1	Levels and determinants of adipose tissue cadmium concentrations in
2	an adult cohort from Southern Spain
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#### 25 Abstract

This study was conceived as a first step to evaluate the suitability of adipose tissue Cadmium (Cd) concentrations as a biomarker for the assessment of long-term exposure. Specifically, the aim of this work was to explore the socio-demographic, dietary, and lifestyle determinants of adipose tissue Cd concentrations.

The study population is a subsample of GraMo cohort. Adipose tissue samples were intraoperatively collected from 226 adult volunteers recruited in two public hospitals from Granada, Spain. Cd Concentrations in adipose tissue were analyzed by High-Resolution Inductively Coupled Plasma Mass Spectrometry (HR-ICP-MS). Data on socio-demographic characteristics, lifestyle, diet and health status were collected by face-to-face interviews. Predictors of Cd concentrations were assessed by multivariable linear regression with a stepwise variable selection.

36 We found detectable levels of Cd in the adipose tissue of all the study participants, with a mean concentration (±standard deviation) of 12.66±18.91 µg/Kg. Smoking habit at recruitment was 37 associated with increased adipose tissue Cd concentrations ( $\beta$  for smokers=0.669 p<0.001;  $\beta$  for 38 former smokers=0.502, p<0.001; reference=non-smokers). Age was positively associated with Cd 39 concentrations ( $\beta$ =0.014, p<0.001), and men showed lower concentrations than women ( $\beta$ =-0.424, 40 p<0.001). Obesity, measured as Body Mass Index (BMI), showed an inverse association with Cd 41 concentrations ( $\beta$ =-0.038, p<0.001). Egg consumption ≥2 portions/week ( $\beta$ =0.241, p=0.025) was 42 positively associated with Cd concentrations. Perceived exposure to paints was also positively 43 44 associated with Cd concentrations. The observed associations with age, smoking habit, BMI and egg and meat consumption did not substantially change after sex/gender stratification. Our results 45 46 are consistent with currently-known Cd sources and suggest other potential pathways, which might 47 be population-specific.

As a whole, our findings underline the potential relevance of adipose tissue as a biological matrix
for exposure characterization to Cd, as well as for the assessment of long-term clinical implications
of the exposure, particularly in obesity-related diseases.

- **Keywords:** Cadmium; adipose tissue; exposure predictors; sociodemographic characteristics; diet;
- 53 lifestyle.

56 **1. Introduction** 

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Cadmium (Cd) is a metallic element included in the group 12 of the periodic table, with atomic 58 59 number 48. Cd has been widely used in industrial processes, such as certain paints, Cd-Nickel 60 batteries, as an estabiliser in thermoplastics (e.g. polyvinyl chloride), as well as in photography, lithography, tyres and photoelectric cells in solar panels (Herron, 2003). Cd can also be found as 61 an impurity in zinc, lead and copper ores and alloys, iron, steel, fossil fuels, cement and some 62 fertilizers (IARC, 2012). Phosphorous fertilizers are frequently used in greenhouses (Rodríguez-63 Martín et al., 2013), which can be a source of Cd contamination of soils, water, and, consequently, 64 65 the trophic chain (Pan et al., 2010). Other antropogenic sources include recycling, mining and smelting of zinc-bearing ores, the incineration of waste, the combustion of fossil fuels, and the 66 releases of landfills, among others. The global Cd emmission in the mid-90s was estimated in 3000 67 Tm, decreasing by a half in Europe in the period 1990-2003 (UNEP, 2008). Natural sources of Cd 68 can be found in the Earth's crust and oceans with an average abundance of 0.1-0.2 mg/kg. 69

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Cd is considered a relevant persistent pollutant which is slowly degraded in the environment and living organisms (Järup, 2003). Consequently, some of the abovementioned uses have been restricted under the REACH regulation in the European Union in recent years (European Parliament and Council Directive 2013/56/EU, European Commission Regulation 494/2011). However, the majority of the general population still show detectable levels of Cd in their blood and urine (López-Herranz et al., 2016; Pirard et al., 2018).

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Food ingestion and smoking are considered the main sources of Cd exposure in the general population. Ambient air, drinking-water, contaminated soils and dust can also contribute to the exposure, although to a lesser extent, but they could be more relevant for children (Schwartz, 2004). High concentrations of cadmium are commonly found in leafy vegetables, starchy roots, cereals/grains, nuts and pulses, and also in specific animal products, e.g. kidney, liver and certain shellfish (IARC, 2012). The mean dietary exposure for the European general population is estimated in 2.3 µg/kg b.w. per week (EFSA, 2009), which is in a similar range than the Spanish

85 population (AESAN, 2011). Vegetarians frequently have an increased dietary exposure (up to 5.4 µg/kg b.w. per week), as well as regular consumers of bivalve molluscs and wild mushrooms, that 86 also show increased exposure levels (4.6 and 4.3 µg/kg b.w. per week respectively) (IARC, 2012). 87 88 Smoking is considered a relevant source of Cd exposure because of the relatively high 89 concentrations in tobacco leaves. In 1988 the Joint FAO/WHO Expert Committee on Food Additives established a health based guidance value for cadmium of 7 µg/kg b.w. per week. In 90 2009 the CONTAM Panel of the European Food Safety Authority (EFSA) established a tolerable 91 weekly intake of 2.5 µg/kg b.w., which can be easily exceeded in vegetarians, smokers, children 92 93 and people living in contaminated areas (EFSA, 2009).

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Although the health effects of the chronic exposure to low doses of Cd (such as those occuring in the general population) remain unclear, there are certain evidences of renal dysfunction and urinary stone disease, hypertension, lung and prostate cancer, osteoporosis, low birth weight in the offspring, spontaneus miscarriage, obesity, and diabetes (Prozialeck and Edwards, 2012; IARC, 2012; Tinkov et al., 2017). Indeed, Cd is classificated in the category 1 (carcinogenic to humans with a sufficient evidence in humans) by the International Agency for Research on Cancer (IARC, 2012).

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Biomonitoring studies are frequently used to assess human exposure to environmental pollutants 103 104 and their health implications, since internal levels of a pollutant typically account for the overall exposure from different sources and exposure pathways (Needham et al., 2007). The most 105 common human biological matrices used to assess internal levels of Cd are blood and urine. 106 although some research has been performed on hair, nails and saliva. The half-life of Cd strongly 107 108 differs among biological compartments, e.g. from up to 60 years in the kidneys (Ramírez, 2002) to 3-4 months in blood (Talio et al., 2010). However, there is scant research on Cd concentrations in 109 110 the adipose tissue, which is considered the main reservoir of lipophilic pollutants as well as an 111 important biological matrix in the development of chronic non-infectious diseases, e.g. cancer and 112 metabolic syndrome. Indeed, adipose tissue concentrations are considered the most precise and 113 stable estimator for the evaluation of accumulated exposure to other moderately lipophilic

114 pollutants, such as persistent organic pollutants (Kohlmeier and Kohlmeier, 1995; Artacho-Cordón et al., 2015). This might be extended to Cd considering its lipophilic characteristics, e.g. log 115 K<sub>ow</sub>=3.86 (Sakultantimetha et al., 2009). Additionally, adipose tissue is a target for essential trace 116 elements, as an organ in which they can perform their biological effects. Particularly, in vitro 117 experiments with adipocytes have shown that chromium enhances GLUT 4 translocation, which 118 can turn out in increased glucose transport in fat cells (Tinkov et al., 2015; Wiernsperger and 119 Rapin, 2010). Additionally, vanadium has shown an insulin-mimetic potential, although its role as 120 an essential trace element in humans remains unclear (Wiernsperger and Rapin, 2010). Moreover, 121 122 there are experimental evidences that some trace elements in adipose tissue can induce insuline 123 resistance and hypertriglyceridemia in *in vivo* models, suggesting a potential metabolic disrupting effect that should be confirmed in further epidemiologic studies (Hubler et al., 2015; Tinkov et al., 124 2015). Nevertheless, research on the adipose tissue concentrations of toxic and essential trace 125 elements, as well as their biological implications, is still very scarce. Thus, our study aims to shed 126 127 some light in these aspects.

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Considering the abovementioned statements, the present study was conceived as a first step to evaluate the suitability of adipose tissue Cd concentrations as a biomarker for the assessment of long-term exposure. Specifically, the aim of this work was to explore the socio-demographic, dietary, and lifestyle determinants of adipose tissue Cd concentrations.

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#### 134 **2. Materials and methods**

#### 135 **2.1. Study area, design and characteristics of participants**

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This study is part of a wider research which aims to analyse and identify environmental factors affecting the development of chronic diseases in GraMo, an adult cohort in Southern Spain. Study subjects were recruited in two public hospitals from Granada province: San Cecilio University Hospital in the city of Granada (240,000 inhabitants, urban area), and Santa Ana Hospital in the town of Motril (50,000 inhabitants, semi-rural area). The recruitment of the population has been extensively described elsewhere (Arrebola, 2009). The cohort was recruited in 2003-2004, from patients undergoing non-cancer-related surgery [hernias (41%), gallbladder diseases (21%),
varicose veins (12%) and other conditions (26%)]. All subjects signed the informed consent forms
and the study was approved by the Ethics Committee of Granada (Comité de Ética de la
Investigación Biomédica de la Provincia de Granada).

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Out of the 409 individuals contacted, 387 (95%) agreed to participate and were included in the initial cohort. Adequate adipose tissue samples for TE analyses were obtained from 226 (58%), that were used for cross-sectional analyses in the present study. Adipose tissue was collected from pelvic waist (50%), front abdominal wall (40%) and limbs (10%). There were no statistically significant differences in sex/gender or age distributions between participants and non-participants. The main characteristics of the participants are shown in Table 1.

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#### 155 2.2. Independent variables

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Socio-demographic characteristics, life-style, diet and health status data were gathered in face-toface interviews, conducted by trained personnel at the recruitment during their hospital stay. Research procedures were standarised and validated in a pilot study with 50 subjects. The questionnaire was designed and validated in a previous investigation (Buckland et al., 2009; González and Riboli, 2010).

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Body mass index (BMI) was expressed as weight/height squared (kg/m<sup>2</sup>). Participants were considered smokers or alcohol consumers at any level of consumption. Residence in the city of Granada at the time of the surgery was considered "urban" and residence in the area of Motril was considered "semi-rural". The dietary section comprised a food frequency questionnaire that included the following food groups: meat, cold meats, fats, fish, eggs, milk, cheese, vegetables, legumes, fruit, bread, and pasta.

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170 2.3. Sampling and Cd analyses

172 Samples of 5-10 g of adipose tissue were intra-operatively collected and immediately coded and stored at -80 °C until chemical analysis. The adipose tissue samples were freeze-dried in a 173 liophilizator for a minimum of 72 hours, until they reached a plateau weight, at the Slovenian 174 National Building and Civil Engineering Institute (ZAG) in Ljubljana (Slovenia). Then, they were 175 kept at -80°C until their analysis. Subsamples of 0.1 g of adipose tissue underwent total digestion 176 177 with a mixture of 7 mL of HNO<sub>3</sub> and 0.1 mL of HF in a microwave oven (Multiwave 3000, Anton 178 Paar, Graz, Austria). The internal standard added to the digested samples was 1 µg/L of Indium. The multielement analyses of the samples were performed in 2015 at the Laboratory for Inorganic 179 Environmental Geochemistry and Chemodynamics of Nanoparticles, Ruđer Bošković Institute, 180 181 Zagreb (Croatia), by High-Resolution Inductively Coupled Plasma Mass Spectrometry (HR-ICP-MS) using an Element 2 instrument (Thermo, Bremen, Germany). Protocols, conditions and 182 parameters have been described elsewhere (Rodríguez-Pérez et al., 2018). 183

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### 185 2.4. Statistical analyses

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First, the shapes of the relationships between potential continuous predictors and Cd 187 188 concentrations were visually evaluated through locally weighted scatterplot smoothing (LOWESS) 189 and Generalized Additive Models (GAM). Next, predictors of Cd concentrations were examined in multivariable linear regression models, using a combination of backward and forward stepwise 190 variable selection techniques. In order to assess sex/gender-specific predictors, models were 191 stratified by sex/gender (male/female), and women-specific predictors (i.e., accumulated lactation 192 193 time, gravidity, and menopausal status), were entered in the model for females. In pursuance to 194 assess the potential influence of the different sources of adipose tissue (waist, abdomen and limbs) in the associations found, multivariable models were adjusted for this variable, but no 195 relevant modifications in the coefficients were observed (data not shown in tables). The level of 196 197 statistical significance was set at p=0.050 (p=0.100 as borderline significant). In the multivariable 198 models, Cd concentrations were entered as a log-transformed variable.

Generalized standard-error inflation factors were used to verify the absence of collinearity between
 independent variables, while the homoscedasticity was tested by plotting residual against fitted

values. The linearity of quantitative independent variables was checked with partial regression
plots, and the normality of errors was verified using normal QQ plots with 95% confidence
intervals. Data analyses were performed using R statistical computing environment v3.0 (R Core
Team, 2013) and SPSS Statistics 22.0 (IBM Corp., 2013).

## 205 **3. Results and discussion**

## **3.1. Adipose tissue Cd concentrations in the study population**

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The adipose tissue concentrations of Cd and the characteristics of the study population are summarized in Table 1. We detected Cd in all (100%) the analyzed samples, which are depicted in Figure 1.

TABLE 1. Study popula	tion characteristics ar	nd Cd concentrations		
			N	%
	Sox/gondor	Women	99	43.8
	Sex/gender	Men	127	56.2
	Pasidanca	Urban	109	48.2
SOCIODEMOGRAFHIC	Residence	Rural/semi-rural	117	51.8
	Education	Up to primary studies	71	31.4
	Lucation	Secondary/university	155	68.6
	Smoking	Non-smoker	90	40.0
		Current smoker	68	30.0
		Former smoker	68	30.0
	Alcohol consumer		114	50.4
LIFESTYLE AND DIFT		<2 portions/week	107	47.1
	Cheese consumption	2-6 portions/week	70	31.1
		>6 portions/week	49	21.8
		<2 portions/week	86	37.5
	Egg consumption	=2 portions/week	73	32.6
		>2 portions/week	67	29.9
	Meat consumption	≤2 portions/week	85	37.4

		>2 portions/week	141	61.2
	Processed meat		201	88.9
	consumer			
	Chicken consumer		175	77.4
	Vegetables	≤1 portion/week	60	26.9
		=2 portions/week	61	27.4
	consumption	>2 portions/week	105	45.7
		<1 portions/week	13	5.8
	Legumes consumption	1-2 portions/week	120	52.7
		>2 portions/week	93	41.5
		≤1 portions/week	69	30.7
	Fish consumption	=2 portions/week	84	36.9
		>2 portions/week	73	32.4
		Fatty fish	51	22.5
	Fish type	Lean fish	70	31.0
		Both	105	46.5
	Self-perceived	Yes	43	19.0
	exposure to paints			
WOMEN SPECIFIC	Post-menopausal		45	45.5
	Breastfeeding Yes		77	89.5
	Mean	Standard Deviation	Perce	ntiles
			25 <sup>th</sup>	75 <sup>th</sup>
Age (yrs)	53.6	11.8	47.0	62.0
BMI (kg/m²)	26.7	4.4	23.4	29.3
Water consumption	5.8	4 2	3.0	8.0
(glasses/day)	0.0		0.0	0.0
Cd concentration in	12 7	18 9	5.0	14.0
adipose tissue (ng/kg)	12.1	10.0	0.0	17.0
WOMEN SPECIFIC	Mean	Median	25 <sup>th</sup>	75 <sup>th</sup>
Number of pregnancies	2.8	3.0	2.0	4.0





## Figure 1. Adipose tissue Cd concentrations in the study population (µg/kg)

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To the best of our knowledge, this is one of the very first epidemiogical studies exploring Cd concentrations in human adipose tissue. Qin et al. (2010) observed mean adipose tissue Cd concentrations of 0.47 µg/kg in patients with uterine leiomyoma and of 0.38 µg/kg in a control group. The concentrations in our population are higher, possibly due to the strong differences between the populations under study in terms of clinical background but also in the study regions, diet and lifestyle of Southern Spain vs. Hong Kong, where the study of Qin et al. (2010) was conducted.

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Our results of detectable levels of Cd in all the study participants are in agreement with previous studies in other biological matrices. Previous researches with healthy volunteers found Cd 5<sup>th</sup>-95<sup>th</sup> percentiles of 0.15-2.04  $\mu$ g/L in blood, 0.01-0.05  $\mu$ g/L in plasma, 0.06-0.79  $\mu$ g/L in urine, 0.004227 0.17 ng/mg in hair (Goullé, 2005). Other authors report reference levels (95th percentile) of 0.55 228  $\mu$ g/L in blood, 0.32  $\mu$ g/L in urine, 0.41 ng/mg in hair, 0.018 ng/mg in nails and 0.43  $\mu$ g/L in saliva 229 (Tirado et al., 2015). Our results in adipose tissue (12.66  $\mu$ g/kg) are extremely difficult to compare 230 to those obtained in other matrices. Differences are expected according to the matrix composition 231 (with different affinities of Cd to certain molecules), sample accessibility and available amount, etc.

233 Cd is considered a persistent and bioaccumulable element. After ingestion and absorption, Cd is mainly transported bound to certain low-molecular-weight thiols, such as metallothionein and 234 glutathione (Zalups and Ahmad, 2003). Around one third of Cd intake is transported to the kidneys, 235 236 where Cd-MT complex is filtrated in the glomerulus and is reabsorbed in the proximal tubule, remaining in the tubulus cells for years (Cucu et al., 2011; Sabolic et al., 2010). Indeed, only 237 0,007% of the body burden is excreted via urine and 0.009% via feces (Kjellström and Nordberg, 238 1978). Cd is also partly secreted into the biliary tract in the form of Cd-Glutathione but, after 239 enzimatic degradation to Cd-Cysteine complexes, Cd is reabsorbed in the small intestines. Liver 240 and kidneys accumulate an estimated 50% of Cd body burden, while the rest is distributed widely 241 in other organs and tissues (Hammond and Beliles, 1980). Cd can also form large aggregates with 242 lipids (Kerek et al., 2017). 243

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#### **3.2. Predictors of adipose tissue Cd concentrations**

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Table 2 shows the results from the multivariable analysis of the potential predictors of adipose tissue Cd concentrations.

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Table 2. Predictors of log-transformed adipose tissue Cd concentrations (µg/kg) in GraMo cohort (men and women). Multivariable linear regression analysis. Standard Error p-value Beta Sex/gender=male -0.424 0.106 <0.001 Age (years) 0.014 0.003 < 0.001 BMI (Kg/m<sup>2</sup>) -0.038 0.009 < 0.001

Smoking habit <sup>1</sup>			
Former smoker	0.502	0.124	<0.001
Current smoker	0.669	0.127	<0.001
Egg consumption²≥2 portions/week	0.241	0.107	0.025
Meat consumption <sup>3</sup> >2 portions/week	-0.210	0.105	0.046
Fish consumption⁴≥2 portions/week	0.160	0.108	0.140
Self-perceived exposure to paints	0.175	0.123	0.156

- Multiple R<sup>2</sup>=0.2596; adjusted R<sup>2</sup>= 0.2284; p =<0.0001
- 252 <sup>1</sup> Ref. category: non-smoker

<sup>2</sup> Ref. category: egg consumption < 2 portions/week

<sup>3</sup> Ref. category: meat consumption ≤2 portions/week

<sup>4</sup> Ref. category: fish consumption <2 portions/week

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258 Smoking habit emerged as a substantial predictor of Cd adipose tissue concentrations, in an apparently dose-response manner, with the current smokers showing the highest exposure levels 259 260 (2-fold higher levels than non-smokers), followed by former smokers (1.7-fold higher levels). This 261 finding is in agreement with previous investigations in other biological matrices, e.g. in blood, 262 where smokers have been reported to exert even 4-5-fold higher Cd concentrations than nonsmokers (Järup et al., 1998). Actually, each cigarette is estimated to contain 1.7 µg of Cd 263 accumulated in tobacco leaves, and the smoker inhales up to 10% of this quantity (Morrow, 2001). 264 265 The relatively strong positive association between smoking habit and adipose tissue Cd 266 concentracions is depicted in Figure 2, with an increasing number of smokers/former smokers in 267 the upper deciles of Cd concentrations. Our results support the consideration of smoking as the 268 most relevant determinant of Cd concentrations in the body, as previously suggested by other 269 researchers (Marano et al., 2012; Satarug and Moore, 2004). In fact, Ratelle et al. have recently 270 identifed smoking as the main predictor of Cd concentrations in urine and blood in Northwestern Canadian communities population, even more important than a diet rich in animal organs, e.g. liver 271 272 and kidney (Ratelle et al., 2018).





Figure 2. Smoking habit distribution according to deciles of adipose tissue Cd concentrations.

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In the present study, we evidenced a positive association between age and Cd concentrations in 279 280 adipose tissue. Despite the very limited previous research on adipose tissue, this finding is 281 consistent with the bioaccumulative characteristics of this metal (Kjellström and Nordberg, 1978). Previous mathematical models in adult women predicted a lifespan monotonic curvilinear increase 282 of urinary Cd with age until it leveled off at the age of 60-70 (Amzal et al., 2009). A similar plateau, 283 around the age of 50, was found by Satarug et al., consistent with the degeneration of the kidney 284 285 reabsortion function (Satarug et al., 2002). Results showing the increase of urine Cd with age have been reported in the United States, Canada, China, Korea and Spain (USCDC, 2009; Health 286 Canada, 2010; Sun et al., 2016; Lee et al., 2012; Huang et al., 2013; López-Herranz et al., 2016). 287

289 In our opinion, the abovementioned findings in the multivariable models point to a certain degree of potential Cd bioaccumulation in adipose tissue, i.e. higher levels in the older participants as well as 290 increased concentrations in former smoker vs non-smokers. Further research is warranted to 291 confirm the existance and the extent of the potential bioaccumulation of Cd in adipose tissue, as 292 293 well as to characterize if there is any degree of speciation and interaction with other metals, that could affect the biological implications of Cd concentrations (Yokel et al., 2006). For example, Cd 294 can displace Zn already binded to MT to form Cd-MT (Funk et al., 1987), which could influence Zn 295 bioaccumulation (Smidt et al., 2007). Furthermore, the complex Cu-Glutathione has the potential 296 to remove both Zn and Cd from their MT-binded forms (Ferreira et al., 1993). In addition, co-297 298 exposure to Cd and Cr, or Cd and Pb, has been reported to reduce the glomerular filtration rate even more than when the exposure is limited to the individual TEs (Tsai et al., 2017). 299

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Obesity was negatively associated Cd concentrations, being consistent with the inverse associations found with Cd concentrations in urine (Padilla et al. 2010) and in blood (Garner and Levallois, 2016). A dilution effect might partially account for this observation, so that Cd concentrations in the adipose tissue would be diluted at very high levels of obesity, resulting in negative associations between BMI and the concentration of certain pollutants in adipose tissue (Arrebola et al, 2014; Wolff et al., 2000).

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Another feasible explanation of the negative association with BMI involves alterations found in the 308 regulation of MT gene expression in obese and type 2 diabetes patients, and their influence on Zn 309 concentrations (Do et al., 2002; Haynes et al., 2013; Szrok et al., 2016). Although results on this 310 311 issue are not totally clear yet, similar mechanisms could alter Cd concentrations in adipose tissue, as Cd can easily displace Zn binded to MT. Indeed, the median BMI in GraMo cohort was 26.70, 312 which is in the range of overweight as defined by the WHO (WHO, 2018). However, Cd has also 313 314 been acknowledged as a potential obesogen and, therefore, capable of inducing obesity and obesity-related diseases (Green et al., 2018). Future research should include other obesity 315 measures, such as Waist Circumference, Waist-to-Hip Ratio, or Bioelectric Impedance, which 316 could shed more light on this issue. 317

In our study we also evidenced some dietary predictors of Cd concentrations, although their influence was less evident in comparison to the rest of variables. We found a positive association of Cd adipose tissue concentrations with egg and meat consumption, but no statistically-significant association was found with the intake of vegetables and legumes. Previous studies have reported a major contribution of grains (26.9%) to dietary Cd intake in Europe, followed by vegetables (16.0%) and meat (7.7%) (EFSA, 2012).

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Meat consumption is considered a relevant source of Cd exposure based on the sorption and retention of Cd<sup>2+</sup> by amine groups in meat proteins (Lopes et al., 2007). Indeed, a previous study performed in Tenerife Island (Spain), González-Weller et al. (2006) identified meat as a relevant exposure source. In contrast, we found a negative association in our population that needs to be further investigated, considering the specific types of meat products as well as the potential residual confounding effect of unmeasured variables.

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Furthermore, our findings of a positive association between egg consumption and Cd concentrations was somehow unexpected. The grain-based diet of hens and the content of fat and albumin (which exerts high affinity to Cd) might explain this finding. The absorption of persistent pollutants by chickens and its excretion through eggs has been a matter of concern for decades (Van Eijkeren et al., 2006; Lovett et al., 1998). Further research is needed on this issue. In fact, metal levels in bird eggs have been previously used as a marker of environmental contamination (Tsipoura et al., 2011; Gochfeld, 1997).

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Despite the recent increasing concern regarding fish as a potential source of metal exposure (Perera et al., 2015), we found a positive but non significant association of Cd adipose tissue concentrations with the frequency of fish consumption. This might be related to the fact that Cd concentrations in saltwater fish are lower than in fresh water fish. This is because Cd combines with chlorides to form CdCl<sub>2</sub>, with less bioavailability for fish (the best form for direct uptake through fish gills is the free ion Cd<sup>2+</sup>, which is more abundant in low salinity water) (Perera et al., 2015).

347 Although we did not gather information on the fish origin, our study population is based on the Mediterranean coast, so that it is expected that saltwater fish is predominant in their diets. 348 Noteworthy, because of its geological history, the Mediterranean Sea has an increased salinity in 349 350 comparison to the Atlantic and other oceans in the world (lorga and Lozier, 1999). Furthermore, 351 large variability in the content of persistent pollutants in fish can be detected among fishing sites due to the differences in the exposure of sea fauna (Nicklisch et al., 2017), which might induce 352 increased variability and, therefore, a certain degree of exposure misclassification. On the other 353 side, previous studies have reported a predominant consumption of white fish (49.5% women, 354 61.2% men) in inhabitants of Granada province, and a relatively low consumption of crustacea 355 (18.0% woman, 12.1% men) or fatty fish (32.5% women, 26.5% men) (Welch et al., 2002), the 356 latter traditionally considered to have a higher metal content (Olmedo et al., 2013). Indeed, our 357 study population declared a higher consumption of white than blue fish (Table 1). 358

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In the multivariable models, males showed lower Cd concentrations than females. This is in 360 agreement with previous studies of Cd concentrations in blood, urine and kidneys (Vahter et al., 361 2007; López-Herranz et al., 2016). These differences might be related to sex- (physiological) 362 363 and/or gender-specific aspects (sociological). Interestingly, women at fertile age frequently show 364 relatively decreased iron stores, which has been reported to induce gastrointestinal absorption of Cd (Åkesson et al., 2002). In order to elucidate sex/gender-related differences in the predictors of 365 the exposure, as well as the influence of women-specific characteristics, we further stratified the 366 multivariable analyses by sex (Table 3), and tested women-specific variables in the models for 367 368 females. We did not evidence any significant association of the number of pregnancies, number of 369 children, months of breastfeeding, or menopausal status with Cd concentrations.

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Interestingly, those males that declared a frequent direct exposure to paints (including occupational and non-occupational exposure) showed increased Cd concentrations, although the association was only marginally significant. Other authors observed increased urinary Cd concentrations in individuals working with paints (Awodele et al., 2014), but the exposure can also be domestic since Cd is a common component in a number of pigments on the market (Faulkner and Schwartz,

376 2009). We believe that this finding also relates to socio-occupational aspects in the study region. In our study population, 29 men (23%) and only 14 women (14%) declared a frequent exposure to 377 paints, which might hamper the finding of significant associations in women. However, the model 378 coefficient in females was very far from the one in males, which also points to differences in the 379 380 perception of the exposure among genders. Indeed, from those declaring frequent exposure, 7 (24%) men and only 1 (7%) woman were involved in occupations implying a daily handling of 381 paints (professional painters). This could explain an increased occupational exposure in the male 382 population. 383

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# Table 3. Predictors of log-transformed adipose tissue Cd concentrations (µg/kg) in GraMo cohort stratificated by sex/gender. Multivariable linear regression analysis.

		Men			Women	
	Beta	Standard error	p-value	Beta	Standard error	p-value
Age (yrs)	0.017	0.004	<0.001	0.010	0.005	0.065
BMI (kg/m²)	-0.029	0.013	0.021	-0.041	0.015	0.008
Smoking habit <sup>1</sup> :						
Former smoker	0.491	0.181	0.008	0.440	0.202	0.032
Current smoker	0.642	0.179	<0.001	0.640	0.195	0.001
Egg consumption²≥2 portions/week	0.210	0.151	0.167	0.197	0.161	0.226
Meat consumption <sup>3</sup> >2 portions/week	-0.162	0.149	0.280	-0.282	0.150	0.063
Fish consumption⁴≥2 portions/week	0.237	0.149	0.114	0.056	0.165	0.735
Self perceived	0.287	0.165	0.085	-0.009	0.204	0.966

	exposure to paints					
386	Men: multiple R <sup>2</sup> =0.3011;	adjusted R <sup>2</sup> =0.2529;	p<0.001. Women:	multiple R <sup>2</sup> =0.1981;	adjusted R <sup>2</sup>	=0.1260;
387	p=0.009					

388 <sup>1</sup> Ref. category: non-smoker

389 <sup>2</sup> Ref. category: egg consumption < 2 portions/week

<sup>3</sup>Ref. category: meat consumption<=2 portions/week

<sup>4</sup> Ref. category: fish consumption<2 portions/week

392

Our findings reveal a generalized exposure to Cd in the study population, and highlight the 393 potential relevance of adipose tissue Cd concentrations as a biomarker for exposure assessment. 394 395 Despite our hospital-based cohort may not be be entirely representative of the general population, 396 as well as the cross-sectional design, which is susceptible of reversed-causality, most of the reported associations seem robust and consistent with previous research. Although this matrix 397 might pose limitations in comparison to other biological compartments in terms of accessibility, we 398 399 believe that adipose tissue Cd concentrations can provide relevant results. In this regard, the analysis of adipose tissue can shed light on the potential health effects of Cd, and can complement 400 previous results in relevant matrices, such as urine or blood, which might have different biological 401 meanings. Indeed, previous studies have suggested potential associations of long-term Cd 402 403 exposure with the induction of oxidative stress (Liu et al., 2009), inflammation (Riemschneider et al., 2015), and endocrine disruption (Takiguchi and Yoshihara, 2006), as well as further 404 development of certain obesity-related pathologies, such as metabolic syndrome (Tinkov et al., 405 2017) and cancer (Johnson et al., 2003), in which adipose tissue is believed to play an important 406 (though not fully understood) role. 407

408

Further research is warranted on the clinical relevance of the adipose tissue Cd concentrations, as well as on specific aspects of their use as an exposure biomarker, such as the potential speciation in the matrix and concentrations in other populations from different regions. Considering previous studies, potential interactions between Cd and other trace elements should be taken into account in the assessment of possible health effects (Tsai et al., 2017).

414

415 **4. Conclusions** 

416

We detected Cd in all the adipose tissue samples from the study population, and identified certain predictors of the exposure, such as age, sex/gender, BMI, smoking habit, eggs and meat consumption and exposure to paints (the latter only in men). This points to a potential relevance of adipose tissue Cd concentrations for exposure characterization as well as for the assessment of long-term effects of chronic exposure to low levels of Cd, which are currently being studied in GraMo cohort.

423

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425

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433

#### 434 **Conflict of interest**

435

- 436 The authors declare no conflict of interest.
- 437

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