462 potential risk of toxic elements is a strong argument for the continuous monitoring of studied sediments. The
463 methodology applied here, namely the combined use of pollution indices, geochemical background values,
464 magnetic susceptibility, statistical analysis, and Monte Carlo analysis, gives a suitable information about
465 contamination levels, sources of PTEs, as well as the current and future potential risks of mining activities to the
466 downstream environments. Those methods have been used separately earlier, but their combined use is proven to
467 be a more valuable tool for the analysis of the situation since if the several independent methods point to same
468 result, the certainty of the obtained data is more pronounced.

469

470 **Author Contributions**

471 S.S: conceptualization, formal analysis, methodology, data curation, supervision, validation, visualization, writing
472 original draft. A.M.Z: investigation, validation, visualization, writing-review and editing. N.S: methodology,
473 software, writing-review and editing. S.F.B: investigation, formal analysis, writing-review and editing. I.K: formal
474 analysis, writing-review and editing. D.Đ: conceptualization, methodology, supervision, writing-review and
475 editing.

476
477
478
479
480
481
482
483
484
485
486
487
488
489
490
491
492
493
494
495
496
497
498
499
500
501
502
503
504
505
506
507
508
509
510
511
512
513
514
515
516
517
518
519
520
521
522
523
524
525
526
527
528
529
530
531
532
533
534
535

Funding

This research has been financially supported by the Ministry of Science, Technological Development and Innovation of Republic of Serbia (Contract No: 451-03-47/2023-01/200026).

Data availability

The datasets used and/or analyzed during the current study are available from the corresponding author on reasonable request.

Declarations

Ethics approval and consent to participate

Not applicable.

Consent for publication

Not applicable.

Competing interests

The authors declare no competing interests.

Appendix A. Supplementary data

References

- Acevedo-Figueroa D, Jiménez BD, Rodríguez-Sierra CJ (2006) Trace metals in sediments of two estuarine lagoons from Puerto Rico. *Environ Pollut* 1414 (2):336-342. <https://doi.org/10.1016/j.envpol.2005.08.037>
- Arain, MB, Kazi, TG, Jamali, M.K, Jalbani N, Afridi, HI, Baig JA (2008) Speciation of Heavy Metals in Sediment by Conventional, Ultrasound and Microwave Assisted Single Extraction Methods: A Comparison with Modified Sequential Extraction Procedure. *J Hazard Mater* 154:998–1006. doi:10.1016/j.jhazmat.2007.11.004
- Ayari J, Barbieri M, Agnan Y, Sellami A, Braham A, Dhaha F, Charef A (2021) Trace element contamination in the mine-affected stream sediments of Oued Rarai in north-western Tunisia: a river basin scale assessment. *Environ Geochem Health* 43 (10):4027–4042. <https://doi.org/10.1007/s10653-021-00887-1>
- Ayari J, Barbieri M, Barhoumi A, Boschetti T, Braham A, Dhaha F, Charef A (2023) Trace metal element pollution in media from the abandoned Pb and Zn mine of Lakhouat, Northern Tunisia. *J Geochem Explor* 247:107180. <https://doi.org/10.1016/j.gexplo.2023.107180>
- Barago N, Cristiano M, Pavoni E, Floreani F, Parisi F, Lenaz D, Covelli, S. (2023) Environmental impact of potentially toxic elements on soils, sediments, waters, and air nearby an abandoned Hg-rich fahlrore mine (Mt. Avanza, Carnic Alps, NE Italy). *Environ Sci Poll Res* 30:63754–63775. <https://doi.org/10.1007/s11356-023-26629-7>
- Chaparro M, Ramirez-Ramirez M, Chaparro M, Miranda-Aviles R, Puy-Alquiza M, Bohnel H, Zanol G (2020). Magnetic parameters as proxies for anthropogenic pollution in water reservoir sediments from Mexico: an interdisciplinary approach. *Sci Total Environ* 700:134343. <https://doi.org/10.1016/j.scitotenv.2019.134343>
- Chen ZL, Zhang JQ, Huan, L, Yuan ZH, Li ZJ, Liu MC (2019) Removal of Cd and Pb with biochar made from dairy manure at low temperature. *J Integr Agric* 18(1):201–210. [https://doi.org/10.1016/S2095-3119\(18\)61987-2](https://doi.org/10.1016/S2095-3119(18)61987-2)
- Cotter-Howells JD, Champness PE, Charnock JM, Patrick RAD (1994) Identification of pyromorphite in mine-waste contaminated soils by ATEM and EXAFS. *Eur J Soil Sci* 45:393-402. <https://doi.org/10.1111/j.1365-2389.1994.tb00524.x>.
- Desai CC, Patel KN, Shukla, MJ (1989) Some aspects of the magnetic susceptibility of ferroelectric magnesium hydrogen phosphate crystals. *J Mater Sci Lett* 8:1391-1392. <https://doi.org/10.1007/BF00720197>
- Donkor AK, Bonzongo J-CJ, Nartey VK, Adotey DK (2005) Heavy metals in sediments of the gold mining impacted Pra River Basin, Ghana, West Africa. *Soil Sediment Contam* 14:479–503. <https://doi.org/10.1080/15320380500263675>

536 Dos Santos S N, Alleoni LRF (2013) Reference Values for Heavy Metals in Soils of the Brazilian Agricultural
537 Frontier in Southwestern Amazônia. *Environ Monit Assess* 185:5737–5748.
538 <https://doi.org/10.1007/s10661-012-2980-7>

539 Duodu GO, Goonetilleke A, Ayoko GA (2016) Comparison of pollution indices for the assessment of heavy metal
540 in Brisbane River sediment. *Environ Pollut* 219:1077–1091.
541 <https://doi.org/10.1016/j.envpol.2016.09.008>

542 Đokić BV, Jovanović M (2008) Tailing dumps of Blagodot – need to establish a true relationship between man
543 and nature. *Recycl. Sustain. Dev.* 1(1):60–66. (in Serbian) https://www.rsd.tfbor.bg.ac.rs/?page_id=45
544 (assessed 25.06.2023).

545 Đokić B (2012) Geochemical characteristics of the Grot mine flotation tailing (southeast Serbia). Doctoral
546 Dissertation, University of Belgrade, Faculty of Mining and Geology. (in Serbian)
547 <https://nardus.mpn.gov.rs/handle/123456789/2657?show=full>

548 Đokić BV, Jović V, Jovanović M, Ćirić A, Jovanović D (2012) Geochemical behaviour of some heavy metals of
549 the Grot flotation tailing, Southeast Serbia. *Environ Earth Sci* 66:933–939.
550 <https://doi.org/10.1007/s12665-011-1303-6>

551 Đokić BV, Jovanović M, Đokić O (2013) Heavy metals in the soil around the flotation tailing dump of the lead-
552 zinc mine Grot mine, southeast Serbia. *Min. Metall. Eng. Bor* 2:135–146.
553 <https://doi.org/10.5937/mmeb1302135D>

554 Đorđević L, Živković N, Živković L, Đorđević A (2012) Assessment of heavy metals pollution in sediments of
555 the Korbevačka river in Southeastern Serbia. *Soil Sediment Contam* 21:889–900.
556 <https://doi.org/10.1080/15320383.2012.699110>

557 Fakhri Y, Mousavi Khaneghah A, Conti GO, Ferrante M, Khezri A, Darvishi A, Ahmadi M, Hasanzadeh V,
558 Rahimizadeh A, Keramati H, Moradi B, Amanidaz N (2018) Probabilistic risk assessment (Monte Carlo
559 simulation method) of Pb and Cd in the onion bulb (*Allium cepa*) and soil of Iran. *Environ Sci Pollut Res*
560 25:30894–30906.
561 <https://doi.org/10.1007/s11356-018-3081-0>

562 Gantayat RR, Viswanathan PM, Ramasamy N, Sabarathinam C (2023) Distribution and fractionation of metals in
563 tropical estuarine sediments, NW Borneo: Implication for ecological risk assessment. *J Geochem Explor*
564 252:107253.
565 <https://doi.org/10.1016/j.gexplo.2023.107253>

566 García-Giménez R, Jiménez-Ballesta R (2017) Mine tailings influencing soil contamination by potentially toxic
567 elements. *Environ Earth Sci* 76:51.
568 <https://doi.org/10.1007/s12665-016-6376-9>

569 Government of Republic of Serbia (2012) Regulation on Limit Values for Pollutants in Surface and Ground
570 Waters and Sediments and the Deadlines for Their Achievement. Off Gaz Rep Serbia No 50/12 (in
571 Serbian)

572 Hakanson L (1980) Ecological risk index for aquatic pollution control, a sedimentological approach, *Water Res*
573 14:975–1001.
574 [https://doi.org/10.1016/0043-1354\(80\)90143-8](https://doi.org/10.1016/0043-1354(80)90143-8)

575 Haris H, Looi LJ, Aris AZ, Mokhtar NF, Ayob NAA, Yusoff FM, Salleh AB, Praveena SM (2017) Geo-
576 accumulation index and contamination factors of heavy metals (Zn and Pb) in urban river
577 sediment. *Environ Geochem Health* 39:1259–1271.
578 <https://doi.org/10.1007/s10653-017-9971-0>

579 Jamshidi-Zanjani A, Saeedi M (2017) Multivariate analysis and geochemical approach for assessment of metal
580 pollution state in sediment cores. *Environ Sci Pollut Res* 24:16289–16304.

581 Karimi R, Ayoubi S, Jalalian A, Sheikh-Hosseini AR, Afyuni M (2011) Relationships between magnetic
582 susceptibility and heavy metals in urban topsoils in the arid region of Isfahan, central Iran. *J Appl*
583 *Geophys* 74:1–7.
584 <https://doi.org/10.1016/j.jappgeo.2011.02.009>

585 Kowalska JB, Mazurek R, Gsiorek M, Zaleski T (2018) Pollution indices as useful tools for the comprehensive
586 evaluation of the degree of soil contamination – A review. *Environ Geochem Health* 40:2395–2420.
587 <https://doi.org/10.1007/s10653-018-0106-z>

588 Lidman J, Olid C, Bigler C, Berglund AMM (2023) Effect of past century mining activities on sediment properties
589 and toxicity to freshwater organisms in northern Sweden. *Sci Total Environ* 872:162097.
590 <https://doi.org/10.1016/j.scitotenv.2023.162097>

591 MacDonald DD, Ingersoll CG, Berger TA (2000) Development and Evaluation of Consensus-Based Sediment
592 Quality Guidelines for Freshwater Ecosystems. *Archives Environ Contam Toxicol* 39:20–31.
593 <https://doi.org/10.1007/s002440010075>

- 594 Maftai AE, Iancu OG, Buzgar N (2014) Assessment of minor elements contamination in Bistrița River sediments
595 (upstream of Izvorul Muntelui Lake, Romania) with the implication of mining activity. *J Geochem*
596 *Explor* 145:25-34.
597 <http://dx.doi.org/10.1016/j.gexplo.2014.05.001>
- 598 Magiera T (2022) ISO 21226:2019 "Soil quality – Guideline for the screening of soil polluted with toxic elements
599 using soil magnetometry". 17th "Castle Meeting" on Palaeo, Rock and Environmental Magnetism. 28
600 August-3 september 2022, Castle Trakošćan, Croatia. Book of Abstracts, P.78, ISBN 978-953-7941-44-
601 4. <https://castle2020.irb.hr/Program-and-Book-of-Abstracts>.
- 602 Milošević V, Miličević S, Trumić M, Trumić M, Bartulović Z, Todorović D, Jovanović V, Ivošević B, Čarapić J,
603 Jogrić R, Adamović V (2015) Defining a new production line for the flotation concentrating of copper,
604 lead and zinc minerals from polymetallic ore of the deposit "Podvirovi and Conjev Kamen" - Bosilmetal
605 applying modern technical and technological solutions. Technical solution. Institute for Technology of
606 Nuclear and Other Mineral Raw Materials, Decision 13/6-7, 27.11.2015. (in Serbian)
607 http://www.itnms.ac.rs/downloads/tehnicka_resenja/Kompletno%20resenje%20Bosilmetal%20TR3300
608 7.pdf. (assessed 15.06.2023)
- 609 Müller G (1979) Schwermetalle in den Sedimenten des Rheins: Veränderungen seit 1971. *Umschau* 79,778–783.
610 Oyarzun R, Lillo J, López-García JA, Esbrí JM, Cubas P, Llanos W, Higuera P (2011) The Mazarrón Pb–(Ag)–
611 Zn mining district (SE Spain) as a source of heavy metal contamination in a semiarid realm:
612 Geochemical data from mine wastes, soils, and stream sediments. *J Geochem Explor* 109:113–124.
613 <https://doi.org/10.1016/j.gexplo.2010.04.009>
- 614 Paradelo R, Moldes AB, Barral MT, Moldes AB (2009) Magnetic susceptibility as an indicator of heavy metal
615 contamination in compost. *Waste Manag Res* 27:46–51.
616 <https://doi.org/10.1177/0734242X07082962>
- 617 Pekey H, Karakaş D, Ayberk S, Tolun L, Bakoğlu M (2004) Ecological risk assessment using trace elements from
618 surface sediments of Izmit Bay (Northeastern Marmara Sea) Turkey, *Mar Pollut Bull* 48:946-953.
619 <https://doi.org/10.1016/j.marpolbul.2003.11.023>
- 620 Pujiwati A, Nakamura K, Wang J, Watanabe N, Komai T (2022) Potentially toxic elements pose significant
621 and long-term human health risks in river basin districts with abandoned gold mines. *Environ Geochem*
622 *Health* 44:4685–4702.
623 <https://doi.org/10.1007/s10653-022-01229-5>
- 624 Putra R, Rifai H, Wurster CM (2019) Relationship between magnetic susceptibility and elemental composition of
625 Guano from Solek Cave, West Sumatera. *IOP Conf Series: J Phys: Conf Ser* 1185:012011 IOP
626 Publishing.
627 <https://doi.org/10.1088/1742-6596/1185/1/012011>
- 628 Qu L, Huang H, Xia F, Liu Y, Dahlgren RA, Zhang M, Mei K (2018) Risk analysis of heavy metal concentration
629 in surface waters across the rural-urban interface of the Wen-Rui Tang River, China. *Environ Pollut* 237:
630 639–649.
631 <https://doi.org/10.1016/j.envpol.2018.02.020>
- 632 Request for determining the scope and content of the Environmental Impact Assessment Study of the Pb, Zn and
633 Cu ore exploitation project from the "Podvirovi" and "Popovica" deposits in the area of Karamanica near
634 Bosilegrad (2021) Submitted by „BOSIL-METAL“ d.o.o. to the Ministry of Environmental Protection
635 of the Republic of Serbia. (in Serbian) [https://www.ekologija.gov.rs/sites/default/files/inline-](https://www.ekologija.gov.rs/sites/default/files/inline-files/Zahtev%20o%20obimu%20i%20sadr%C5%BEaju_BOSIL-METAL_tekst%20Bprilozi.pdf)
636 [files/Zahtev%20o%20obimu%20i%20sadr%C5%BEaju_BOSIL-METAL_tekst%20Bprilozi.pdf](https://www.ekologija.gov.rs/sites/default/files/inline-files/Zahtev%20o%20obimu%20i%20sadr%C5%BEaju_BOSIL-METAL_tekst%20Bprilozi.pdf)
637 (assessed 20.08.2023)
- 638 Sakan S, Popović A, Anđelković I, Đorđević D (2016) Aquatic sediments pollution estimate using the metal
639 fractionation, secondary phase enrichment factor calculation, and used statistical methods. *Environ*
640 *Geochem Health* 38:855-867.
641 <https://doi.org/10.1007/s10653-015-9766-0>
- 642 Sakan S, Frančišković-Bilinski S, Đorđević D, Popović A, Sakan N, Škrivanj S, Bilinski H (2021) Evaluation of
643 Element Mobility in River Sediment Using Different Single Extraction Procedures and Assessment of
644 Probabilistic Ecological Risk. *Water*, 13:1411.
645 <https://doi.org/10.3390/w13101411>
- 646 Sakan S, Mihajlidi-Zelić A, Škrivanj S, Frančišković-Bilinski S, Đorđević D (2022) An Integrated Approach in
647 the Assessment of the Vlasina River System Pollution by Toxic Elements. *Front Environ Sci* 10:909858.
648 <https://doi.org/10.3389/fenvs.2022.909858>
- 649 Sakan S, Mihajlidi-Zelić A, Ašković K, Sakan N, Trifunović S, Đorđević D (2023) The significance of applying
650 different factors for the evaluation of sediment contamination by toxic elements and estimation of the
651 ecological risk. *Environ Sci Pollut Res* 30:53461–53477.
652 <https://doi.org/10.1007/s11356-023-26111-4>

653 Shen F, Mao L, Sun R, Du J, Tan Z, Ding M (2019) Contamination Evaluation and Source Identification of Heavy
654 Metals in the Sediments from the Lishui River Watershed, Southern China. *Int J Environ Res Public*
655 *Health* 16 (3):336.
656 <https://doi.org/10.3390/ijerph16030336>

657 Shen T, Tang Y, Li YJ, Liu Y, Hu H (2020) An experimental study about the effects of phosphorus loading in
658 river sediment on the transport of lead and cadmium at sedimentwater interface. *Sci Total Environ* 720:
659 137535.
660 <https://doi.org/10.1016/j.scitotenv.2020.137535>

661 Shikazono N, Zakir HM, Sudo Y (2008) Zinc contamination in river water and sediments at Taisyu Zn–Pb mine
662 area, Tsushima Island, Japan. *J Geochem Explor* 98:80–88.
663 <https://doi.org/10.1016/j.gexplo.2007.12.002>

664 Simić M (2001) Metallogeny of Mačkatica-Blagodot-Karamanica Zone, Special Editions of Geoinstitute, Book
665 28, Geoinstitute, Belgrade, ISBN 86-7156-012-0 (in Serbian)

666 Simić MM (2002) Lead and zinc potential of the Karamanica ore field, Serbia. *Tehnika* , 53(4-5), 7-16. (in
667 Serbian)
668 <https://scindeks.ceon.rs/article.aspx?artid=0350-26270205007S&lang=en>. (assessed 01.07.2023)

669 Stanojević D, Filipović-Petrović L (2014) A contribution to integrated metal valorisation in zinc hydrometallurgy.
670 *Zastita Materijala* 55 (1):11-25. (in Serbian)
671 <https://doi.org/10.5937/ZasMat1401011S>

672 Stojmenović M, Pašalić S, Kragović M (2017) Influence of the underground mining waste on the environmental
673 tailings and wastewater characterization. *Undgrd Min Eng* 31:85-100.

674 Sutherland RA (2010) BCR-701: A Review of 10-years of Sequential Extraction Analyses. *Anal Chim Acta* 680:
675 10–20.
676 <https://doi:10.1016/j.aca.2010.09.016>

677 Tomlinson DL, Wilson JG, Harris CR, Jeffrey DW (1980) Problems in the assessment of heavy-metal levels in
678 estuaries and the formation of a pollution index. *Helgolander Mar Res* 33:566-575.
679 <https://doi.org/10.1007/BF02414780>

680 Ustaoglu F, Islam MdS (2020) Potential toxic elements in sediment of some rivers at Giresun, Northeast Turkey:
681 A preliminary assessment for ecotoxicological status and health risk. *Ecol Indic* 113:106237.
682 <https://doi.org/10.1016/j.ecolind.2020.106237>

683 Wedepohl KH (1995) The composition of the continental crust. *Geochim Cosmochim Acta* 59 (7):1217–1232.
684 [https://doi.org/10.1016/0016-7037\(95\)00038-2](https://doi.org/10.1016/0016-7037(95)00038-2)

685 Werner F, Prietze, J (2015) Standard Protocol and Quality Assessment of Soil Phosphorus Speciation by P K-
686 Edge XANES Spectroscopy. *Environ Sci Technol* 49:10521–10528.
687 <https://doi.org/10.1021/acs.est.5b03096>

688 Wolfenden, PJ, Lewin J (1978) Distribution of metal pollutants in active stream sediments. *Catena* 5:67–78.
689 [https://doi.org/10.1016/S0341-8162\(78\)80007-1](https://doi.org/10.1016/S0341-8162(78)80007-1)

690 Yang Y, Li S, Bi X, Wu P, Liu T, Li F, Liu C (2010) Lead, Zn, and Cd in slags, stream sediments, and soils in an
691 abandoned Zn smelting region, southwest of China, and Pb and S isotopes as source tracers. *J Soils Sedim*
692 *10:1527–1539*.
693 <https://doi.org/10.1007/s11368-010-0253-z>

694 Yucel DS, Baba A (2018) Determining water and sediment quality related to lead-zinc mining activity.
695 *Arch Environ Prot* 44 (3):19–30.
696 <http://dx.doi.org/10.24425/aep.2018.122294>
697 <https://www.mindat.org/loc-62233.html> (assessed 22.07.2023)
698 <https://www.serbiemap.net> (assessed 03.06.2023)

699
700
701
702
703
704
705
706
707
708
709
710
711
712

713
714

Table 1
Sampling locations with the GPS coordinates

Label	Name of locality	GPS coordinates
1	The nameless stream that flows into the Karamanička river (upstream of the flotation tailing dump of the Podvirovi ore deposit flotation pilot plant)	N 42.349999°, E 22.3384774°
2	The Nameless stream that flows into the Karamanička river (downstream of the flotation tailing dump of the Podvirovi ore deposit flotation pilot plant)	N 42.3444682°, E 22.3449898°
3	Golema river (near village Omatica)	N 42.3410672°, E 22.3668551°
4	Brankovačka river (village Brankovci, before the Gradište small hydropower plant dam)	N 42.4150061°, E 22.4915979°
5	Krasnodolska river (near the confluence with Crna river)	N 42.5171703°, E 22.248264°
6	Crna river (upriver from the confluence with Krasnodolska river)	N 42.516755°, E 22.248210°
7	Crna river (from the confluence with Krasnodolska river)	N 42.516702°, E 22.248553°
8	Ljubatska river (before the water intake for Lisina lake)	N 42.5028795°, E 22.30225°
9	Lisina lake (near the dam)	N 42.5508947°, E 22.3880402°
10	Vlasina lake (near the Božica canal)	N 42.6792604°, E 22.3519681°

715
716
717
718
719
720
721
722
723
724
725
726
727
728
729
730
731
732
733
734
735
736
737
738
739
740
741
742
743
744
745
746
747
748
749
750

751
752
753
754
755
756
757
758
759
760
761
762
763
764
765
766
767
768
769
770
771
772
773
774
775
776
777
778
779
780
781
782
783
784

Table 2

Descriptive statistics of the content of the studied elements, Crust values and Quality Guidelines [mg kg⁻¹]

	Min	Max	Mean	SD ^a	Median	CV ^{**}	HM ^b	GM ^c	TEC ^d	PEL ^e	MAV ^f	RV ^g	Crust ^h
Zn	74.3	7,892	3,521	2,992	4,074	85.0	403	1586	121	459	430	720	65
Ni	17.7	45.8	32.3	10.10	30.00	31.3	29.39	30.83	22.7	48.6	44	210	56
Cu	21.2	3,224	1,183	1,316	506	111	92	367	31.6	149	110	190	25
Cr	29.6	91.5	43.8	20.19	36.6	46.1	38.6	40.7	43.4	111	240	380	126
Pb	15	36,790	6,848	10,986	4,497	160	128	1,663	35.8	128	310	530	14.8
Cd	0.06	64.2	25.5	25.6	18.0	100	0.5	7.7	0.99	4.98	6.4	12	0.1
As	4.2	1,444	396	588	50	148	25	88	9.79	33	42	55	1.7
Al	9,877	33,011	17,954	6,977	16,132	38.9	16,030	16,911	nd ^{***}	nd	nd	nd	79,600
B	<DL*	111	24.4	45.9	0.1	188	0.1	0.8	nd	nd	nd	nd	11
Ba	34.5	431	129	110	98	85.3	89	104	nd	nd	nd	nd	584
Ca	4,388	46,397	18,869	16,410	11,311	87.0	9,227	12,945	nd	nd	nd	nd	38,500
Co	4.24	39.6	18.2	9.2	17.5	50.5	13.4	16.0	nd	nd	nd	nd	24
Fe	27,549	108,639	65,615	25,608	58,539	39.0	56,466	61,059	nd	nd	nd	nd	43,200
K	1987	94,094	39,894	40,797	27,922	102	5,242	14,478	nd	nd	nd	nd	21,400
Li	9.16	21.8	14.6	4.3	13.9	29.5	13.6	14.1	nd	nd	nd	nd	18
Mg	6,200	16,265	11,266	2,972	10,660	26.4	10,486	10,889	nd	nd	nd	nd	22,000
Mn	753	6,161	3,193	1,977	4,114	61.9	1,804	2,474	nd	nd	nd	nd	716
Na	2,431	13,558	5,427	4,178	3,273	77.0	3,669	4,328	nd	nd	nd	nd	23,600
P	1,282	2,122	1,527	262	1,478	17.2	1,493	1,509	nd	nd	nd	nd	nd
S	5	13,122	5,676	5,071	7,304	89.3	47	1,443	nd	nd	nd	nd	nd
Sr	5.67	42.4	15.9	10.2	13.3	64.2	12.3	13.8	nd	nd	nd	nd	333
V	26.7	88.9	46.9	21.0	41.6	44.8	40.1	43.2	nd	nd	nd	nd	98

*below detection limit; ** in %; ***nd - no data

^aSD (standard deviation); ^bHM (harmonic mean); ^cGM (geometric mean); ^dTEC (threshold effect concentration)—MacDonald et al. (2000); ^ePEL (probable effect concentration)—MacDonald et al. (2000); ^fMAV, maximum allowable value (Government of Republic of Serbia, 2012); ^gRV, remediation value (Government of Republic of Serbia, 2012); ^hElement concentrations in the Continental Crust—K.H. Wedepohl, (1995).

785
786

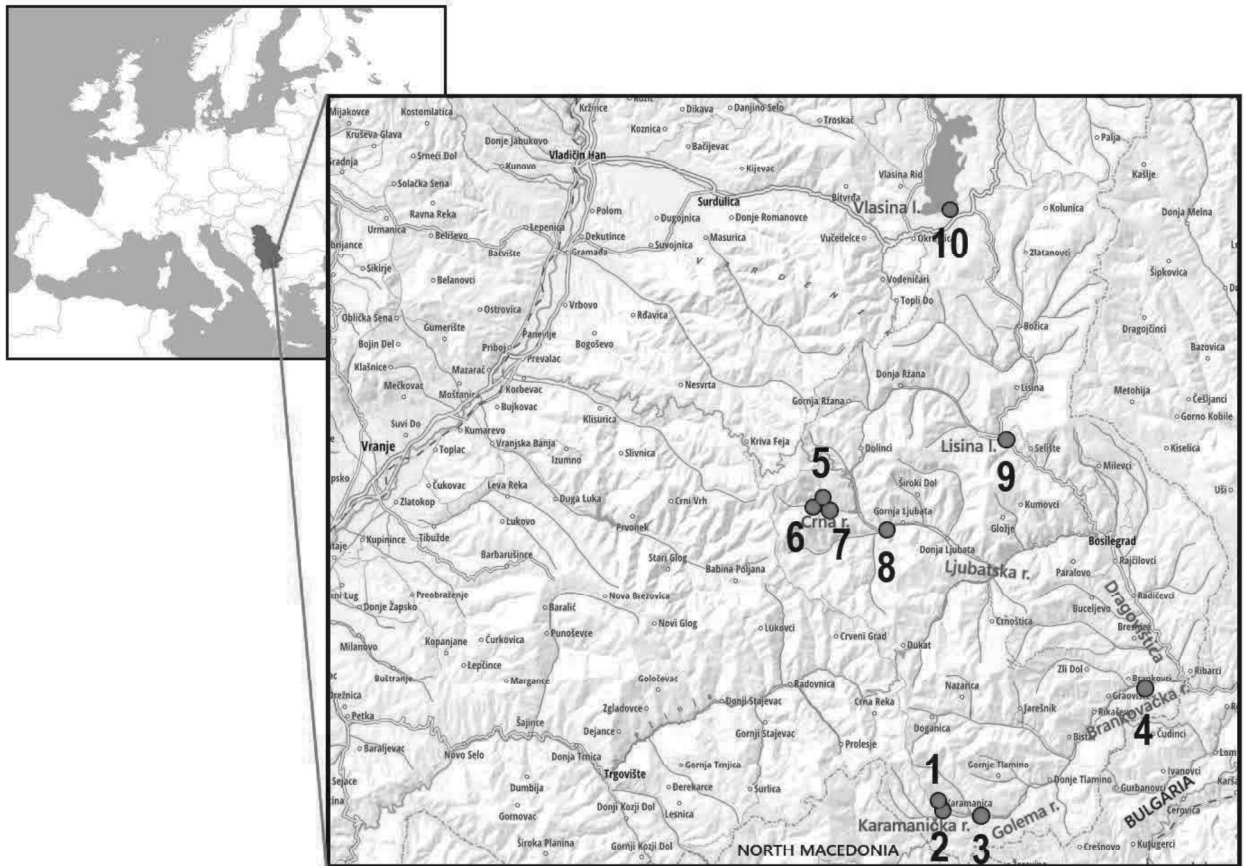
Table 3

Comparison with results for similar river sediments and background values [mg kg⁻¹]

	This study	RS, China ^a	SS, Tunisia ^b	RS, Romania ^c	RS, Japan ^d	RS, UK ^e	SS, China ^f	SS, Spain ^g	BV ^h
Zn	Min	74.3	46.9	nd*	50	86.8	nd	nd	
	Max	7892	325	nd	1117	7490	6,955	30,425	
	Mean	3521	113	5790	126.2	1607	nd	nd	1092.7 74.25
Ni	Min	17.7	18.4	nd	16	nd	nd	nd	
	Max	45.8	45.5	nd	48	nd	nd	nd	
	Mean	32.3	28.9	16.08	30.10	nd	nd	nd	20.57
Cu	Min	21.2	16.7	nd	17	nd	nd	nd	
	Max	3224	92.1	nd	451	nd	2,557	nd	
	Mean	1183	36.7	31.67	45.63	nd	nd	nd	122.2 24.4
Cr	Min	29.6	44.7	nd	39	nd	nd	nd	
	Max	91.5	129	nd	99	nd	nd	nd	
	Mean	43.8	68.8	31.42	71.44	nd	nd	nd	19.88
Pb	Min	15	16.1	nd	17	nd	nd	nd	
	Max	36790	192	nd	139	nd	6,411	21,850	
	Mean	6848	42.5	3542	36.35	nd	nd	nd	2708.6 20.73
Cd	Min	0.06	0.100	nd	0.11	nd	nd	nd	
	Max	64.2	0.910	nd	2.38	nd	44	97	
	Mean	25.5	0.336	16.80	0.43	nd	nd	nd	3.7 0.46
As	Min	4.2	2.46	nd	8.2	nd	nd	nd	
	Max	1444	15.0	nd	170	nd	nd	nd	
	Mean	396	5.10	81.71	24.44	nd	nd	nd	119.8 11.1

787
788
789
790
791
792
793
794
795
796
797
798
799
800
801

*nd – no data; ^a River sediments, China, origin of elements: natural, agricultural and industrial activities, mining, and vehicular traffic (Shen et al. 2019); ^b Stream sediments, Northern Tunisia, near abandoned Pb and Zn mine (Ayari et al. 2023); ^c River sediments, Romania, presence of many waste dumps and underground mining works (closed or still active) which cause the well-known process of acid mine drainage (Maftei et al. 2014); ^d River sediment in the Tsushima Island, Japan, Taisyu Zn–Pb mine area (Shikazono et al. 2008); ^e RS, Afon Twymyn, Wales, river bed sediments from different metal mining regions in the UK (Wolfenden and Lewin, 1978); ^f Stream sediment in Guizhou province, China (Abandoned Zn–Pb smelter, Hezhang Country) (Yang et al. 2010); ^g Stream sediments in Mazarrón Pb–(Ag)–Zn mining district (Oyarzun et al. 2011); ^h Background values (this study): BV for Al is 1264.42 mg kg⁻¹.



802
803
804
805
806
807
808

Figure 1. Map with sampling locations: (1), (2) The Nameless stream that flows into the Karamanička river; (3) Golema river; (4) Brankovačka river; (5) Krasnodolska river; (6) (7) Crna river; (8) Ljubatska river; (9) Lisina lake; and (10) Vlasina lake (adapted from <https://www.serbiamap.net>).

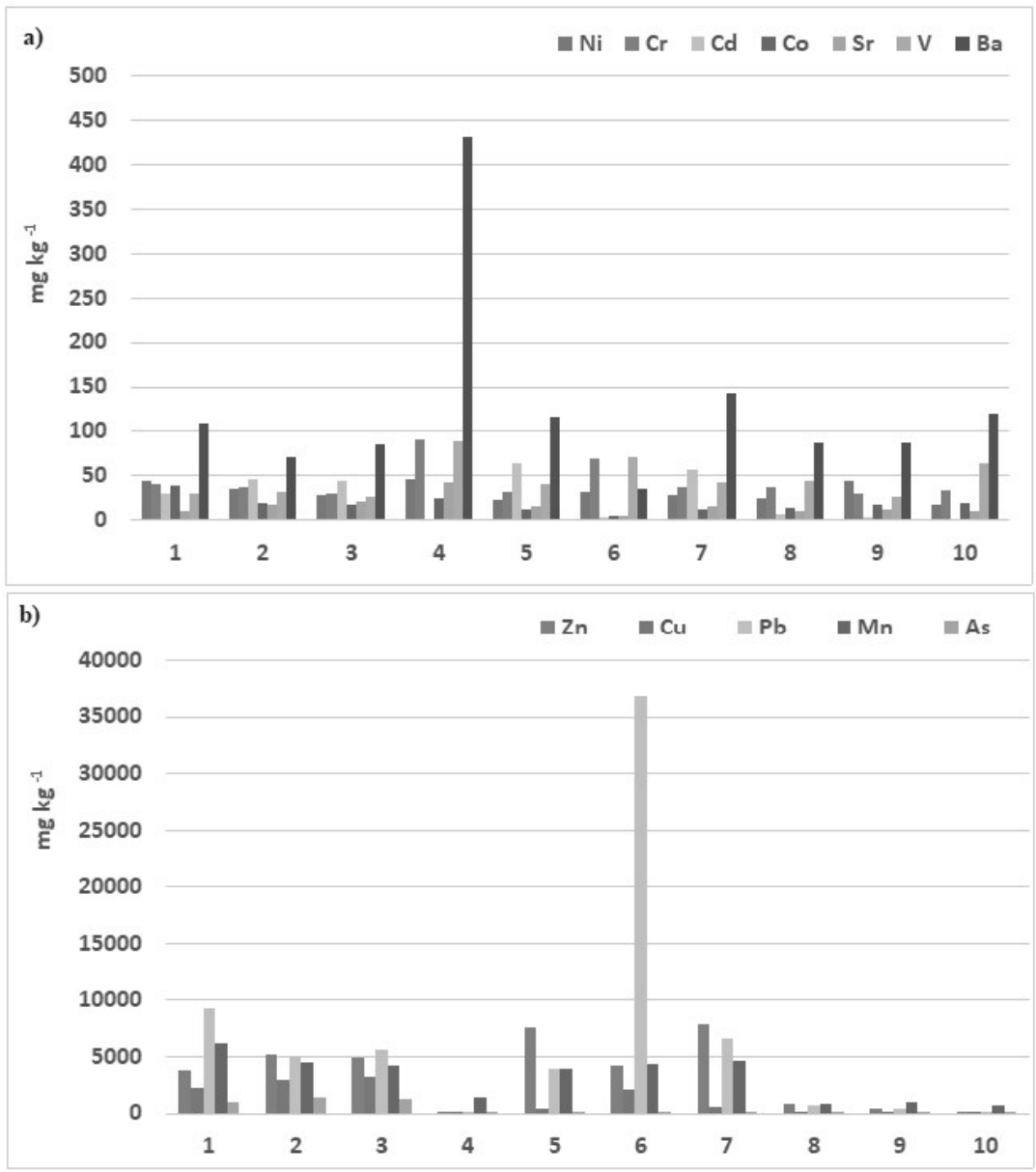


Figure 2. Total content of studied elements in river sediments: (a) Ni, Cr, Cd, Co, Sr, V, and Ba; and (a) Zn, Cu, Pb, Mn, and As.

809
810
811

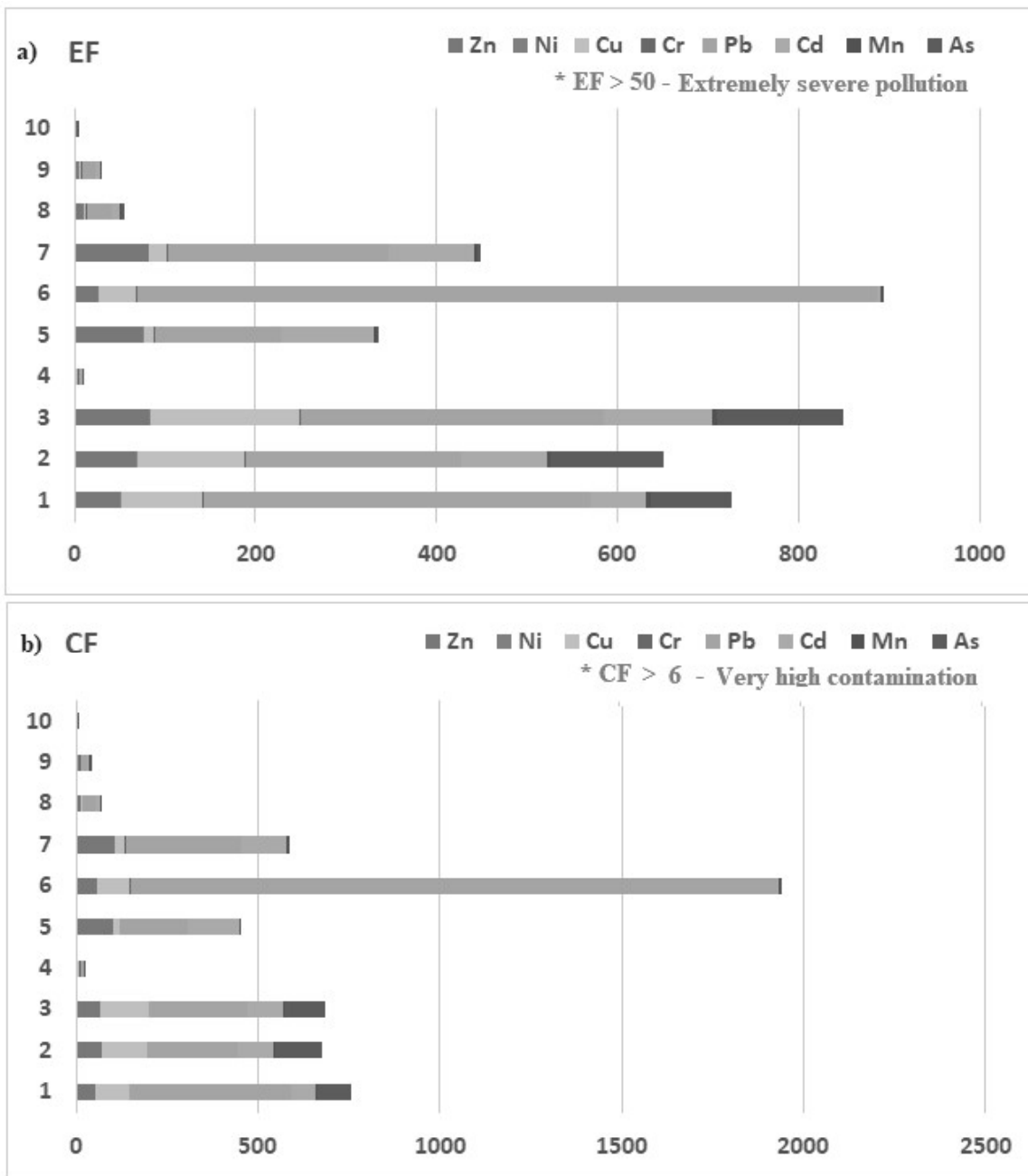


Figure 3. Contamination indexes for studied elements: a) Enrichment factor (EF), and b) Contamination factor (CF).

812
 813
 814
 815
 816
 817

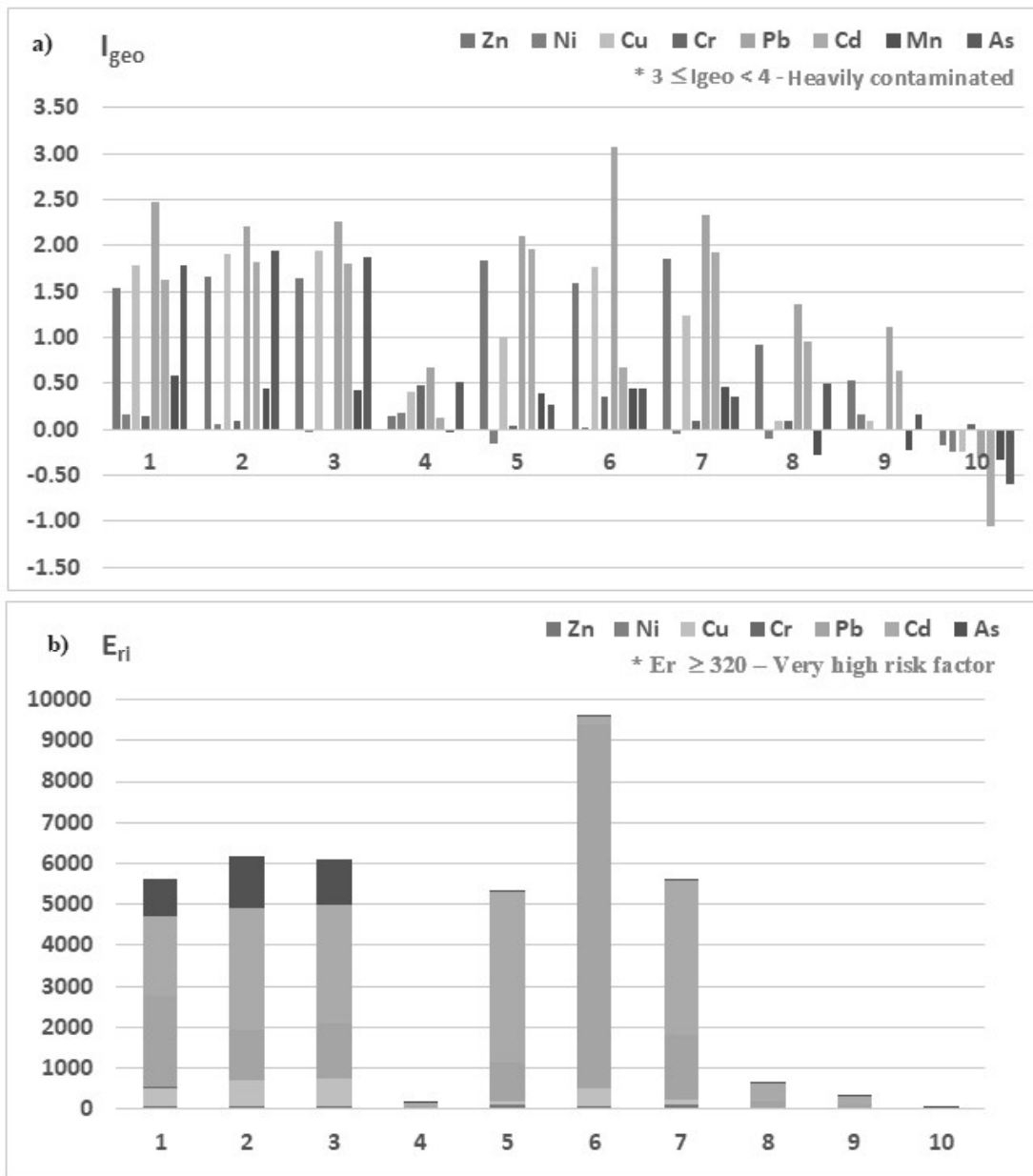
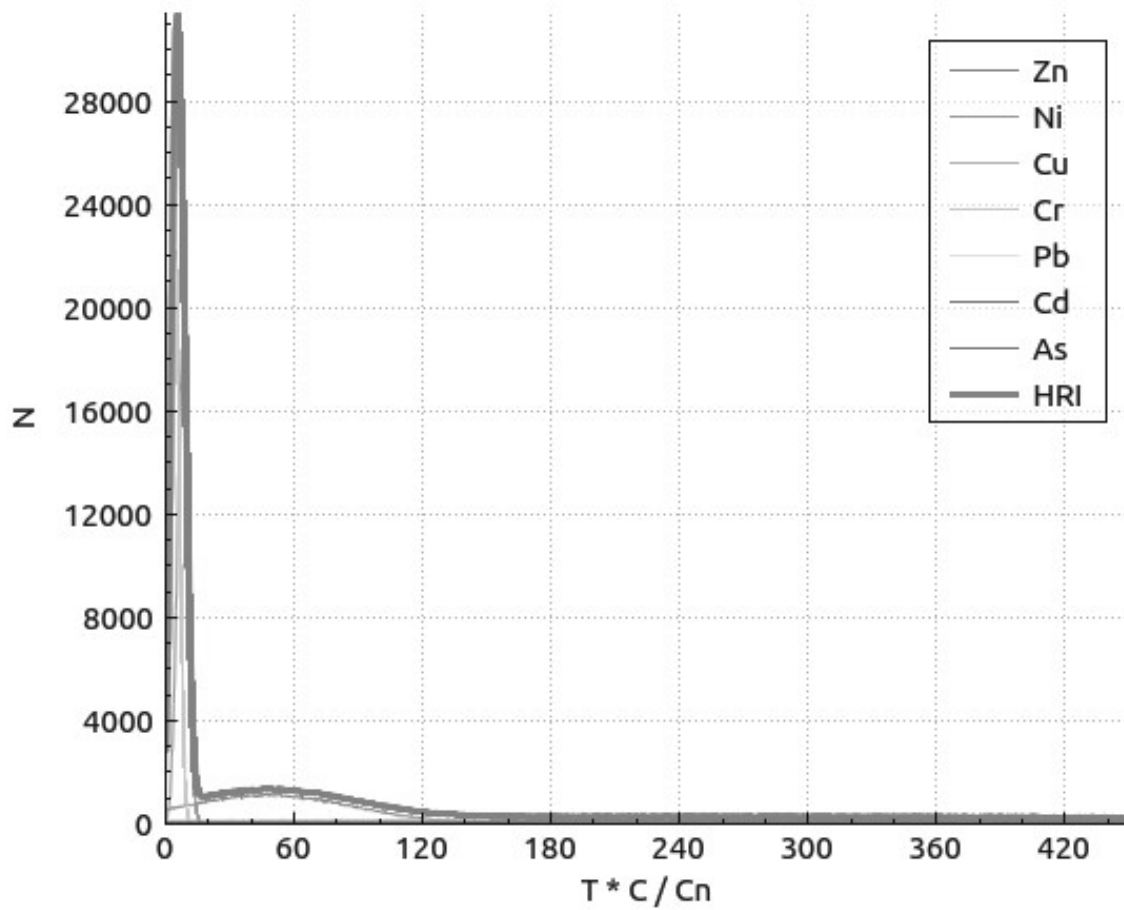


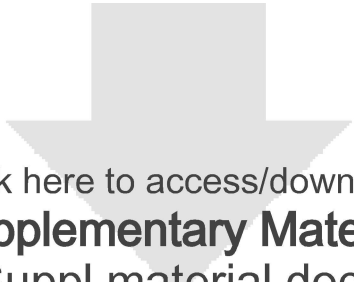
Figure 4. Contamination indexes for studied elements: a) Index of geoaccumulation (I_{geo}); and b) Ecological risk factor (E_{ri}).

818
819
820
821
822
823
824
825
826



827
 828
 829
 830
 831
 832

Figure 5. Distribution curve and exceedance probability curves of the risk index (RI) and total ecological risk comprehensive index HRI based on a Monte Carlo simulation run 100,000 times. Local backgrounds are the reference values for the calculation of Eri.



Click here to access/download
Supplementary Material
Suppl material.docx

