

Alkaline phosphatase activity related to phosphorus stress of microphytoplankton in different trophic conditions.

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Running title: Microphytoplankton adaptation to P-stress

1 *Abstract*

2 The northern Adriatic (NA) is a favorable basin for studying the adaptive strategies of
3 plankton to a variety of conditions along the steep gradients of environmental parameters over
4 the year. Earlier studies identified phosphorus (P)-limitation as one of the key stresses within
5 the NA that shape the biological response in terms of biodiversity and metabolic adjustments.
6 A wide range of reports supports the notion that P-limitation is a globally important
7 phenomenon in aquatic ecosystems. In this study P stress of marine microphytoplankton was
8 determined at species level along a trophic gradient in the NA. In P-limitation all species with
9 considerable contributions to the diatom community expressed alkaline phosphatase activity
10 (APA), compared to only a few marginal dinoflagellate species. Nevertheless, APA
11 expressing species did not always dominate the phytoplankton community, suggesting that
12 APA is also an important strategy for species to survive and maintain active metabolism
13 outside of their mass abundances. A symbiotic relationship could be supposed for diatoms
14 that did not express APA themselves and probably benefited from APA expressed by attached
15 bacteria. APA was not expressed by any microphytoplankton species during the autumn when
16 P was not limiting, while most of the species did express APA during the P-limitation. This
17 suggests that APA expression is regulated by orthophosphate availability. The methods
18 employed in this study allowed the microscopic detection of APA for each
19 microphytoplankton cell with simultaneous morphologic/taxonomic analysis. This approach
20 uncovered a set of strategies to compete in P-limited conditions within the marine
21 microphytoplankton community. This study confirms the role of P-limitation as a shaping
22 factor in marine ecosystems.

1. Introduction

Growing evidence suggests that phosphorus (P) is frequently the limiting nutrient in coastal systems (Thingstad et al., 1993; 1998) and oligotrophic oceans (Cotner et al., 1997; Karl and Yanagi, 1997). Furthermore, it is believed that the open ocean, far from the continental inputs of nutrients, should evolve towards P-limited conditions due to N₂ fixation, while P is a non-renewable limiting nutrient (Tyrrell, 1999). The rise in N₂ fixation and the simultaneous disappearance of P stocks observed in the subtropical North Pacific Ocean between 1989 and 2004 is an illustration of this phenomenon (Karl et al., 2001, Karl, 2007). Therefore, the importance of alkaline phosphatase activity (APA) with regard to the transformation and turnover of organic compounds in marine environments has been investigated with growing attention (for review see Hoppe, 2003; Yamaguchi and Adachi, 2010).

A number of studies in the northern Adriatic (NA) evidenced that this region is currently P-limited (citations reported in the following text). This shallow (up to 50 m) coastal sea is characterized by significant freshwater input, mainly from the Po River. Although in the Po River waters both nitrogen (N) and P concentrations (total nitrogen 263 $\mu\text{mol L}^{-1}$ and total phosphorus 4.8 $\mu\text{mol L}^{-1}$; Cozzi and Giani, 2011) are more than one order of magnitude higher than in the NA waters, the inorganic N:P molar ratio (about 84:1; Cozzi and Giani, 2011) provides a strongly unbalanced N versus P supply for phytoplankton requirements (presumed balanced N:P=16:1; Redfield *et al.*, 1963). Earlier studies in this area showed that organic phosphorus concentrations markedly exceeded orthophosphate (PO₄) concentrations, representing an important source of P for microbial communities (Ivančić and Degobbis, 1987). A significant contribution to the pool of dissolved P in the NA could be due to the release of dissolved DNA by nanoflagellates grazing on nucleic acid-rich bacterial biomass (Turk et al., 1992). These authors reported that during stratified conditions a significant part

of the P demand by the autotrophs may be sustained by the rapid turnover of DNA-mediated release of inorganic phosphorus. The importance of dissolved organic phosphorus (DOP) in the area during periods of P-limitation was evidenced also by a recent mesocosm experiment (Malfatti et al., 2014). However, the authors pointed out that not all DOP is directly hydrolysable by alkaline phosphatase (AP) and multiple hydrolase activities were available with potential to transform various organic matter pools into molecules accessible to AP and other hydrolases that are the terminal step in P regeneration.

Bioassays based on selective nutrient additions and nutrient molar ratios identified P-limitation of the NA (Maestrini et al., 1997 and citation therein). Enzyme assays with soluble substrate showed that extracellular AP is important in providing P for phytoplankton growth, particularly during blooms induced by freshwater imported nutrients (Ivančić et al., 2009; 2010). However, although measurements with soluble substrate allowed detection of the phytoplankton P status, they provided only the APA of the whole community and did not yield information on the individual species. Furthermore, it was not possible to mark the exact location of the enzymatic activity in the cells or to determine the contribution of each species to the overall enzyme signal.

In the present study an AP substrate with insoluble fluorogenic product, termed ELF (Enzyme Labelled Fluorescence), was used to investigate how the APA status varies among phytoplankton species at two locations along the trophic gradient. This method yields a stable, highly fluorescent precipitate at the site of enzyme activity and thus has the capability to determine the APA status of individual cells (González-Gill et al., 1998). Hence, it can give insights into the mechanisms of P nutrition of phytoplankton in natural waters. In the previous study it was found that low PO_4 concentrations can modify the structure of plankton communities and constrain the phytoplankton distribution (Ivančić et al., 2012) as also reported for other areas (Egge, 1998; Moutin et al., 2005; Mackey et al., 2007). In addition P-

limitation can affect phytoplankton abundance and composition, viral and prokaryotic abundance, microbial activity and organic matter production and utilization (Larato et al., 2010). However, not all microorganisms are equally affected. By comparing estimated taxon-specific C:P utilization rates with particulate C:P ratios in P-replete and P-deplete cultures it was shown that different phytoplankton groups experienced different degrees of P stress in the same nutrient environment (Casey et al., 2009).

In this study, investigations are concentrated on microphytoplankton as it is often the dominating component of the phytoplankton community biomass in the NA (Center for Marine Research, unpubl. data) and it is taxonomically well defined. Microphytoplankton also seems to play the major role in benthic pelagic coupling in the area (Pfannkuchen et al., 2009; 2010). This study aims to answer the questions: do all microphytoplankton species unequivocally contribute to the bulk APA? Or perhaps, are there alternative strategies to cope with the gradients of P-availability in the NA? Another covered topic is the behavior in the planktonic community of (i) species which do follow the observed bulk APA and (ii) species which follow alternative strategies to cope with P-limitation.

2. Methods

2.1. Sampling strategy

Measurements were carried out at two stations (SJ101, RV001) in the northern Adriatic (Fig. 1) during eight cruises performed from March to November 2013. These stations were specifically chosen as they are in areas with different nutrient regimes. SJ101 (bottom depth 32 m), situated in the western area, is usually under riverine nutrient pressure while RV001 (bottom depth 27 m), situated in the eastern coastal area, is only intermittently experiencing freshwater influence. Further, SJ101 is usually under more direct freshwater nutrient influence, while freshwater reaching RV001 is already impoverished from nutrients.

Fig. 1

During all cruises sea temperature and salinity were determined with an SBE25 conductivity-temperature-depth probe (Sea-Bird Electronics, Washington, USA). Samples for analysis of nutrients, bulk APA, and phytoplankton were collected with 5 L Niskin bottles at three depths (surface i.e. 0-1 m, 10 m, and 2 m above the bottom). Samples for the determination of single cell APA (ELF assay) were collected by vertical tow of the phytoplankton net (pore size 20 μm) from 10 to 0 m depth (i.e. upper waters). Data for the daily Po flow mean measured at Pontelagoscuro, Italy, were kindly supplied by Assessorato Programmazione, Pianificazione e Ambiente of the Emilia Romagna region (Italy).

2.2. *Analytical protocol*

Inorganic nutrient analyses were performed onboard, immediately after sample collection, using methods described earlier (Strickland and Parsons, 1972; Ivančić and Degobbis, 1984). Samples for total dissolved phosphorus were filtered (Whatman GF/F, precombusted at 500 °C) and stored in autoclavable polypropylene tubes at -30 °C. In the laboratory ashore analyses were performed using a wet combustion oxidation method with persulphate (Menzel and Corwin, 1965). DOP was calculated by subtracting PO_4 from the total dissolved phosphorus. Dissolved inorganic nitrogen (DIN) was calculated as the sum of nitrate, nitrite and ammonia.

Determination of APA was performed aboard the research vessel immediately after sample collection. Sea water (5 mL per one sample) was filled into plastic syringes directly from samplers and filtered through a 200 μm mesh (to remove mesozooplankton), 3 μm (polycarbonate Nuclepore filter, Whatman) and 0.22 μm filters (mixed cellulose esters Millipore membrane filter, Millex-GS). All filtrations were performed manually and gently applying light pressure using manual Millipore filter units. Determined activity in the 0.22 μm filtrate was very low and in upper waters often below the detection limit, thus indicating that

gentle filtration did not cause cell rupture. Filters of 3 μm were preferred (passing through both bacteria and picocyanobacteria) instead of 2 μm at which an unknown part of picocyanobacteria is retained. The nanophytoplankton fraction was retained on the 3 μm filter as confirmed by microscopic inspection. The phytoplankton APA was calculated by subtracting activity in the 3 μm from the activity in the 200 μm filtrate.

All APA measurements were performed using the fluorogenic substrate analogue methylumbelliferyl-phosphate (MUF-P) dissolved in methylcellosolve and diluted with water immediately before addition, following the procedure of Hoppe (1983). Aliquots of 5 mL of all the filtrates, in duplicate, were used for APA measurements. The final concentration of the substrate in the samples was 50 $\mu\text{mol L}^{-1}$ (Ivančić et al., 2009). Incubation of the samples was performed in dark in baths filled with water from the sampling depths, i.e. at *in situ* temperature. Fluorescence was measured immediately after substrate addition and after ~1h of incubation using a Turner Designs-700 fluorometer with excitation at 365 nm and emission at 460 nm. APA ($\text{nmol L}^{-1} \text{h}^{-1}$) was calculated as the difference between those measurements divided by the incubation time after calibration of the fluorometer with methylumbelliferone, the product of MUF-P degradation. Additionally, in the 200 μm filtrate, single measurements were performed using various MUF-P concentrations: from 0.5 $\mu\text{mol L}^{-1}$ to 50 $\mu\text{mol L}^{-1}$. Those measurements were used to determine the half saturation constant (K_m) and maximum activity (V_{max}) of the Michaelis-Menten kinetics, employing the Wolf-Hanes linearization. P turnover time (T) was estimated by the $K_m:V_{\text{max}}$ ratio (Labry et al., 2005).

Single cell APA was detected utilizing the ELF[®]97 Endogenous Phosphatase Detection Kit (E6601) (Thermo Fisher Scientific, Waltham, USA). Cells were fixed in 3.5% buffered (pH 8) formaldehyde at 4 °C for 30 min and not permeabilized with detergents as suggested by the ELF-kit manufacturer. Afterwards, cells were rinsed in sterile seawater and subjected to microscopic analysis. ELF substrate was diluted 20 fold in detection buffer and

mixed with fixed cells to reach a final dilution of 40 fold, directly prior to microscopic examination. Chloroplast autofluorescence was recorded at 580-600 nm (555 nm excitation) and ELF signal at 550-555 nm (405 nm excitation) on a Zeiss (Oberkochen, Germany) LSM (laser scanning microscope) 700 setup. Nominal thickness of optical sections under confocal conditions was 1.2 μm . Thresholding signal intensity and three dimensional signal reconstructions allowed to identify cell surface associated signal after the analysis of axial cross sections and correction for signal point spread functions (ZEN black, Zeiss). When an ELF positive cell was found, the species was determined and 99 additional intact cells of the same species were examined for ELF signal. Cells with leaking cytoplasm or without chloroplasts were disregarded. Each cell was tallied as either positive or negative for ELF labelling on the basis of the presence or absence of the fluorescent ELF[®]97 precipitate. Cells were regarded as ELF positive as soon as an ELF signal (signal intensity above the threshold as determined from the negative controls without addition of substrate) was located on the cell surface. The average standard error for triplicate counts for 100 cells using the ELF[®]-97 labelling technique was determined to be 3% (Dyhrman and Palenik, 1999), and a similar error is expected in the present study. The respective figures show tangential optical sections best representing cell surface associated ELF signal.

Samples for the determination of phytoplankton composition and relative abundance were filtered through a 200 μm mesh to remove zooplankton, and filtrates were preserved with formaldehyde solution (2% final concentration) buffered with sodium acetate. After 38 h sedimentation of 50 mL of filtrate, cell counts were performed on an inverted Axiovert 200 microscope (Zeiss GmbH) following the Utermöhl method (1958). During counting, phytoplankton cells were attributed to microplankton or nanoplankton fractions, based on observed cell dimensions (Sieburth et al., 1978) and counted at 200 X and 400 X magnifications, respectively. Phytoplankton cells were identified at the lowest possible

taxonomical rank. Diatoms and dinoflagellates within microphytoplankton were determined on the level of species.

Comparison of the contribution of ELF labelled organisms in microphytoplankton communities with bulk APA in fraction $> 3 \mu\text{m}$ was tested using least square linear regression and principal component factor analysis (PCA). One-way ANOVA was performed in order to test if there was a difference between the contribution of ELF labelled species in P-limited and P-repleted conditions. The conditions of normal distribution were tested with the Shapiro-Wilk test.

3. Results

3.1. Hydrological conditions, nutrient status and phytoplankton community composition

From March to the end of June the Po River flow was high (on average $3023 \text{ m}^3 \text{ s}^{-1}$) with maximal impulses up to $7120 \text{ m}^3 \text{ s}^{-1}$ in May (Fig. 2A). At the end of June the Po flow decreased to typical summer values and remained relatively constant until the end of October (in average $1070 \text{ m}^3 \text{ s}^{-1}$), with frequent small impulses (up to $1437 \text{ m}^3 \text{ s}^{-1}$). From the end of October the Po flow increased up to $2743 \text{ m}^3 \text{ s}^{-1}$ at the end of November.

Fig. 2

In March the entire water columns at both stations were cold ($9.78\text{--}10.65^\circ\text{C}$) and almost homogenous (Fig. 2B). During April–August, temperatures gradually increased ($13.21\text{--}24.67^\circ\text{C}$ at the surface and $10.05\text{--}17.70^\circ\text{C}$ at the bottom) with the establishment of thermal stratification at both stations, being the sharpest between the depths of 12 and 20 m. From September a progressive cooling of the surface waters ($21.84\text{--}14.15^\circ\text{C}$; Fig. 2B) started a progressive mixing in the water columns. At the end of November temperature of the water columns was homogenous.

At SJ101 freshwater influence, extending down to 10-15 m, was considerable during the whole studied period (surface salinity 15.90-36.54) (Fig. 2C). At RV001, freshwater influence was detected during May-October (surface salinity 35.82-37.00) extending down to 5-15 m, while in March-April and November it was low (surface salinity 37.21-37.54; Fig. 2C). In deeper layers of both stations more saline water was observed during the whole investigation period (salinity 37.20-38.41).

PO₄ concentrations at RV001 were exceptionally low ($< 0.03 \mu\text{mol L}^{-1}$; Fig. 3A; S1A) in the entire water column. At the surface of SJ101 somewhat higher PO₄ concentrations were found during freshets in early spring and autumn (March-April, October-November; $0.07\text{-}0.17 \mu\text{mol L}^{-1}$), while during the summer values were exceptionally low ($< 0.03 \mu\text{mol L}^{-1}$). At this station a moderate accumulation of PO₄ in the bottom waters started with the onset of stratification in May and persisted during the summer and autumn ($0.05\text{-}0.16 \mu\text{mol L}^{-1}$; Fig. S1A). DOP concentrations ($0.10\text{-}0.59 \mu\text{mol L}^{-1}$) always exceeded those of PO₄ (Fig. 3A; S1B) at both stations. Lower DOP concentrations were characteristic for summer and autumn, while during spring they were markedly higher.

Fig. 3

During summer, DIN concentrations ($0.77\text{-}1.18 \mu\text{mol L}^{-1}$) in upper waters of both stations (Fig. 3A, S1C) were markedly lower than in spring and autumn ($1.75\text{-}4.28$ and $2.57\text{-}76.15 \mu\text{mol L}^{-1}$ at RV001 and SJ101, respectively), with the highest values during the spring freshets at SJ101. At SJ101, a moderate accumulation of DIN was observed in the bottom waters during the summer (up to $5.88 \mu\text{mol L}^{-1}$), while at RV001 no significant variation with depth was observed (S1C).

At the surface and 10 m depth at both stations DIN was present in marked surplus with respect to PO₄ (average inorganic N:P ~ 200). In bottom waters they were generally more balanced for phytoplankton requirements (average inorganic N:P ~ 21).

The phytoplankton abundance was in the range of $\sim 10^4$ - 10^6 cell L⁻¹ (Fig. 3B; S2). The highest abundances were found in the surface waters and generally decreased with depth. The highest abundances were found from March to May in upper waters of SJ101 (up to 2.0×10^6 cell L⁻¹), generally for an order of magnitude higher than at RV001. From July to November the abundances were generally similar at both stations (5.2×10^4 - 4.8×10^5 cell L⁻¹). In the bottom waters, phytoplankton abundance (2.8×10^4 - 1.5×10^5 cell L⁻¹) was lower than in upper waters and without noteworthy changes during the year or by station (Fig. S2C).

Generally, the phytoplankton community at both stations was dominated by the microphytoplankton in the upper waters, while the nanophytoplankton was dominating at the bottom (Fig. 3B; S2). Diatoms were generally much more abundant than dinoflagellates. In the microphytoplankton, diatoms and dinoflagellates were the two most diverse and abundant groups, while other groups accounted for < 3% of the respective community. During spring (March-April) the community composition in upper waters differed between the two stations (Fig. 3B). Microphytoplankton, comprised of almost only diatoms, strongly predominated at SJ101 (91%), while at RV001 domination of nanophytoplankton (75-98%) was recorded. In May-November the phytoplankton community composition was rather similar at both stations. During May and July nanophytoplankton dominated in the community, especially at SJ101 (60-80%). Dinoflagellates during this period reached their maximal contribution at both stations (10-22% and 6-16% at RV001 and SJ101, respectively). During August-November diatoms strongly predominated (generally 70-90%), while dinoflagellates accounted for a minor part of the community (0.2-4.6%).

3.2. APA and P turnover time

In upper waters bulk APA gradually increased from March to high summer values (Fig. 4A; S3; up to 362.8 and 2916 nmol L⁻¹ h⁻¹, at RV001 and SJ101, respectively). During these months, values at SJ101 were constantly higher than at RV001, especially in March

when APA at RV001 was exceptionally low ($< 10 \text{ nmol L}^{-1} \text{ h}^{-1}$). In autumn (October-November) a decrease in APA was observed (18.8 and $53.8 \text{ nmol L}^{-1} \text{ h}^{-1}$ at SJ101 and RV001, respectively). In upper waters most of bulk APA was observed in the fraction $> 3 \mu\text{m}$ (phytoplankton fraction; generally 70-90%), except in November when most of the activity was found in the fraction $< 3 \mu\text{m}$ (Fig. 4A). At the bottom, APA was generally for an order of magnitude lower than in upper waters (Fig. S3C).

Fig. 4

APA-mediated PO_4 turnover time (T) at the bottom was generally far above 5 h (up to 135 h; 4B), indicating that phytoplankton was not P-limited (Nausch et al., 2004; Xu et al., 2008). On the contrary, in upper waters T was generally below 5 h, with exceptionally low values in August at the surface of SJ101 (about 1 min), indicating P-limitation. In surface waters $T > 5 \text{ h}$ were observed only in March at RV001 (about 63 h), then in October and November at both stations (17-257 h), indicating that P-limitation did not occur.

ELF signal was found localized around associated bacteria (Fig. 4Ca), diffused across the cell surface (Fig. 4C b), in irregular shaped structures on the cell surface (Fig. 4C c), in regular shaped islands on the cell surface (Fig 4C d), in small aggregations across the cell surface (Fig. 4C e,f), diffusely associated with chloroplasts but on the cell surface (Fig. 4C g) and in regular shaped cell surface islands associated to chloroplast (Fig. 4C h).

In the microphytoplankton community a variety of diatoms showed a positive ELF signal (Table 1). The lowest contribution of ELF positive species in diatom community (about 15-38%) was found during spring at RV001 (Fig. 5A,B).

Fig. 5

In March, during the water column mixing, only few species in relatively low abundance (*Chaetoceros constrictus*, *Cerataulina pelagica*, *Thalassiosira* sp.), and in April

the dominant *Chaetoceros circinalis*, showed ELF signal at RV001 (Fig. 5A,B). *Chaetoceros brevis* was not ELF labelled, but bacteria attached to it showed strong ELF signal (Fig. 5A). Neither the dominant species *Bacteriastrum furcatum* in March nor the co-dominant *Cerataulina pelagica* in April were ELF labelled. Contrary, during the spring freshets at SJ101 the dominant *Skeletonema marinoi*, as well as the major part of diatom species (up to 99.5% in diatom community), showed ELF signal (Fig. 5A,B). The co-dominant *Chaetoceros socialis* was ELF positive in April, but not in March when bacteria attached to it had a strong positive signal. In May the diatom community was strongly dominated by ELF labelled species (about 83-91%) at both stations (Fig. 5C). During summer the maximal number of ELF labelled species was found (Fig. 6). The dominant species were ELF labelled, with exception of *Chaetoceros* sp. in August at SJ101. However, at both stations, during summer a variety of species with relatively high contribution to the diatom community did not show any ELF signal.

Fig. 6

A variety of species with minor contribution and no detectable ELF signal (“Other not ELF labelled”, Figs. 5, 6) were found at both stations during all sampling terms (Table 1).

Dinoflagellates were scarcely ELF labelled (Table 2). The labelled species made up 3-10% of the dinoflagellate community in June and August, but generally labelled species contributed less than 2% to the dinoflagellate community. Dinoflagellates in total generally contributed < 10% to the microphytoplankton abundance.

In October no ELF signal was detected in any microphytoplankton species.

The contribution of ELF positive species to the diatom community was strongly correlated with the bulk APA in the > 3 µm fraction (phytoplankton APA; $p=0.009$), in contrast to species from the dinoflagellates community ($p=0.538$). Principal Component

Analysis (Fig. 7) also indicated a tight relationship between the contribution of ELF positive diatoms and phytoplankton APA. In addition, positive relations with nitrate and temperature, and inverse relations with PO_4 , DOP and salinity were found. The contribution of ELF positive species to the dinoflagellate community was loosely related with phytoplankton APA, but showed strong positive relationship with temperature. ANOVA showed that during P-limitation the contribution of ELF positive species in both diatom and dinoflagellate communities was significantly higher than during period when P was not limiting ($p < 0.001$ and $p = 0.014$, respectively).

Fig. 7

4. Discussion

The seasonal development of bulk APA in 2013 was similar to those described in earlier studies (Ivančić et al., 2009; 2010; 2012). As observed earlier, during periods of high APA, DOP was consumed, and higher temperature and freshwater nitrate input incited, while PO_4 repressed APA expression. This strongly indicates a seasonally reoccurring pattern of APA and that DOP is an important source of P in the region. Maximal hydrolysis rates and P-limitation were found in upper water column where single cell APA in microphytoplankton was identified using the ELF assay. This assay identifies which members of the phytoplankton community express APA, offering substantially enhanced resolution along the size fractionation. This method is being increasingly used to identify single cell APA in field populations (Dyhrman et al., 2002; Lomas et al., 2004; Dyhrman and Ruttenberg, 2006; Ou et al., 2006; Girault et al., 2013).

The distribution of APA across different species was highly variable between sites and seasons. The contribution of ELF positive species to the diatom community generally related with the phytoplankton APA, and a variety of labelled species were observed during P-limitation (Table 1).

During spring freshets, the western area (SJ101) provided conditions with high nutrient availability in which diatoms dominated the phytoplankton community. However, high bulk APA and very fast APA-mediated PO_4 turnover time (hereinafter termed P turnover) implied a severe P-limitation. In such conditions ELF labelled species represented up to 99.7% of the diatom community. The dominant species *Skeletonema marinoi* uses APA to create an advantage in the competition for organic P sources. It appears safe to suggest, that expressing APA is a successful strategy for diatoms to compete for organic P under the influence of riverine nutrient input, since at that time they reached the highest abundance during the year.

Markedly different phytoplankton strategies of P acquisition were found at the eastern area (RV001). In this area, the cyclonic circulation (Artegiani et al., 1997) advects the water from the more southern parts of the Adriatic Sea, with low nutrient content and microphytoplankton abundances (Viličić et al., 2009; Zavatarelli et al., 1998). Contrary to the western area, at the time microphytoplankton community in this area does not experience pronounced fluctuations of nutrient availability. In March low bulk APA and exceptionally high P turnover time (about 63 h) showed that PO_4 was not limiting phytoplankton growth. The water column was mixed and a modest phytoplankton biomass, mainly nanophytoplankton, developed on PO_4 regenerated during the winter. In the diatom community only few larger species did already express APA (*Chaetoceros constrictus*, *Cerataulina pelagica* and *Thalassiosira* sp.), indicating them being P-stressed, or even P-limited since their abundance was low. *C. pelagica* in April did not show any APA signal, indicating that expression of APA in March was due to P-stress/limitation. Probably the other two species follow the same pattern. These larger species have probably a higher half saturation constant of PO_4 uptake (K_m) than the other present species. Larger cells require more nutrients for growth, implicating higher K_m (Eppley et al., 1969) and consequently may

be at a disadvantage in nutrient uptake competition compared to smaller phytoplankton. Oh et al. (2010) reported that K_m might be an important index of nutrient affinity as well as a threshold for APA induction. In fact, in oligotrophic conditions found in March at the eastern area, the dominating smaller species *Bacteriastrum furcatum* (60% in diatom community) did not express APA. Probably a lower K_m of PO_4 uptake and hence an adaptation for success in more oligotrophic conditions, allowed growth of *B. furcatum* without the use of APA. This species co-dominated the diatom community expressing APA in May under P-limited conditions. Consequently, in this species APA regulation according to environmental conditions is supposed. In April P-limiting conditions were observed also at the eastern site and the dominant *Chaetoceros circinalis* expressed APA to thrive on DOP. However, a variety of diatoms contributed considerably to the overall diatom abundance (about 60%) without utilizing APA. At the time overall oligotrophic conditions were observed at the eastern site. Hence most of the diatoms did not compete for DOP due to low nutrient availability, but probably slowed down their metabolisms, as supported by their exceptionally low abundances.

From May on phytoplankton communities at both sampling sites share a common history. Typically, at that time of the year the general circulation transports water from the western coast towards the eastern coast, where the current turns northwards (Russo et al., 2005; Lyons et al., 2007), considerably reducing the influence of southern Adriatic waters at the eastern area. In such circumstances, microphytoplankton developed near the western coast under considerable influence of riverine nutrient input, subsequently reach the eastern area. Consequently, both sampling sites show microphytoplankton communities under the influence of unbalanced nutrient input. At the western site, species experienced the unbalanced nutrient input directly, and under nutrient rich conditions they grew to larger numbers exhausting PO_4 . Consequently, water spreading toward the east was poor in PO_4 , but

enriched in organic matter produced at the western area. In both cases P-limitation occurred, as confirmed by very fast P turnover (down to 1 min). In such strong P-limited conditions diatoms reacted by expressing APA to sustain (May, July with low abundance), or even to further develop their abundance and activity (August, September, high abundance). As soon as these species again reach the river plume (closed circulation), they will benefit immediately (without lag phase due to dormancy or very low abundances) from the rich nutrient inputs. Additionally, diatoms are known for longer generation times, and hence they would profit from the capability of P pool formation to compete with faster growing species.

Despite of general P-limiting conditions during the summer, a variety of species that significantly contributed, or even co-dominated the diatoms community did not express APA. Except for *Chaetoceros* sp., the other species were found to express APA under at least some conditions. This indicates that those species were not P-limited when found ELF negative. They probably either thrived on intracellular P pools (most likely for species observed at the western site) or had alternative mechanisms to overcome P-limitation (particularly at the eastern site). Possible alternatives indicated by obtained results are (i) the use of PO_4 liberated by free/attached enzymes from different species and (ii) the production of non-phospholipids (Ivančić et al., 2012; Van Mooy et al., 2009). During P-limitation, *Chaetoceros socialis* and *Chaetoceros brevis* did not show APA themselves when bacteria attached to them showed a strong APA signal. Since *C. socialis* was found able to use DOP when P is limiting, it is possible that associated bacteria provided P for its growth. Nevertheless, it is not possible to exclude, that *C. socialis* (found during freshets at the western area), deposited P in intracellular pools while it was abundantly available and that at the sampling time the intracellular availability of P did not yet set on the expression of APA. Phytoplankton accumulate P in response to pulses in P supply (Andersen et al., 1991; Tantanasarit et al., 2013; Wasmund et al., 2014) and stored P can ensure numerous divisions (Silkin et al., 2013).

P accumulation in the phytoplankton community of the NA was observed in earlier experiments (Ivančić et al., 2004). Furthermore, phospholipid:non-phospholipid ratio at the investigated sites during the summer (P-limitation) was markedly lower than during November (not P-limitation) (B. Gašparović, unpubl. data). This indicates a preferential synthesis of non-phospholipids during P-limitation, as already reported for the NA (Ivančić et al., 2012).

At the investigated sites a variety of species have not yet been found ELF labelled (Table 1). Except for *Chaetoceros* sp., they were not successful during P-limitation, contributing < 3% to the diatom community. This observation again fosters the hypothesis, that APA is a necessary prerequisite for diatom species to maintain larger abundances during P-limited conditions.

After the onset of vertical mixing in October, bulk APA decreased being minimal in November when water columns were completely mixed. In these months long P turnover (up to 257 h) showed that phytoplankton was not P-limited. During the vertical mixing none of the microphytoplankton species expressed APA. Many of the most abundant species in this period (*Lioloma pacificum*, *Bacteriastrum hyalinum*, *Rhizosolenia fragilissima*, *Chaetoceros socialis*, *Thalassionema nitzschioides*, *Rhizosolenia imbricata*, *Skeletonema marinoi*, *Cyclotella* sp.) on the contrary did express APA during the P-limitation in summer. That means there is so far no evidence for APA to be expressed constitutively. APA is rather expressed only when P is limiting, while under other conditions APA is down regulated or not expressed. It seems that during this period the constant supply of PO_4 from the bottom prevented P-limitation, as observed earlier (Ivančić, et al., 2012).

In the case of dinoflagellates, the contribution of ELF positive species to the community was low regardless of high or low hydrolytic rates or establishment of P-limited conditions. Dinoflagellates expressed APA rarely, and if so, those species had really low abundances

(Table 2). Even in May and July, when dinoflagellates approached diatom abundances, only marginal species (< 8% in dinoflagellate community) expressed APA. This indicates that dinoflagellates in the NA do not employ AP as a means for reaching large abundances. Dinoflagellates did not reach large abundances at the sampled stations, as usual for the investigated area (Godrikan et al., 2013). In contrast to our findings, in other areas a high percentage of dinoflagellates were ELF labelled as reported for the Sargasso Sea (Lomas et al., 2004), Oregon coastal area (Dyhrman and Ruttenberg, 2006), the China Sea (Ou et al., 2006) and subtropical North Pacific (Girault et al., 2013). Some of the reported species in the mentioned areas were identified also in the present study in the NA. Consequently, a low contribution of ELF labelled dinoflagellate species in the respective community in the NA probably does not reflect their inability to employ AP, but rather a so far unidentified limitation or possibility, prevent the certain dinoflagellate species from expressing APA. Their reported capability of heterotrophy might make the expression of extracellular APA redundant (Jeong et al., 2010). In fact, the Oregon coast field data indicate that cell-specific APA of dinoflagellates may not be strongly P regulated, and unlike for diatoms, the presence of cell-specific APA does not support an interpretation of P stress for dinoflagellates (Dyhrman and Ruttenberg, 2006).

5. *Conclusions and future directions*

Overall, in P-limited conditions all species with considerable contributions to the diatom community, with only a few exceptions, expressed APA, while only few dinoflagellate species expressed APA. Diatoms often dominated the phytoplankton community, showing APA to be seemingly a very important prerequisite for their success in the NA. APA-expressing species do not necessarily dominate the phytoplankton community, since species in very low relative abundances that nevertheless did express APA were found. This suggests that APA is also an important strategy for species to survive and maintain

active metabolic state outside of their mass abundances or blooms. This allows them to immediately react to short term nutrient availability (e.g. riverine input). This feature appears to be very beneficial in a complex and fast changing environment with gradients of nutrient availability and point sources for nutrient input. Findings of co-dominating species in the diatom community that did not express APA, let presume that those species store P intracellularly, especially at the western site (close to the Po plume), and at the sampling time still lived on those pools. Other species did not express APA themselves, however bacteria attached to those cells did show APA. For these species a symbiotic relationship could be supposed, where the larger host diatom cell makes use of APA expressed by the attached bacteria.

Further research should address strategies for successful competition of phytoplankton in P-limiting conditions. Such strategies are likely to be found globally. From a variety of possible phytoplankton strategies we propose here 4 of them (Fig. 8) that in the light of results presented herein, authors' experiences and earlier work (Gašparović et al., 2013; Godrijan et al., 2013; Ivančić et al., 2004; 2012; Viličić et al., 2013) appear to be the most plausible.

Fig. 8

Strategy 1: Efficient PO_4 uptake. K_m for PO_4 uptake should be lower than PO_4 concentration. Species dominate in oligotrophic conditions without APA when P is not limiting. Cell division time is not essential but rather endurance and uptake efficiency. Such species use APA for survival or bloom formation/prolongation after unbalanced nutrient input.

Strategy 2: Sustaining abundances and activity during P-limitation with APA. Fast bloom formation when nutrients become available due to sustained abundances and active state. The absence of APA in oligotrophic conditions might be due to limitation with other resources.

Strategy 3: Sustaining abundances or bloom formation on P-pools or/and production of non-phospholipids. Species co-dominating in P-limited conditions without APA.

Strategy 4: Symbiotic relationship with bacteria. Species that co-dominate in P-limited conditions without APA, but with attached bacteria that do express APA. When experiencing eutrophic riverine input those species grow fast and co-dominate the community. Under oligotrophic conditions those species are moderately abundant but attached bacteria are virtually absent or do not express APA.

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References

- Andersen T., Schartau, A.K.L., Paasche, E., 1991. Quantifying external and internal nitrogen and phosphorus pools, as well as nitrogen and phosphorus supplied through remineralization, in coastal marine plankton by means of a dilution technique. *Marine Ecology Progress Series* 69, 67-80.
- Artegiani, A., Bregant, D., Paschini, E., Pinardi, N., Raicich, F., Russo, A., 1997. The Adriatic sea general circulation. Part II: baroclinic circulation structure. *Journal of Physical Oceanography* 27 (8), 1515-1532.
- Casey, J.R., Lomas, M.W., Michelou, V.K., Dyhrman, S.T., Orchard, E.D., Ammerman, J.W., Sylvan, J.B., 2009. Phytoplankton taxon-specific orthophosphate (Pi) and ATP utilization in the western subtropical North Atlantic. *Aquatic Microbial Ecology* 58 (1), 31-44.
- Cotner, J.B., Ammerman, J.W., Peele, E.R., Bentzen, E., 1997. Phosphorus-limited

493 bacterioplankton growth in the Sargasso Sea. *Aquatic Microbial Ecology* 13 (2), 141-149.

494 Cozzi, S., Giani, M., 2011. River water and nutrient discharges in the Northern Adriatic Sea:
 495 Current importance and long term changes. *Continental Shelf Research* 31 (18), 1881-
 496 1893.

497 Dyhrman, S., Palenik, B., 1999. Phosphate stress in cultures and field populations of the
 498 dinoflagellate *Prorocentrum minimum* detected by a single-cell alkaline phosphatase
 499 assay. *Applied and Environmental Microbiology* 65 (7), 3205-3212.

500 Dyhrman, S.T., Webb, E.A., Anderson, D.M., Moffett, J.W., Waterbury, J.B., 2002. Cell-
 501 specific detection of phosphorus stress in *Trichodesmium* from the Western North
 502 Atlantic. *Limnology and Oceanography* 47 (6), 1832-1836.

503 Dyhrman, S., Ruttenberg, K., 2006. Presence and regulation of alkaline phosphatase activity
 504 in eukaryotic phytoplankton from the coastal ocean: implications for dissolved organic
 505 phosphorus remineralization. *Limnology and Oceanography* 51 (3), 1381–1390.

506 Egge, J.K., 1998. Are diatoms poor competitors at low phosphate concentrations? *Journal of*
 507 *Marine Systems* 16 (3-4), 191-198.

508 Eppley, R.W., Roger, J.N., McCarthy, J.J., 1969. Half-saturation constants for uptake of
 509 nitrate and ammonium by marine phytoplankton. *Limnology and Oceanography* 14 (6),
 510 912-920.

511 Gašparović, B., Godrijan, G., Frka, S., Tomažić, I., Penezić, A., Marić, D., Djakovac, T,
 512 Ivančić, I., Paliaga, P., Lyons, D., Precali, R., Tepić, N., 2013. Adaptation of marine
 513 plankton to environmental stress by glycolipid accumulation. *Marine Environmental*
 514 *Research* 92, 120-132.

515 Girault, M., Arakawa, H., Hashihama, F., 2013. Phosphorus stress of microphytoplankton
 516 community in the western subtropical North Pacific. *Journal of Plankton Research* 35
 517 (1), 146-157.

518 Godrijan, J., Marić, D., Tomažić, I., Precali, R., Pfannkuchen, M., 2013. Seasonal
 519 phytoplankton dynamics in the coastal waters of the north-eastern Adriatic Sea. *Journal of*
 520 *Sea Research* 77, 32-44.

521 González-Gill, S., Keafer, B.A., Jovine, R.V.M., Aguilera, A., Lu, S., Anderson, D.M., 1998.
 522 Detection and quantification of alkaline phosphatase in single cells of phosphorus-starved
 523 marine phytoplankton. *Marine Ecology Progress Series* 164, 21-35.

524 Hoppe, H.-G., 1983. Significance of exoenzymatic activities in the ecology of brackish water:
 525 Measurements by means of methylumbelliferyl-substrates. *Marine Ecology Progress*
 526 *Series* 11, 299-308.

527 Hoppe, H.-G., 2003. Phosphatase activity in the sea. *Hydrobiologia* 493 (1), 187-200.

528 Ivančić, I., Degobbi, D., 1984. An optimal manual procedure for ammonia analysis in natural
 529 waters by the indophenol blue method. *Water Research* 18, 1143-1147.

530 Ivančić, I., Degobbi, D., 1987. Mechanisms of production and fate of organic phosphorus in
 531 the northern Adriatic Sea. *Marine Biology* 94 (1), 117-125.

532 Ivančić, I., Degobbi, D., Pečar, O., Fuks, D., Manganelli, M., Kraus, R., Djakovac, T.,
 533 Precali, R., Scenati, R., 2004. Northern Adriatic mesocosm experiment Rovinj 2003:
 534 Nutrient dynamics. *Periodicum Biologorum* 106 (1), 17-22.

535 Ivančić, I., Radić, T., Lyons, D.M., Fuks, D., Precali, R., Kraus, R., 2009. Alkaline
 536 phosphatase activity in relation to nutrient status in the northern Adriatic Sea. *Marine*
 537 *Ecology Progress Series* 378, 27-35.

538 Ivančić, I., Fuks, D., Radić, T., Lyons, D. M., Šilović, T., Kraus, R., Precali, R., 2010.
 539 Phytoplankton and bacterial alkaline phosphatase activity in the northern Adriatic Sea.
 540 *Marine Environmental Reseserch* 69, 85-94.

541 Ivančić, I., Godrijan, J., Pfannkuchen, M., Marić, D., Gašparović, B., Đakovac, T., Najdek,
 542 M., 2012. Survival mechanisms of phytoplankton in conditions of stratification induced
 543 deprivation of orthophosphate: Northern Adriatic case study. *Limnology and*

544 Oceanography 57 (6), 1721-1731.

545 Jeong, H., Yoo, Y., Kim, J., Seong, K., Kang, N., Kim, T., 2010. Growth, feeding and
546 ecological roles of the mixotrophic and heterotrophic dinoflagellates in marine planktonic
547 food webs. *Ocean Science Journal* 45 (2), 65-91.

548 Karl, D.M., Letelier, R., Tupas, L., Dore, J., Christian, J., Hebel, D., 1997. The role of
549 nitrogen fixation in biogeochemical cycling in the subtropical North Pacific Ocean.
550 *Nature* 388, 533-538.

551 Karl, D.M., Bidigare, R.R., Letelier, R.M., 2001. Long-term changes in phytoplankton
552 community structure and productivity in the North Subtropical Gyre: the domain shift
553 hypothesis. *Deep-Sea Res II* 48, 1449-1470.

554 Karl, D.M., 2007. The marine phosphorus cycle. In: Hurst, C.J. (ed.) *Manual of*
555 *environmental microbiology*. ASM Press, Washington, DC, pp 523-539.

556 Labry, C., Delmas, D., Herbland, A., 2005. Phytoplankton and bacterial alkaline phosphatase
557 activities in relation to phosphate and DOP availability within the Gironde plume
558 waters (Bay of Biscay). *Journal of Experimental Marine Biology and Ecology* 318 (2),
559 213-225.

560 Larato, C., Celussi, M., Virgilio, D., Karuza, A., Falconi, C., De Vittor, C., Del Negro, P.,
561 Fonda Umani, S., 2010. Production and utilization of organic matter in different P-
562 availability conditions: A mesocosm experiment in the Northern Adriatic Sea. *Journal*
563 *of Experimental Marine Biology and Ecology* 391 (1), 131-142.

564 Lomas, M.W., Swain, A., Shelton, R., Ammerman, J.W., 2004. Taxonomic variability of
565 phosphorus stress in Sargasso Sea phytoplankton. *Limnology and Oceanography* 49 (6),
566 2303-2310.

567 Lyons, D.M., Supić, N., Smolaka, N., 2007. Geostrophic Circulation Patterns in the
568 Northeastern Adriatic Sea and the Effects of Air-Sea Coupling: May-September 2003.

569 Journal of Geophysical Research 112 (C3), 1-17.

570 Mackey, K.R.M., Labiosa, R.G., Calhoun, M., Street, J.H., Post, A.F., Paytan, A., 2007.

571 Phosphorus availability, phytoplankton community dynamics, and taxon-specific

572 phosphorus status in the Gulf of Aqaba, Red Sea. Limnology and Oceanography 52

573 (2), 873-885.

574 Malfatti, F., Turk, V., Tinta, T., Mozetič, P., Manganelli, M., Samo, T.J., Ugalde, J.A.,

575 Kovač, N., Stefanelli, M., Antonioli, M., Fonda-Umani, S., Del Negro, P., Cataletto,

576 B., Hozic, A., Ivošević DeNardis, N., Žutić, V., Svetličić, V., Mišić Radić, T., Radić,

577 T., Fuks, D., Azam F., 2014. Microbial mechanisms coupling carbon and phosphorus

578 cycles in phosphorus-limited northern Adriatic Sea. Science of the Total Environment

579 470-471, 1173-1183.

580 Maestrini, S.Y., Berland, B.R., Bréret, M., Béchemin, C., Poletti R., Rinaldi, A., 1997.

581 Nutrients Limiting the Algal Growth Potential (AGP) in the Po River Plume and an

582 Adjacent Area, Northwest Adriatic Sea: Enrichment Bioassays with the Test Algae

583 *Nitzschia closterium* and *Thalassiosira pseudonana*. Estuaries 20 (2), 416-429.

584 Menzel, D.W., Corwin, N., 1965. The measurement of total phosphorus in seawater based on

585 the liberation of organically bound fractions by persulfate oxidation. Limnology and

586 Oceanography 10 (2), 280-282

587 Moutin, T., Van Den Broeck, N., Beker, B., Dupouy, C., Rimmelin, P., Le Bouteiller, A.,

588 2005. Phosphate availability controls *Trichodesmium* spp. biomass in the SW Pacific

589 Ocean. Marine Ecology Progress Series 297, 15-21.

590 Nausch, M., Nausch, G., Wasmund, N., 2004. Phosphorus dynamics during the transition from

591 nitrogen to phosphate limitation in the central Baltic Sea. Marine Ecology Progress

592 Series 266, 15-25.

593 Oh, S.J., Kwon, H.K., Noh, I.H., Yang, H-S., 2010. Dissolved Organic Phosphorus Utilization

594 and Alkaline Phosphatase Activity of the Dinoflagellate *Gymnodinium impudicum*
595 Isolated from the South Sea of Korea. Ocean Science Journal 45 (3), 171-178.

596 Ou, L., Huang, B., Lin, L., Hong, H., Zhang, F., Chen, Z., 2006. Phosphorus stress of
597 phytoplankton in the Taiwan Strait determined by bulk and single-cell alkaline
598 phosphatase activity assays. Marine Ecology Progress Series 327, 95-106.

599 Pfannkuchen, M., Marić, D., Godrijan, J., Fritz, G., Brümmer, F., Jaklin, A., Hamer, B., Batel,
600 R., 2009. Sponges (Porifera) and eukaryotic, unicellular plankton. A case study on
601 *Aplysina aerophoba*, Nardo 1886 in the Northern Adriatic. Journal of Experimental
602 Marine Biology and Ecology 382 (1), 40-46.

603 Pfannkuchen, M., Schlesinger, S., Fels, A., Brümmer, F., 2010. Microscopical techniques
604 reveal the in situ microbial association inside *Aplysina aerophoba*, Nardo 1886
605 (Porifera, Demospongiae, Verongida) almost exclusively consists of cyanobacteria.
606 Journal of Experimental Marine Biology and Ecology 390 (2), 169-178.

607 Redfield, A.C., Ketchum, B.H., Richards, F.A., 1963. The influence of organisms on the
608 composition of seawater. In Hill, M.N. (ed.), The sea. Interscience, pp. 27-77.

609 Russo, A., Maccaferri, S., Djakovac, T., Precali, R., Degobbis, D., Deserti, M., Paschini, E.,
610 Lyons, D.M., 2005. Meteorological and oceanographic conditions in the northern
611 Adriatic Sea during the period June 1999-July 2002: Influence on the mucilage
612 phenomenon. Science of Total Environment 353 (1-3), 24-38.

613 Sieburth, J.M., Smetacek, V., Lenz, J., 1978. Pelagic ecosystem structure: Heterotrophic
614 compartments of the plankton and their relationship to plankton size fractions.
615 Limnology and Oceanography 23 (6), 1256-1263.

616 Silikin, V.A., Pautova, L.A., Lifanchuk, A.V., 2013. Physiological regulatory mechanisms of
617 the marine phytoplankton community structure. Russian Journal of Plant Physiology
618 60 (4), 541-548.

619 Strickland, J.D.H., Parsons, T.R., 1972. A practical handbook of seawater analysis. Bulletin of
620 the Fisheries Research Board of Canada 167, 310.

621 Tantanasarit, C., Englande, A.J., Babel, S., 2013. Nitrogen, phosphorus and silicon uptake
622 kinetics by marine diatom *Chaetoceros calcitrans* under high nutrient concentrations.
623 Journal of Experimental Marine Biology and Ecology 446, 67-75.

624 Thingstad, T.F., Skjoldal, E.F., Böhne, R.A., 1993. Phosphorus cycling and algal-bacterial
625 competition in Sandsfjord, western Norway. Marine Ecology Progress Series 99, 239-
626 259.

627 Thingstad, T.F., Zweifel, U.L., Rassoulzadegan, F., 1998. P limitation of heterotrophic
628 bacteria and phytoplankton in the northwest Mediterranean. Limnology and
629 Oceanography 43 (1), 88-94.

630 Turk, V., Rehnstam, A.-S., Lundberg, E., Hagstrom, A., 1992. Release of Bacterial DNA by
631 Marine Nanoflagellates, an Intermediate Step in Phosphorus Regeneration. Applied
632 and Environmental Microbiology 58 (11), 3744-3750.

633 Tyrrell, T., 1999. The relative influences of nitrogen and phosphorus on oceanic primary
634 production. Nature 400, 525-531.

635 Utermöhl, H., 1958. Zur Vervollkommnung der quantitativen Phytoplankton-Methodik.
636 Mitteilungen des Internationale Vereinigung für theoretische und angewandte
637 Limnologie 9, 1-38.

638 Van Mooy, B.A.S., Fredricks, H.F., Pedler, B.E., Dyhrman, S.T., Karl, M., Koblížek, M.,
639 Lomas, M.W., Mincer, T.J., Moore, L.R., Moutin, T., Rappe, M.S., Webb, E.A., 2009.
640 Phytoplankton in the ocean use non-phosphorus lipids in response to phosphorus
641 scarcity. Nature 458, 69-72.

642 Viličić, D., Kuzmić, M., Tomažić, I., Ljubešić, Z., Bosak, S., Precali, R., Djakovac, T., Marić,
643 D., Godrijan, J., 2013. Northern Adriatic phytoplankton response to short Po River

644 discharge pulses during summer stratified conditions. *Marine Ecology-an Evolutionary*
645 *Perspective* 34 (4), 451-466.

646 Viličić D, Kuzmić M, Bosak S, Šilović T, Hrustić, E., Burić, Z., 2009. Distribution of
647 phytoplankton along the thermohaline gradient in the northeastern Adriatic channel;
648 winter aspect. *Oceanologia* 51 (4), 495-513.

649 Wasmund, N., Nausch, G., Hansen, A., 2014. Phytoplankton succession in an isolated
650 upwelled Benguela water body in relation to different initial nutrient conditions.
651 *Journal of Marine Systems* 140, 163-174.

652 Xu, J., Yin, K., He, L., Yuan, X., Ho, A.Y.T., Harrison, P.J., 2008. Phosphorus limitation in the
653 northern South China Sea during late summer: Influence of the Pearl River. *Deep-Sea*
654 *Research I* 55 (10), 1330-1342.

655 Yamaguchi, H., Adachi, M., 2010. The utilization of organic phosphorus by eukaryotic
656 phytoplankton in marine environments (review). *Bulletin of the Plankton Society of*
657 *Japan* 57 (1), 1-12.

658 Zavatarelli, M., Raicich, F., Bregant, D., Russo, A., Artegiani, A., 1998. Climatological
659 biogeochemical characteristics of the Adriatic Sea. *Journal of Marine Systems* 18, 227-
660 263.

661

662 Table 1. ELF detection of extracellular alkaline phosphatase activity in
663 microphytoplankton species at SJ101 and RV001 during 2013. ELF positive species (+),
664 ELF negative species (-) and ELF positive species where the signal is located at
665 attached bacteria (a).

Species				
DIATOM	ELF	DINOFLAGGELATE	ELF	
<i>Asterionella formosa</i>	-	<i>Alexandrium minutum</i>	-	
<i>Asteromphalus heptactis</i>	-	<i>Ceratium candelabrum</i>	-	
<i>Bacteriastrium furcatum</i>	+	<i>Ceratium extensum</i>	-	
<i>Bacteriastrium hyalinum</i>	+	<i>Ceratium falcata</i>	-	
<i>Bacteriastrium jadranum</i>	-	<i>Ceratium furca</i>	+	
<i>Bacteriastrium mediterraneum</i>	-	<i>Ceratium furca</i> var. <i>eugrammum</i>	-	
<i>Cerataulina pelagica</i>	+	<i>Ceratium fusus</i>	-	
<i>Chaetoceros affinis</i>	-	<i>Ceratium hexacanthum</i>	-	
<i>Chaetoceros brevis</i>	a	<i>Ceratium pavillardii</i>	-	
<i>Chaetoceros circinalis</i>	+	<i>Ceratium symmetricum</i>	-	
<i>Chaetoceros concavicornis</i>	-	<i>Ceratium trichoceros</i>	+	
<i>Chaetoceros constrictus</i>	+	<i>Ceratium tripos</i>	-	
<i>Chaetoceros curvisetus</i>	-	<i>Dinophysis caudata</i>	+	
<i>Chaetoceros dadayi</i>	-	<i>Diplopsalis complex</i>	-	
<i>Chaetoceros danicus</i>	+	<i>Goniaulax polygramma</i>	-	
<i>Chaetoceros diversus</i>	+	<i>Goniaulax</i> sp.	-	
<i>Chaetoceros lorenzianus</i>	-	<i>Goniodoma acuminatum</i>	+	
<i>Chaetoceros rostratus</i>	-	<i>Gymnodinium</i> sp.	-	
<i>Chaetoceros socialis</i>	+/-a	<i>Gyrodinium fusiforme</i>	-	
<i>Chaetoceros</i> sp.	-	<i>Gyrodinium</i> sp.	-	
<i>Chaetoceros tetrastichon</i>	-	<i>Hermesinium adriaticum</i>	+	
<i>Chaetoceros tortissimus</i>	-	<i>Heterocapsa</i> sp.	-	
<i>Chaetoceros vixvisibilis</i>	+	<i>Kofoedinium velelloides</i>	-	
<i>Chaetoceros wighami</i>	-	<i>Minuscula bipes</i>	-	
<i>Chateoceros contortus</i>	+	<i>Oxyphysis oxytoxoides</i>	+	
<i>Coscinodiscus oculus iridis</i>	-	<i>Oxytoxum caudatum</i>	-	
<i>Coscinodiscus</i> sp.	-	<i>Oxytoxum sceptrum</i>	-	
<i>Cyclotella</i> sp.	+	<i>Oxytoxum variabile</i>	-	
<i>Cylindrotheca closterium</i>	-	<i>Phalacroma biceps</i>	-	
<i>Dactyliosolen mediterraneus</i>	-	<i>Phalacroma rotundatum</i>	-	
<i>Diploneis bombus</i>	-	<i>Podolampas elegans</i>	-	
<i>Eucampia cornuta</i>	+	<i>Prorocentrum compressum</i>	-	
<i>Guinardia flaccida</i>	+	<i>Prorocentrum micans</i>	-	
<i>Guinardia striata</i>	-	<i>Prorocentrum minimum</i>	-	
<i>Hemiaulus hauckii</i>	+	<i>Prorocentrum triestinum</i>	-	
<i>Leptocylindrus danicus</i>	+	<i>Protoperidinium brochii</i>	-	
<i>Leptocylindrus minimus</i>	-	<i>Protoperidinium conicum</i>	-	
<i>Licmophora</i> sp.	-	<i>Protoperidinium diabolus</i>	+	

708	<i>Lioloma pacificum</i>	+	<i>Protoperidinium divergens</i>	-
709	<i>Navicula</i> sp.	-	<i>Protoperidinium leonis</i>	-
710	<i>Nitzschia incerta</i>	-	<i>Protoperidinium ovum</i>	-
711	<i>Paralia sulcata</i>	-	<i>Protoperidinium pallidum</i>	+
712	<i>Pleurosigma</i> sp.	+	<i>Protoperidinium paulseni</i>	-
713	<i>Pseudonitzschia</i> spp.	+	<i>Protoperidinium pyriforme</i>	-
714	<i>Rhizosolenia alata</i> f. <i>gracillima</i>	+	<i>Protoperidinium solidicorne</i>	-
715	<i>Rhizosolenia alata</i> var. <i>indica</i>	-	<i>Protoperidinium steinii</i>	-
716	<i>Rhizosolenia calcar-avis</i>	+	<i>Protoperidinium tuba</i>	-
717	<i>Rhizosolenia fragilissima</i>	+	<i>Pseliodinium vaubanii</i>	-
718	<i>Rhizosolenia imbricata</i>	+	<i>Pyrocystis lunula</i>	-
719	<i>Rhizosolenia robusta</i>	-	<i>Scripsiella</i> sp.	-
720	<i>Rhizosolenia</i> sp.	-	<i>Torodinium</i> sp.	-
721	<i>Rhizosolenia stolterfothii</i>	-		
722	<i>Skeletonema marinoi</i>	+		
723	<i>Striatella unipunctata</i>	-		
724	<i>Thalassionema nitzschioides</i>	+		
725	<i>Thalassiosira angulata</i>	-		
726	<i>Thalassiosira</i> sp.	+		
727	<i>Thalassiothrix frauenfeldii</i>	-		

728

729

730 Table 2. ELF positive dinoflagellates and
731 their contribution to the dinoflagellate
732 abundance at SJ101 and RV001 during
733 2013.

734

735

736 Species Contribution (%)

737

738

739 *Ceratium furca* 0.54-6.99

740 *Ceratium trichoceros* 0.33-1.72

741 *Dinophysis caudata* 0.76

742 *Goniodoma acuminatum* 0.64

743 *Hermesinium adriaticum* 0.76-8.17

744 *Oxyphysis oxytoxoides* 0.64

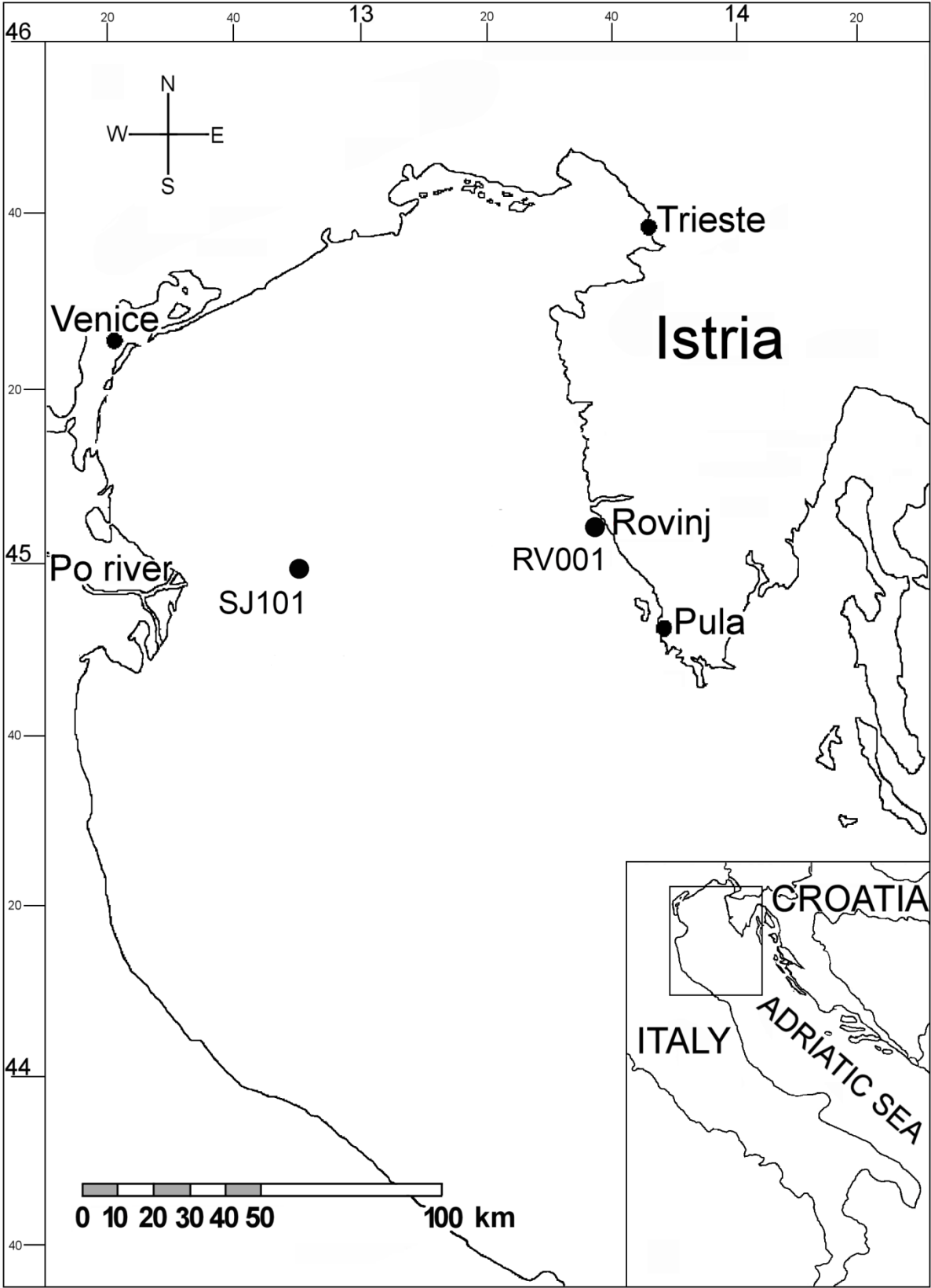
745 *Protoperidinium diabolus* 0.33-0.64

746 *Protoperidinium pallidum* 0.81

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751

752 Fig. 1. Research area and sampling stations in the northern Adriatic Sea.

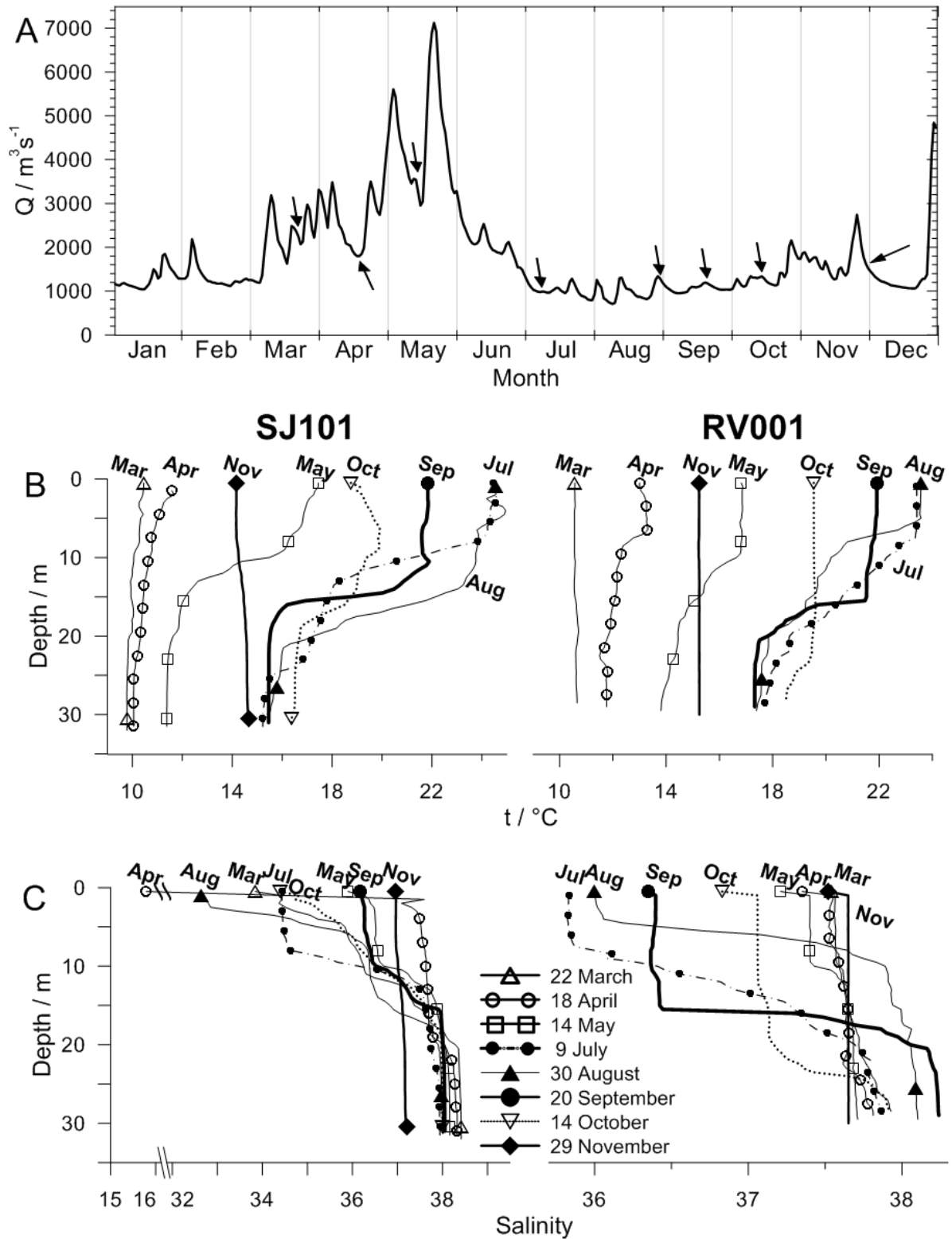


Fig. 2. (A) Daily mean of the Po River discharge rate (Q) with cruise dates denoted by arrows.

(B) Temperature (t) and (C) salinity profiles at SJ101 and RV001 during 2013.

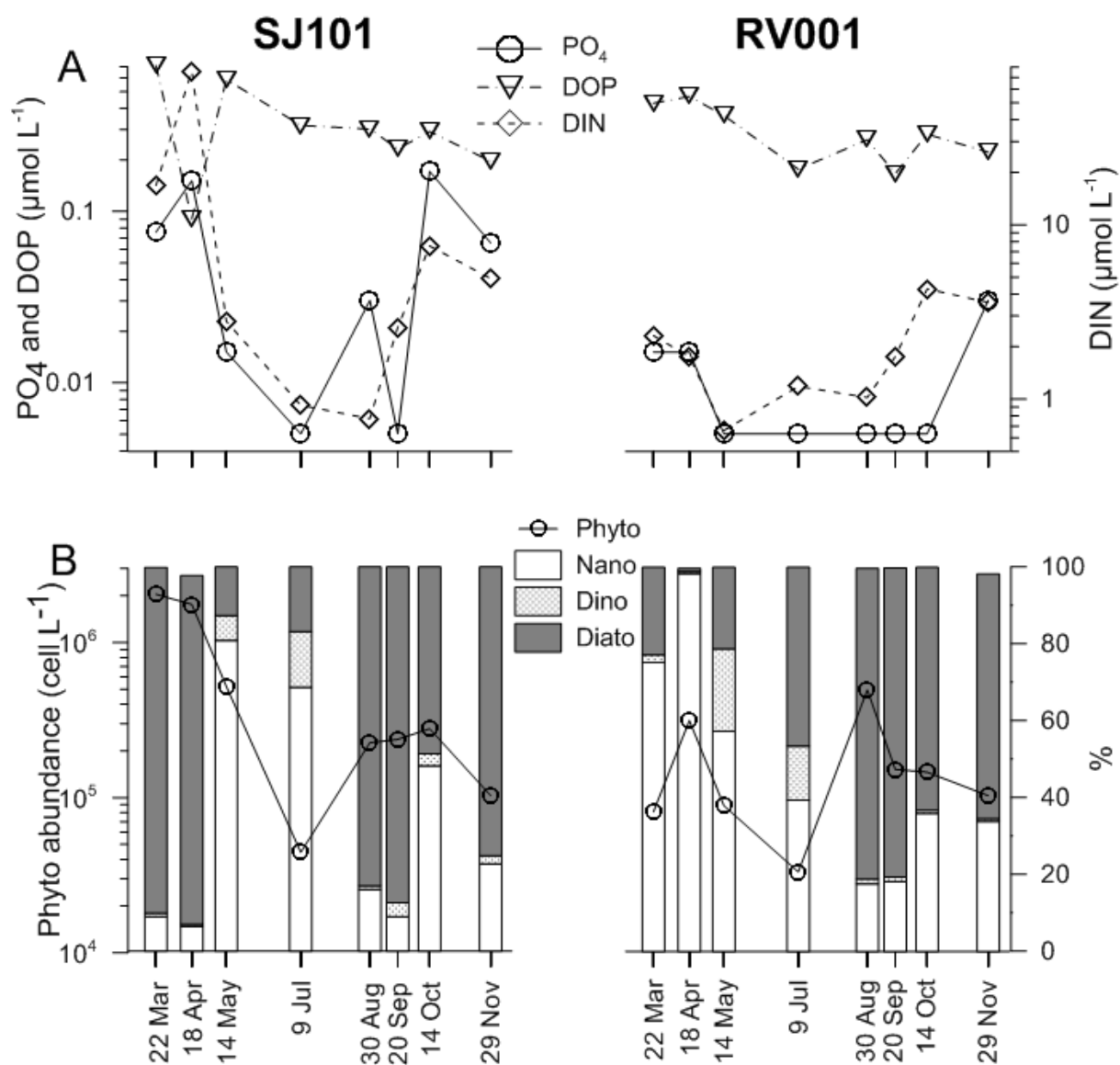
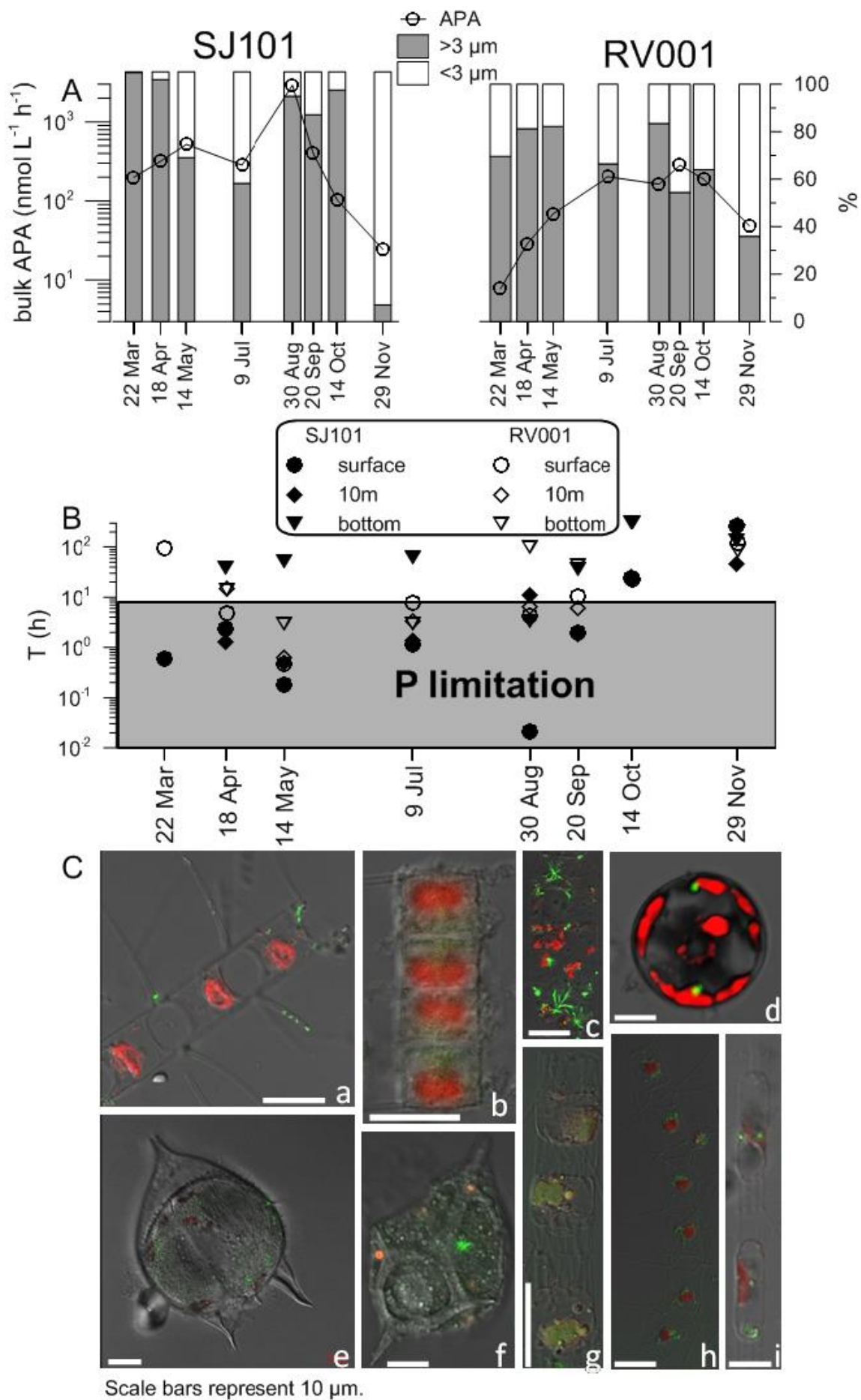


Fig. 3. (A) Concentrations for orthophosphate (PO₄; circle), dissolved organic phosphorus (DOP; triangle), dissolved inorganic nitrogen (DIN; diamond), and (B) phytoplankton abundance (phyto) and community composition at surface of SJ101 and RV001 during 2013.



761 Fig. 4. (A) Bulk APA, contribution of $> 3 \mu\text{m}$ and $< 3 \mu\text{m}$ fractions in the total activity at the
762 surface, and (B) APA-mediated turnover time at SJ101 and RV001 during 2013. P-limitation
763 is denoted by grey area. (C) Chloroplast fluorescence in red, ELF-mediated signal for APA in
764 green. *Chaetoceros brevis* APA with bacteria attached to setae (a). *Chaetoceros* sp. APA
765 diffused on the cell surface, dislocated from the chloroplasts (b). *Cerataulina pelagica* APA
766 on elongated, branching aggregations on the cell surface (c). *Thalassiosira* sp. APA in 2 small
767 aggregations per valve face, dislocated from the chloroplasts (d). *Protoperidinium pyriforme*
768 APA in microaggregations across the cell surface (e). *Hermesinum adriaticum* APA diffused
769 across the cell surface (f). *Skeletonema marinoi* APA on the cell surface where chloroplasts
770 are located (g). *Skeletonema marinoi* APA in few islands on the cell surface and dislocated
771 from chloroplasts (i). *Chaetoceros socialis* APA in concentrated islands on the cell surface
772 adjacent to chloroplasts (h). Scale bars represent $10 \mu\text{m}$.

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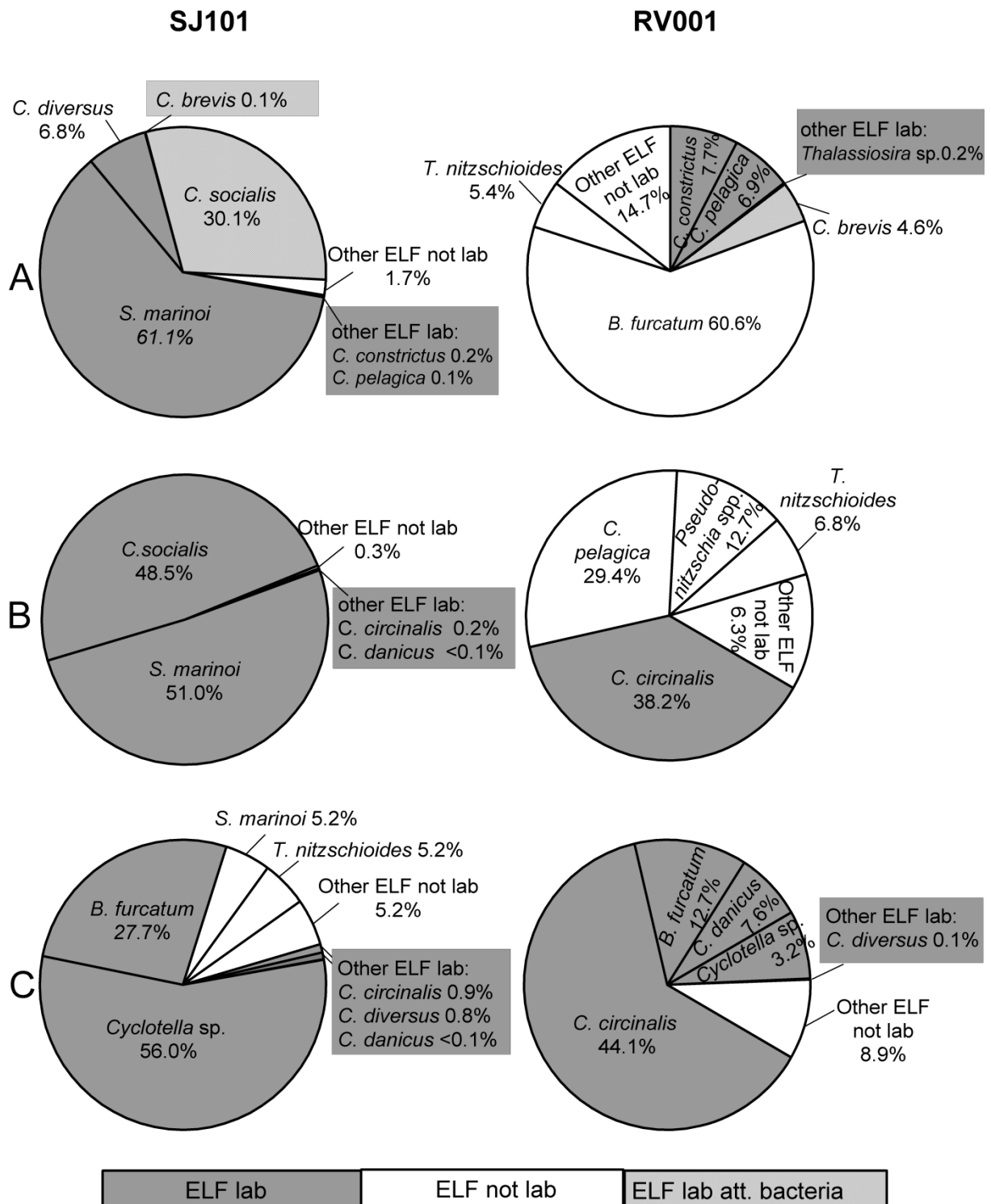
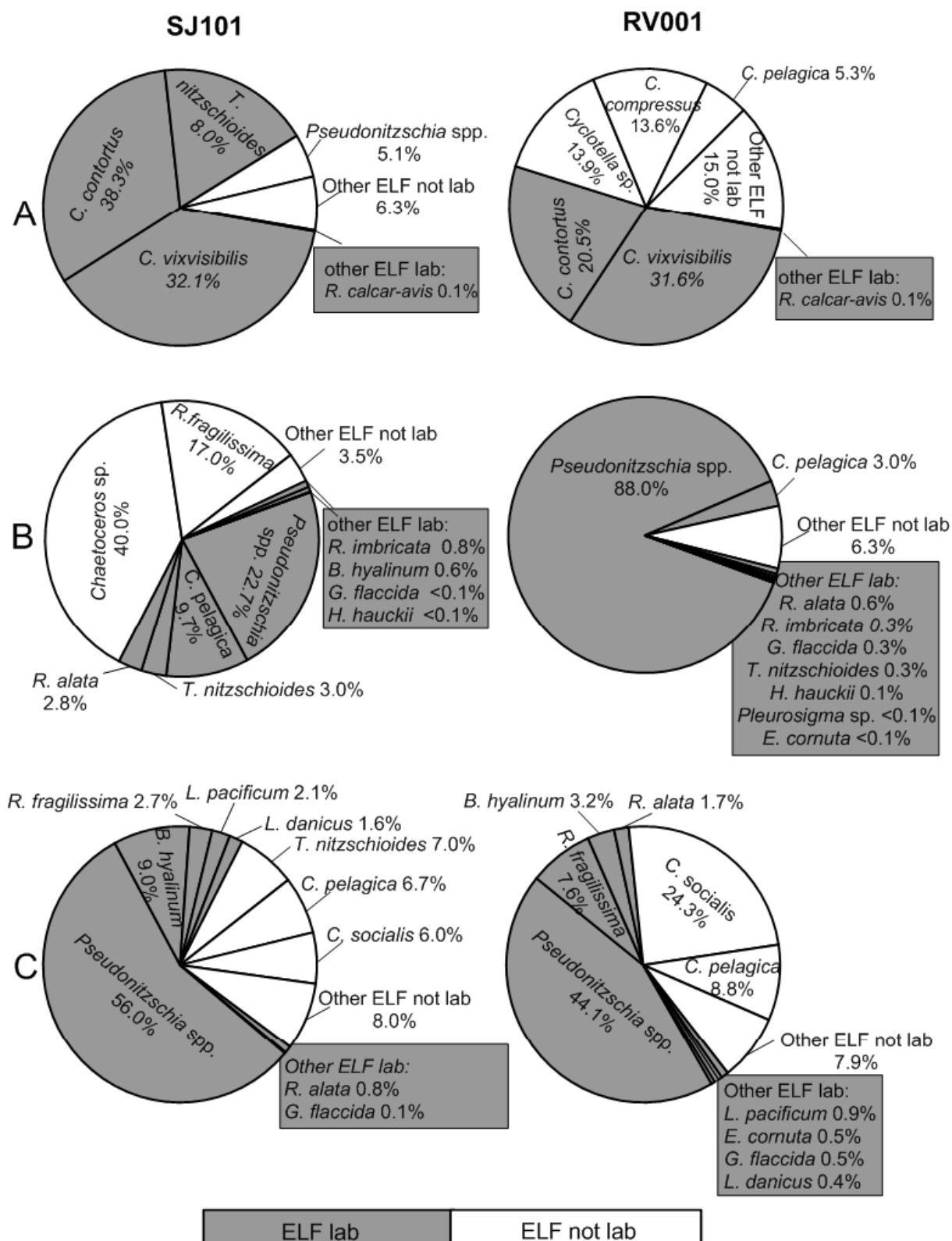


Fig. 5. Contribution (%) of ELF labelled (ELF lab) and ELF not labelled (ELF not lab) species in diatom community, as well as contribution of species with labelled attached bacteria (att. bacteria ELF lab) in the upper waters (0-10 m) at SJ101 and RV001 in (A) March (B) April and (C) May 2013.

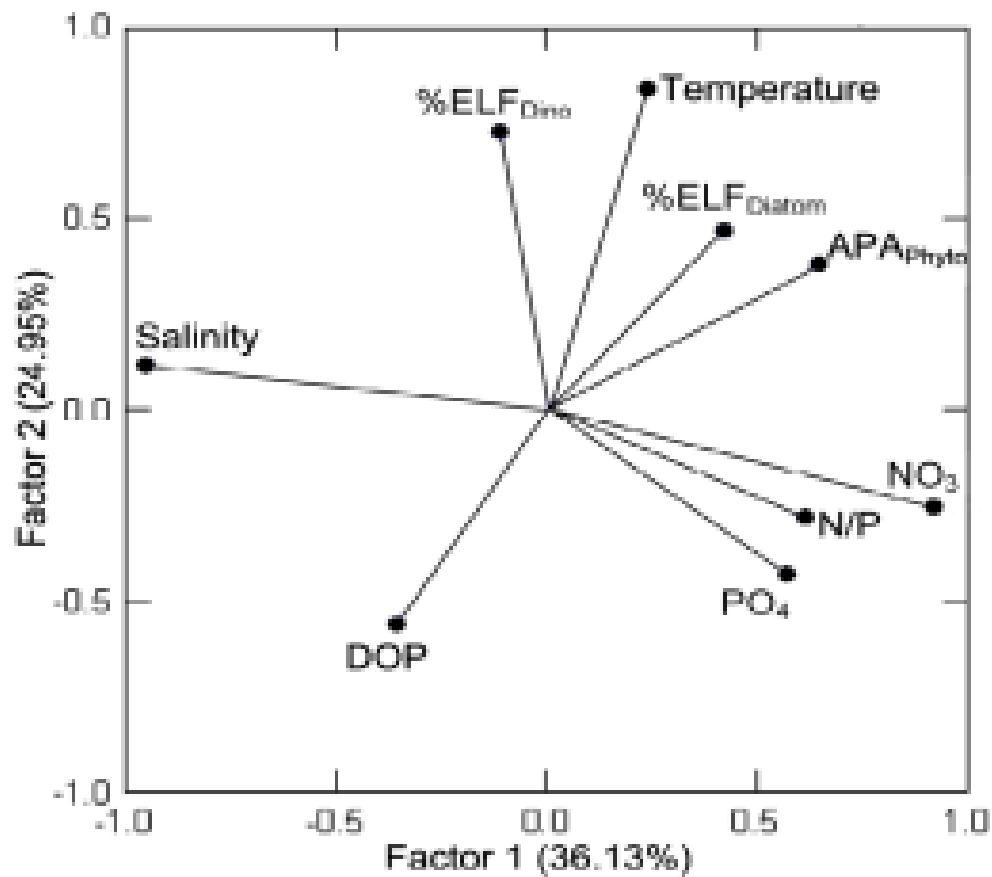


779

780 Fig. 6. Contribution (%) of ELF labelled (ELF lab) and ELF not labelled (ELF not lab)

781 species in diatom community in the upper waters (0-10 m) at SJ101 and RV001 in (A) July

782 (B) August and (C) September 2013.



783

784 Fig. 7. Principal Component Analysis of measured variables: temperature, salinity,
 785 orthophosphate (PO₄), nitrate (NO₃), dissolved organic phosphorus (DOP), inorganic N/P
 786 ratio, contribution of ELF positive species in diatom community (%ELF_{Diatom}), contribution of
 787 ELF positive species in dinoflagellate community (%ELF_{Dino}) and APA in fraction > 3μm
 788 (APA_{Phyto}).

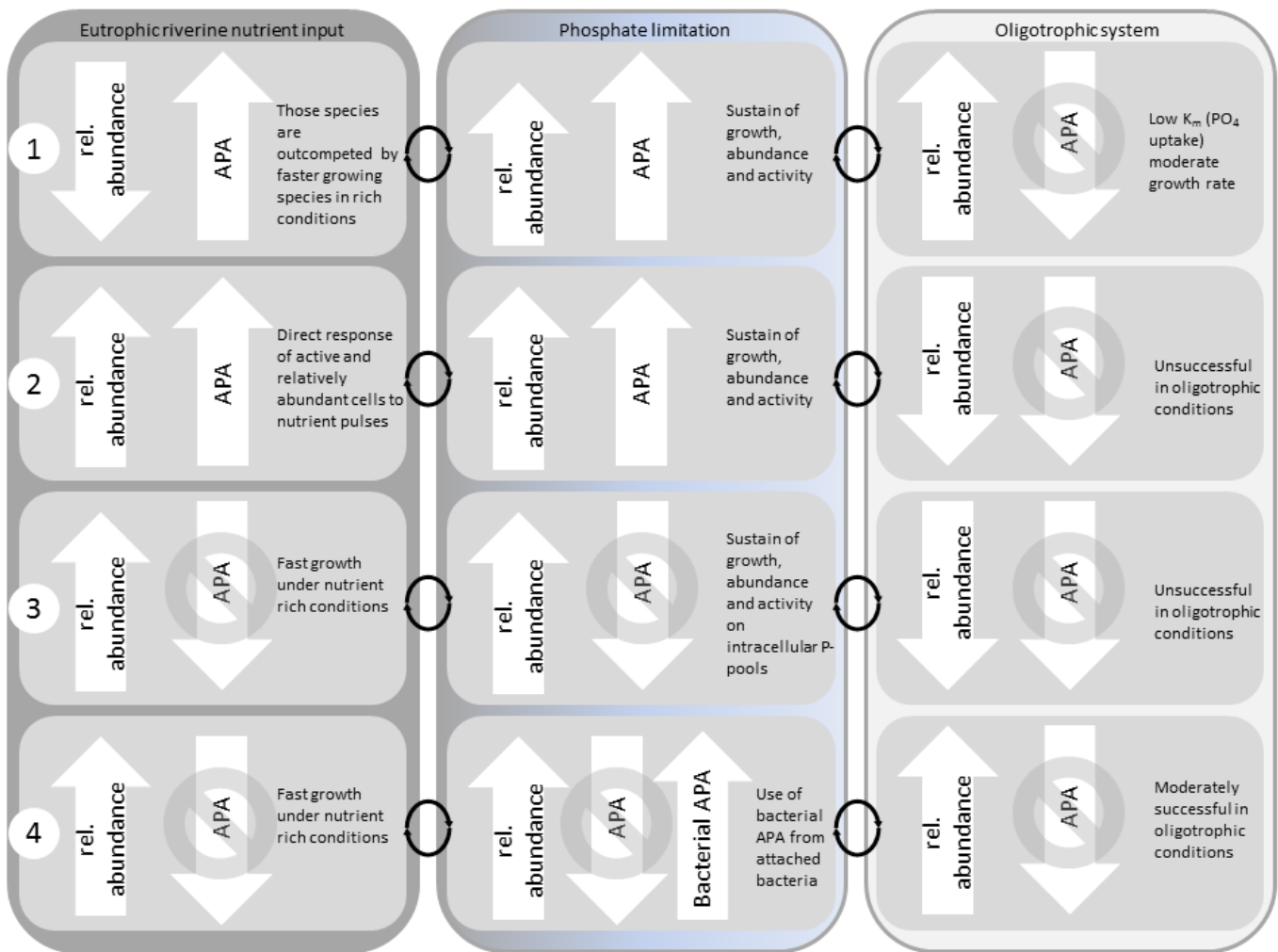


Fig. 8. Four identified strategies for successful competition in the NA. Within one row the figure shows how one strategy addresses the different trophic conditions that plankton species face when following the current systems in the NA (as shown by circular arrows). Although both areas (SJ101 and RV001) can be P-limited, they completely differ in trophic conditions: prevalently eutrophic at SJ101 and more oligotrophic on RV001.

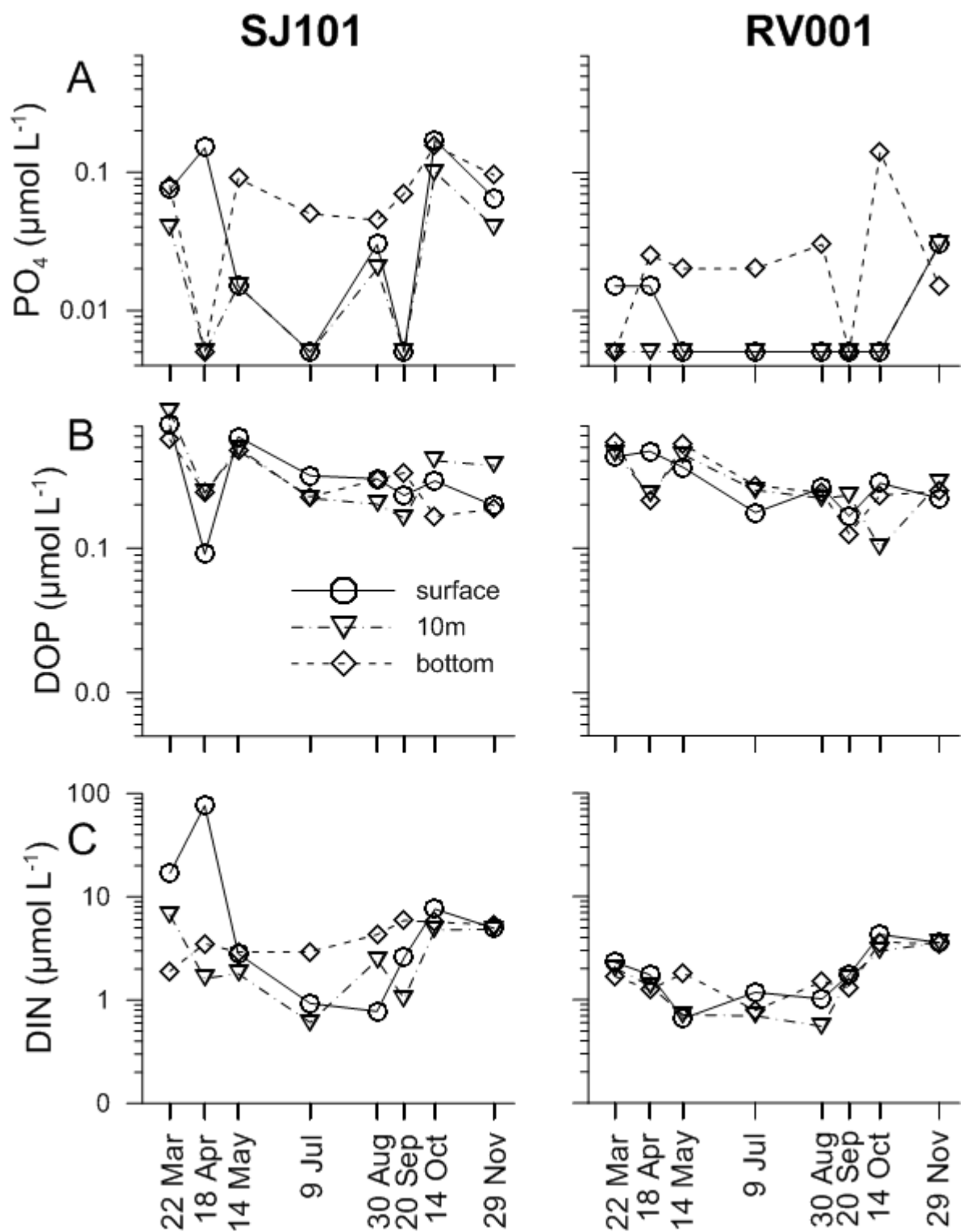


Fig. S1. Concentrations for (A) orthophosphate (PO_4), (B) dissolved organic phosphorus (DOP) and (C) dissolved inorganic nitrogen (DIN) at SJ101 and RV001 during 2013.

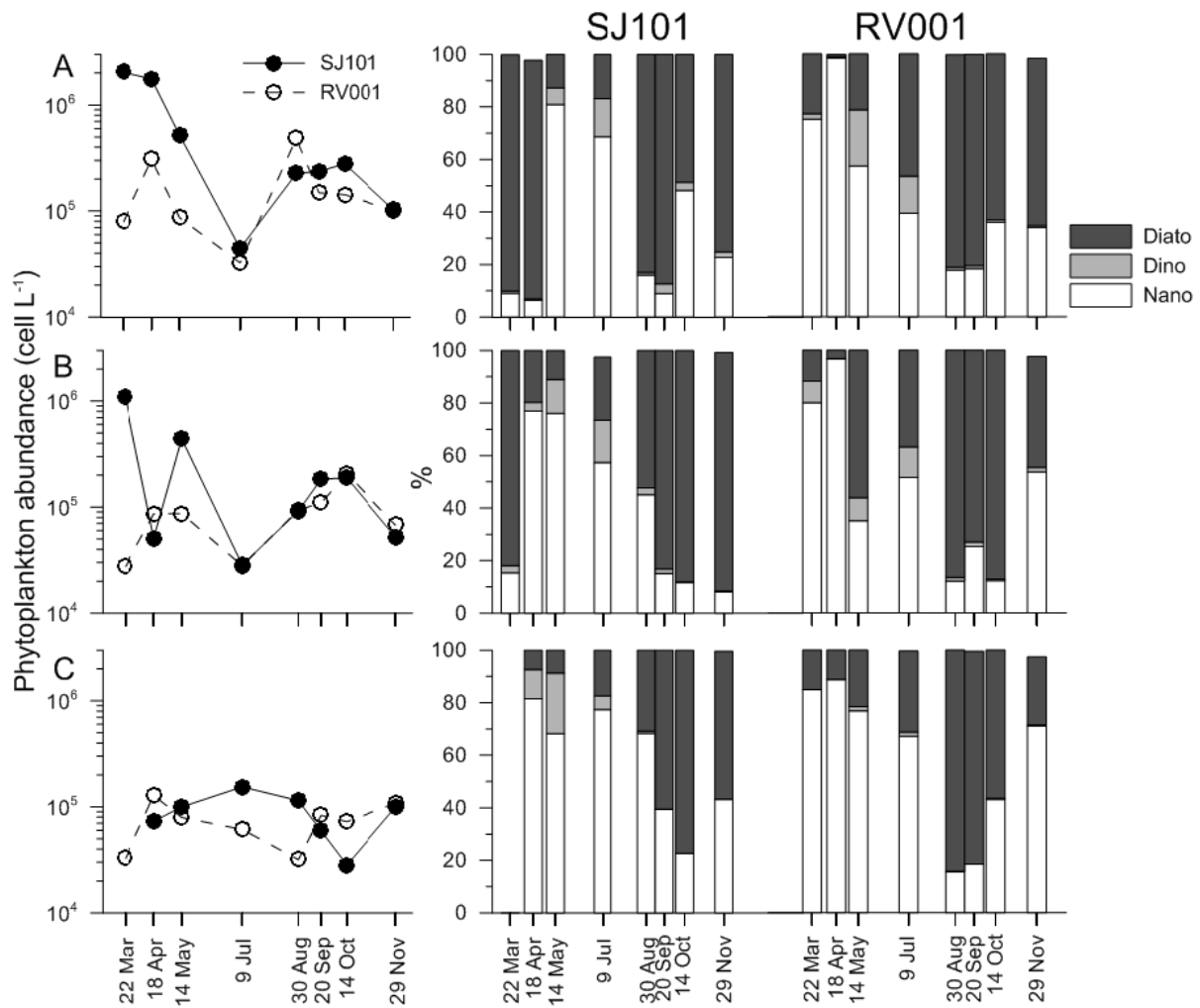


Fig. S2. Phytoplankton abundance (Phyto abundance) and contribution of diatoms, dinoflagellates and nanophytoplankton in phytoplankton community at (A) surface, (B) 10 m depth and (C) bottom at SJ101 and RV001 during 2013. In March it was not possible to count nanophytoplankton at the bottom of SJ101 (due to the high quantity of detritus), and the relative contributions were not calculated.

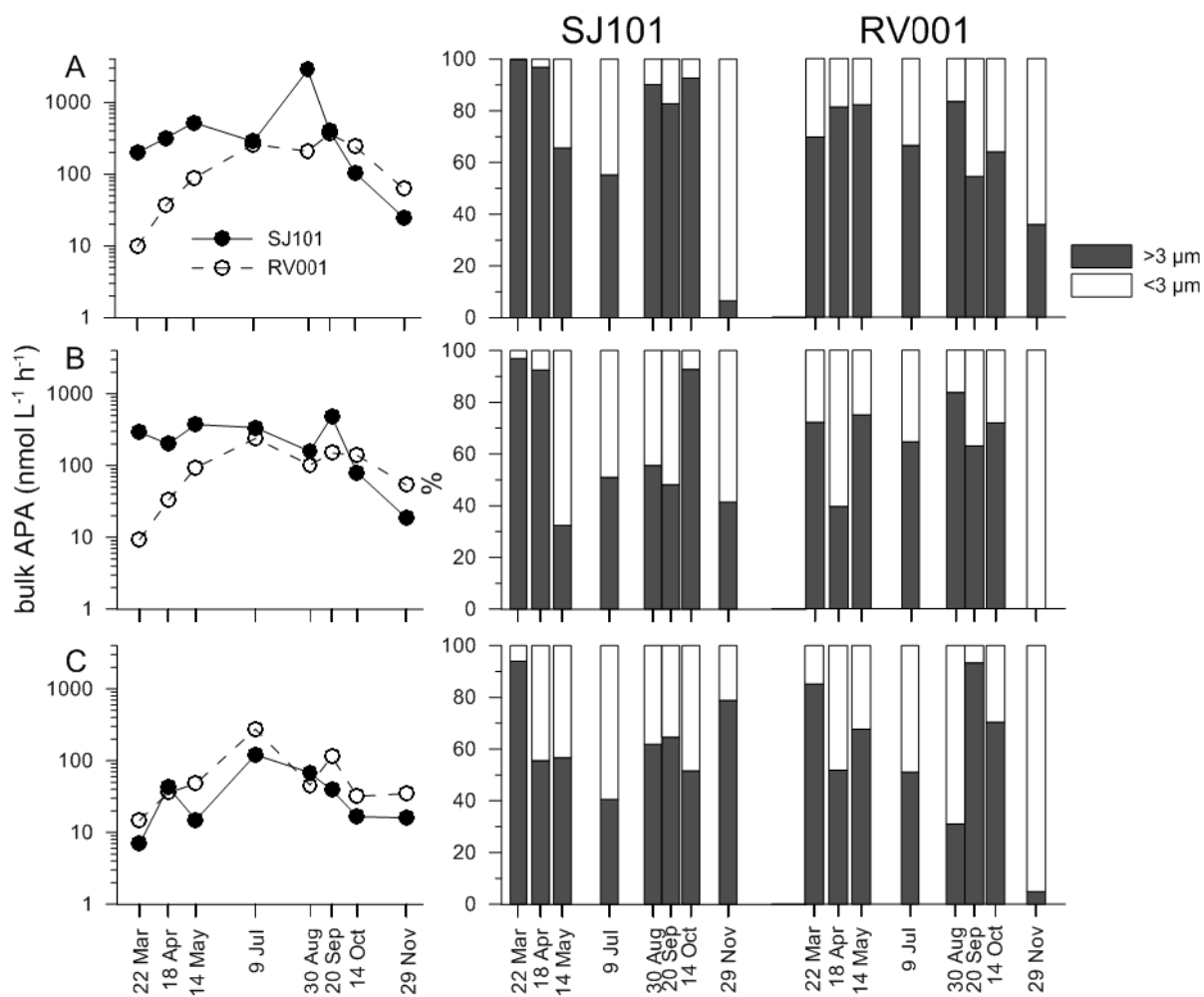


Fig. S3. Bulk APA and contribution of > 3 μm and < 3 μm fractions in the total activity at (A) surface, (B) 10 m depth and (C) bottom at SJ101 and RV001 during 2013.