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Title: OCCURRENCE AND FATE OF DISSOLVED AND PARTICULATE ANTIMICROBIALS IN MUNICIPAL WASTEWATER TREATMENT

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Corresponding Author: Dr. Marijan Ahel, PhD

Corresponding Author's Institution: Rudjer Boskovic Institute

First Author: Ivan Senta, PhD

Order of Authors: Ivan Senta, PhD; Senka Terzic, PhD; Marijan Ahel, PhD

Abstract: Comprehensive study on the occurrence and fate of several classes of antimicrobials, including sulfonamides, trimethoprim, fluoroquinolones and macrolides, in Croatian municipal wastewaters was performed using an integrated approach, which comprised analysis of both dissolved and particulate fractions. A nation-wide screening showed ubiquitous occurrence of human-use antimicrobials in raw wastewater samples with the total concentrations ranging from 2 to 20 µg/L, while veterinary antimicrobials were typically present in much lower concentrations (<100 ng/L). The percentage of the particulate fraction in raw wastewater varied significantly depending on the type of the antimicrobial and the load of suspended solids. A detailed study of the mass flows of dissolved and particulate antimicrobials, performed in the wastewater treatment plant of the city of Zagreb, allowed an improved assessment of the biological and physico-chemical removal mechanisms of investigated compounds during the conventional activated sludge treatment. The overall removal efficiencies of antimicrobials from the water phase were rather variable, ranging from 0% for trimethoprim to 85% for norfloxacin. A significant percentage of fluoroquinolones (norfloxacin and ciprofloxacin) and macrolides (azithromycin and clarithromycin) was associated with the primary and excess secondary sludge, explaining 14 to 77% of the total removal. The removal, which could be attributed to biological transformation, was relatively poor for all antimicrobials, exceeding 30% only for SMX (32%) and clarithromycin (55%).

Suggested Reviewers: Sara Castiglioni Ph.D.

Res. associate, Dept Environm Hlth Sci, Mario Negri Inst Pharmacol Res, Milano, Italy

Sara.Castiglioni@marionegri.it

Specialist in the field of pharmaceuticals in the environment

Mira Petrovic Ph.D.

Research professor, Catalan Institute for Water Research ICRA

mpetrovic@icra.cat

Highly productive researcher in the field of emerging contaminants

Barbara Kasprzyk-Hordern Ph.D.

Dept Chem, Bath, Univ Bath

b.kasprzyk-hordern@bath.ac.uk

Prominent author in the field of emerging contaminants

Tuhkanen Tuula Ph.D.

Prof., Inst Environm Engn & Biotechnol, Tampere Univ Technol

tuula.tuhkanen@tut.fi

Expert in the field of pharmaceuticals in wastewater treatment

Ruder Bošković Institute



DIVISION FOR MARINE AND ENVIRONMENTAL RESEARCH

Tel: ++ 385 1 45 60 940 Fax: ++385 1 468 02 42

Bijenička cesta 54, P.O. BOX 180, 10002 Zagreb, Croatia

Zagreb, 12 June 2012

Dear dr. Loosdrecht,

please find enclosed our manuscript on Occurrence and fate of dissolved and particulate antimicrobials in municipal wastewater treatment by Ivan Senta, Senka Terzic and Marijan Ahel, which we would like to be considered for publication in *Water Research* as a full length research article. The paper represents a comprehensive study on the occurrence and fate of several classes of antimicrobials in municipal wastewaters and provides new insights into the contributions of physico-chemical and biological processes to the overall removal of these compounds during the conventional wastewater treatment. We hope that you will find the manuscript suitable for publication in your journal.

The submitted files include the main text (6629 words), 6 tables, 3 Figures, Supplementary Information and Highlights.

Please send all further correspondence to me (ahel@irb.hr).

Sincerely,

Prof. Marijan Ahel

*Occurrence and fate of dissolved and particulate antimicrobials in
municipal wastewater treatment*

Ivan Senta, Senka Terzic and Marijan Ahel

Division for Marine and Environmental Research, Rudjer Boskovic Institute, Bijenicka c. 54,
10000 Zagreb, Croatia

Highlights:

- Dissolved and particulate antimicrobials were studied in municipal wastewater
- Improved assessment of the removal mechanisms in sewage treatment was achieved
- A significant percentage of fluoroquinolones and macrolides was removed by sorption
- Removal that could be attributed to biological transformation was relatively poor

**OCCURRENCE AND FATE OF DISSOLVED AND PARTICULATE
ANTIMICROBIALS IN MUNICIPAL WASTEWATER TREATMENT**

Ivan Senta, Senka Terzic, Marijan Ahel

*Rudjer Boskovic Institute, Division for Marine and Environmental Research, Bijenicka cesta
54, 10000 Zagreb, Croatia*

Corresponding author:

Dr. Marijan Ahel

Division for Marine and Environmental Research

Rudjer Boskovic Institute

10000 Zagreb, Croatia

Tel. +385-1-4560940

Fax: +385-1-4680242

e-mail: ahel@irb.hr

Abstract

Comprehensive study on the occurrence and fate of several classes of antimicrobials, including sulfonamides, trimethoprim, fluoroquinolones and macrolides, in Croatian municipal wastewaters was performed using an integrated approach, which comprised analysis of both dissolved and particulate fractions. A nation-wide screening showed ubiquitous occurrence of human-use antimicrobials in raw wastewater samples with the total concentrations ranging from 2 to 20 µg/L, while veterinary antimicrobials were typically present in much lower concentrations (<100 ng/L). The percentage of the particulate fraction in raw wastewater varied significantly depending on the type of the antimicrobial and the load of suspended solids. A detailed study of the mass flows of dissolved and particulate antimicrobials, performed in the wastewater treatment plant of the city of Zagreb, allowed an improved assessment of the biological and physico-chemical removal mechanisms of investigated compounds during the conventional activated sludge treatment. The overall removal efficiencies of antimicrobials from the water phase were rather variable, ranging from 0% for trimethoprim to 85% for norfloxacin. A significant percentage of fluoroquinolones (norfloxacin and ciprofloxacin) and macrolides (azithromycin and clarithromycin) was associated with the primary and excess secondary sludge, explaining 14 to 77% of the total removal. The removal, which could be attributed to biological transformation, was relatively poor for all antimicrobials, exceeding 30% only for SMX (32%) and clarithromycin (55%).

Keywords: antimicrobials; fluoroquinolones; macrolides; sulfonamides; trimethoprim; wastewater treatment

1 **1. Introduction**

2 Pharmaceutical compounds represent one of the most important classes of emerging
3 contaminants. In the last decade, they have been detected in different environmental
4 compartments all over the world (Segura et al., 2009). Traces of pharmaceuticals in the
5 environment raised concerns due to their high biological activity and there is an accumulating
6 evidence in the literature that they can cause different adverse effects in non-target species
7 (Fent et al., 2006).

8 Among different classes of pharmaceuticals, antimicrobial agents are under special
9 scrutiny due to the possible formation of the resistant bacterial strains, which can pose a
10 serious threat to the human health (da Costa et al., 2006). Antimicrobials represent the third
11 biggest group among all pharmaceuticals in human medicine and the most prominent group in
12 veterinary medicine (Thiele-Bruhn, 2003), with the estimated annual world consumption
13 between 100,000 and 200,000 tons (Wise, 2002). Antimicrobials with the highest
14 consumption are β -lactams, followed by tetracyclines, macrolides, fluoroquinolones and
15 sulfonamides. However, due to the instability of the β -lactam ring, which is easily
16 hydrolyzed, either chemically or microbiologically, penicilins and cephalosporins are not
17 considered to be a potential threat for the environment (Cha, 2006). On the contrary, due to
18 their comparatively higher stability, sulfonamides, fluoroquinolones and macrolides are
19 widely recognised as important classes of environmental contaminants (Segura et al., 2009).

20 The most important global sources of antimicrobials in the environment are municipal
21 wastewaters. After the intake, antimicrobials are only partially metabolised and their residues,
22 excreted via urine or feces, reach municipal wastewater treatment plants (WWTPs). In
23 addition, it was suggested that a significant percentage of pharmaceuticals in the sewage may
24 derive from the illegal disposal of unused medications down the drain (Daughton and Ruhoy,
25 2009). Numerous reports throughout the world (Hirsch et al., 1999; Miao et al., 2004;

26 Lindberg et al., 2005; Karthikeyan and Meyer, 2006; Xu et al., 2007; Gulkowska et al., 2008;
27 Watkinson et al., 2009; Zuccato et al., 2010; Gros et al., 2010; Gracia-Lor et al., 2012)
28 showed that antimicrobials represent ubiquitous contaminants of municipal wastewaters.
29 Recent study on the occurrence and fate of emerging contaminants in wastewaters of the
30 Western Balkan Region, showed widespread occurrence of different classes of emerging
31 contaminants, including many representatives of antimicrobials (Terzic et al., 2008).

32 Most of the literature reports suggested that the current design of conventional
33 WWTPs does not assure a complete removal of pharmaceuticals, including several classes of
34 antimicrobials (Castiglioni et al., 2006; Watkinson et al., 2007; Xu et al., 2007; Kasprzyk-
35 Hordern et al., 2009; Gros et al., 2010). Moreover, treatment of municipal and industrial
36 wastewaters in many less developed countries, including Croatia, is often incomplete and
37 consists only of the mechanical treatment (Kastelan-Macan et al., 2007). As a consequence, a
38 large percentage of pharmaceuticals introduced into WWTPs is released into the aquatic
39 environment and pose a significant threat to the receiving ambient waters. Despite a growing
40 number of studies on pharmaceutical in WWTPs, only few studies addressed the issue of
41 particle bound antimicrobials (Göbel et al. 2005; Okuda et al., 2009; Jelic et al., 2011). The
42 reason for that is a general perception that most of the pharmaceuticals belong to
43 comparatively polar compounds, which are primarily expected to occur in the dissolved
44 phase. Moreover, determination of organic micropollutants in the complex solid matrices is
45 often analytically more challenging than their determination in the dissolved fraction.
46 Nevertheless, comprehensive approach, which includes both dissolved and particulate
47 fraction, is the key prerequisite for the accurate assessment of the pharmaceutical behaviour in
48 the wastewater treatment and for understanding their ultimate fate in the aquatic environment.

49 The aim of this work was to investigate the occurrence of several important classes of
50 antimicrobials, including sulfonamides, trimethoprim, fluoroquinolones and macrolides, in

51 raw municipal wastewaters, and their fate in conventional wastewater treatment. Unlike most
52 of the reports in the literature, our study involved simultaneous determination of these
53 compounds in both dissolved and particulate fraction, thus providing a basis for a
54 comprehensive assessment of the physico-chemical and biological removal processes.

55

56 **2. Experimental**

57

58 *2.1. Target compounds*

59 This study was focused on antimicrobials with the extensive usage in human medicine
60 in Croatia, but some additional compounds, including several representatives of veterinary
61 antimicrobials, and the main human metabolite of sulfamethoxazole *N*-
62 acetylsulfamethoxazole, were also included. The list of all target compounds, together with
63 their abbreviations and usage is presented in Table 1.

64

65 *2.2. Sampling*

66 All samples were collected in clean amber glass bottles pre-rinsed with methanol and
67 ultrapure water. The nation-wide screening was performed in April and May 2005. Raw
68 wastewater (RW) samples were collected in the largest Croatian cities (Table S1;
69 Supplementary information). Additionally, where available, effluents from wastewater
70 treatment plants were also collected. During this sampling campaign only a few cities
71 (Bjelovar, Cakovec, Varazdin, Velika Gorica and Vinkovci) had facilities for full mechanical
72 and biological treatment, while wastewaters of the cities of Rijeka and Split were treated only
73 mechanically. For those cities, which did not have any wastewater treatment facilities, grab
74 samples were collected directly from the sewerage system. In WWTPs of the cities Cakovec
75 and Bjelovar 24-hour flow-proportional composite samples were collected using automatic

76 devices, while on some other locations (Rijeka, Split, Varazdin) composite samples were
77 prepared by mixing grab samples taken over a diurnal cycle.

78 A detailed study on the occurrence and fate of antimicrobials during conventional
79 wastewater treatment was performed in the central WWTP of the city of Zagreb, which is
80 fully operational since 2007. This WWTP receives a combined municipal and industrial
81 sewage from the entire city and includes full mechanical and biological treatment based on
82 conventional activated sludge (CAS) treatment. It has a designed capacity of 1000000
83 population equivalents and currently serves about 750,000 inhabitants (Schröder et al., 2001).
84 The average hydraulic load of raw wastewater is about 250,000 m³/day. The details about the
85 WWTP are summarized in Table S2 (Supplementary information). During the study in this
86 WWTP, several sampling campaigns in the period from March to September 2009 were
87 performed, in which twenty-four-hour composite samples of both RW and biologically treated
88 wastewater (secondary effluent, SE) were collected. In addition, for the investigation of the
89 partitioning behaviour of the selected antimicrobials during CAS treatment, activated sludge
90 samples were taken directly from the aeration tank.

91

92 *2.3. Sample pre-treatment and instrumental analysis*

93 All samples were filtered through 2.7 µm glass fiber filters (GF/D, Whatman, USA)
94 immediately after being brought back to the laboratory, typically within 3 hrs after sampling.
95 Filtrates were stored in the dark at 4 °C and extracted as soon as possible, typically within 24
96 hours. Filters containing particulate fraction were kept frozen at -18 °C until the extraction.
97 Dissolved fraction was enriched using solid-phase extraction (SPE) on Oasis HLB cartridges
98 using the protocol described in our previous paper (Senta et al., 2008). Briefly, samples were
99 acidified with formic acid to obtain pH 3. After percolation of the samples, adsorbed analytes
100 were eluted from the cartridges with 4 mL of 1% ammonia solution in methanol. After

101 evaporation to dryness, residue was dissolved in 0.1% formic acid just before the instrumental
102 analysis. Absolute recoveries of individual antimicrobials from wastewater samples were
103 between 49% and 119%, with good repeatabilities (1% – 18%).

104 For the determination of antimicrobials incorporated in particulate fraction,
105 pressurized solvent extraction (PLE) followed by subsequent extract cleanup on Oasis HLB
106 columns were used (Senta, 2009). Briefly, samples were extracted using PLE with the mixture
107 of 50 mM *o*-phosphoric acid and methanol (50:50). Temperature was 100 °C and pressure
108 was 138 bars. 3 cycles were performed with static time of 5 minutes. Preheating time was 5
109 minutes and flush volume 120%. PLE extracts were diluted to approximately 300 mL with
110 ultrapure water and additionally extracted on Oasis HLB cartridges using the same protocol as
111 for the dissolved fraction. Absolute recoveries for the analysis of the particulate
112 antimicrobials were between 35% and 65%, with repeatabilities between 1% and 12%. The
113 analyte losses during the extraction procedures and extract work-up, as well as matrix effects,
114 were compensated by using several surrogate standards (at least one for each antimicrobial
115 class), which were added in the samples prior the extraction (SPE or PLE). Instrumental
116 standards sulfamerazine and josamycin were added in the final extracts just prior to analysis
117 as a control of instrument performance.

118 Our previously described method (Senta et al., 2008) based on liquid chromatography
119 – tandem mass spectrometry (LC-MS/MS) was used for determination of the selected
120 antimicrobials. Target compounds were separated on C₁₈ column in reversed-phase system
121 using gradient elution with water and methanol both acidified with 0.1% of formic acid. The
122 mass spectrometric analyses were performed on a TSQ Quantum triple quadrupole instrument
123 (Thermo Electron, San Jose, USA). Electrospray ionization in positive mode was used for the
124 production of ions. Detection and quantification of all analytes were performed using selected
125 reaction monitoring (SRM). Instrumental detection limits were between 1 and 17 pg, while

126 method detection limits were between 1.0 and 13.2 ng/L and 0.5 and 6.6 ng/L for raw
127 wastewater and secondary effluent samples, respectively. The original method was later
128 slightly modified in order to include deuterated azithromycin as a surrogate standard for
129 azithromycin and other macrolides, which significantly improved the reliability of
130 quantification of these compounds in heavily loaded wastewater samples. Furthermore, during
131 the study performed in the Zagreb WWTP, the major metabolite of SMX *N*-
132 acetylsulfamethoxazole, was also included in the method.

133 In addition, ultrahigh performance liquid chromatography (UHPLC, Waters, USA)
134 coupled to quadrupole-time of flight mass spectrometry (QToF MS; Micromass, Manchester,
135 UK) was applied in this study for the confirmation purposes, such as confirmation of the
136 presence of erythromycin and *N*-acetyl derivatives of sulfonamides.

137

138 **3. Results and discussion**

139

140 *3.1. Screening of antimicrobials in Croatian municipal wastewaters*

141 Occurrence of antimicrobials in raw municipal wastewater samples collected during
142 the nation-wide campaign in Croatia is presented in Table 2 (dissolved fraction) and Table 3
143 (particulate fraction). As can be seen, several antimicrobials were detected in all of the
144 analysed samples. The most prominent compounds in the dissolved fraction were SMX, TMP,
145 NOR, CIP, AZI, SPY and ERY-H₂O. The concentrations of major antimicrobials were
146 typically found in concentration range between 100 and 1000 ng/L, with only few exceptions
147 reaching into in the low ug/L range. The highest individual concentration was determined for
148 SMX in the wastewater of Novi Zagreb (11.6 µg/L). These results show a strong
149 predominance of antimicrobials used in the human medicine, while typical veterinary
150 antimicrobials belong to the minor components. The observed levels of antimicrobials for

151 human use are in a good agreement with the official reports on the consumption figures for
152 top prescribed human-use antimicrobials in Croatia (Ferech et al., 2006). The only exception
153 is SPY, which is rarely used as antimicrobial agent for human use. However, SPY is the main
154 metabolite of sulfasalazine, which has a significant usage in human medicine in Croatia.
155 ERY-H₂O is the metabolite and the degradation product of macrolide antibiotic erythromycin
156 and represents its strongly predominant form in raw wastewater and wastewater effluents
157 (Terzic et al., 2008). The results of preliminary screening are in a good agreement with the
158 reports on the occurrence of antimicrobials in raw municipal wastewaters in other countries
159 (Göbel et al., 2005; Lindberg et al., 2005; Watkinson et al., 2009; Zuccato et al., 2010, Gros et
160 al., 2010), with some differences which can be explained with the different local consumption
161 patterns. For example, relatively high concentrations of AZI in Croatian wastewaters can be
162 explained with the fact that this antibiotic is among the top-prescribed pharmaceuticals in
163 Croatia. On the other hand, ROX, which is frequently detected macrolide antibiotic in
164 wastewaters (Göbel et al., 2004; Xu et al, 2007), was rarely detected in our samples due to its
165 very limited usage in Croatia.

166 In contrast to human-use antimicrobials, compounds used exclusively in veterinary
167 medicine (STZ, SMZ, ENR) were detected only occasionally and in much lower
168 concentrations (usually below 100 ng/L). This was actually expected since this study was
169 focused on municipal wastewaters, characterised by low inputs from agricultural sources. The
170 only exception was the city of Belisce, located in an area characterised with the extensive
171 livestock production. As a consequence, wastewater from this city contained enhanced
172 concentration of veterinary antimicrobial SMZ (175 ng/L).

173 The most prominent antimicrobials associated with the particulate fraction were
174 generally the same as in the dissolved fraction, with the exception of SPY, which could not be
175 quantitatively determined in the particulate fraction (Table 3). However, distribution of

176 antimicrobials in the particulate fraction was somewhat different than the distribution in the
177 dissolved fraction, which can be explained by their different physico-chemical properties
178 (Table 1). The percentage of antimicrobials associated with the particulate fraction, ranged
179 from 1% for sulfonamides up to 35% for norfloxacin. The concentration of fluoroquinolones
180 and azithromycin, expressed as mass fraction in the total suspended solids exceeded 1 $\mu\text{g/g}$
181 (Table S3, Supplementary information).

182 The concentrations of individual antimicrobials in secondary effluents were generally
183 of the same order of magnitude as their concentrations in raw wastewater, collected at the
184 same location, indicating their incomplete removal during conventional wastewater treatment
185 (Table S4, Supplementary information). Incomplete removal was observed even in the
186 WWTPs with the full mechanical and biological treatment (Cakovec, Bjelovar), which
187 emphasized recalcitrant nature of investigated compounds. It should be noticed that for the
188 effluent samples only the concentration in the dissolved fraction was determined.

189

190 *3.2. Occurrence of antimicrobials in WWTP of the city of Zagreb*

191 Detailed study on the occurrence and fate of antimicrobials during conventional
192 wastewater treatment was performed in the central WWTP of the city of Zagreb during 2009,
193 after this WWTP became fully operational, including both mechanical and biological
194 treatment. In three sampling campaigns 36 RW and 34 SE samples were collected. Summary
195 of the results for the dissolved fraction of RW is presented in Table 4. As can be seen, 9
196 compounds were detected in all analyzed samples and the most prominent antimicrobials
197 were generally the same as in the preliminary screening of the Croatian wastewaters. In
198 addition, *N*-Ac-SMX, the most prominent metabolite of SMX (Baselt, 2008) and macrolide
199 antibiotic CLA were also included in the method during this study. The highest average
200 concentration in RW was determined for *N*-Ac-SMX (656 ± 87 ng/L), followed by AZI

201 (502±315 ng/L), SMX (484±189 ng/L) and fluoroquinolones NOR (339±181 ng/L) and CIP
202 (333±197 ng/L). The average concentrations of SPY, TMP and CLA were also higher than
203 100 ng/L.

204 The most prominent compounds in the particulate fraction of RW samples from the
205 Zagreb WWTP (Table 5) were NOR, CIP, AZI and CLA with the concentrations ranging
206 from several hundreds ng/L to a few µg/L. Other human-use antimicrobials or their
207 metabolites were also present, but in much lower concentrations, typically below 50 ng/L. As
208 can be seen in Fig. 1A, the contributions of the particulate antimicrobials to the total
209 concentrations in RW varied in a very wide range (from 0.4% to 47 %). All sulfonamides and
210 trimethoprim were characterised by a strong predominance in the dissolved fraction (>95%).
211 On the contrary, the percentage of particulate macrolides and fluoroquinolones was
212 significant and varied from 7% for ERY-H₂O to 47% for NOR. As a consequence, ignoring
213 contribution of their particulate fraction in raw wastewater, can lead to significant
214 underestimation of the real input for these antimicrobials into a WWTP.

215

216 *3.3. Partitioning of antimicrobials during wastewater treatment*

217 Behaviour of organic contaminants during the conventional wastewater treatment can
218 be significantly affected by physico-chemical processes, in particular sorption onto suspended
219 solids (Ternes et al., 2004). Simultaneous determination of antimicrobials in dissolved and
220 particulate fractions provided a basis for the determination of their distribution coefficients
221 (K_d) and, consequently, the assessment of relative importance of the particulate fraction,
222 which is usually overlooked in the studies of the occurrence and fate of antimicrobials in
223 wastewater treatment. In this study we estimated the average distribution coefficients (K_d) of
224 individual antimicrobials in two typical compartments of wastewater treatment, raw
225 wastewater and mixed liquor from the aeration tank, using the following formula:

226
$$K_d = \frac{C_s}{C_d * SS}$$

227 where C_s is the concentration of the compound in the particulate fraction (suspended solids in
228 RW or activated sludge in aeration tank) ($\mu\text{g/L}$), C_d is concentration of the compound in the
229 dissolved fraction ($\mu\text{g/L}$) and SS is suspended solid concentration in raw wastewater or
230 activated sludge concentration in the aeration tank (Ternes et al., 2004). The results are
231 presented in Table 6. It should be noticed that sorption coefficients were calculated only for
232 the major antimicrobials, which could be quantitatively determined in both fractions.
233 Moreover, K_d values for RW shown in Table 6, represent an average value from 31
234 determinations, except for *N*-Ac-SMX ($n = 18$), while the average value for the aeration tank
235 was calculated from 2 independent experiments. As expected, distribution coefficients of
236 investigated antimicrobials varied in a broad range, from 16 L/kg for *N*-Ac-SMX to 6959 for
237 NOR. It should be noted that K_d values are not positively correlated with octanol-water
238 partition coefficients (Table 1; Fig.S1), indicating that mechanisms other than lipophilic
239 partitioning played a predominant role for the strongly adsorbable antimicrobials (Okuda et
240 al., 2009). However, it is well-known that these coefficients can be strongly affected by the
241 type, composition and pH of the activated sludge (Ternes et al., 2004).

242 Calculated distribution coefficients for fluoroquinolones (NOR and CIP) were
243 somewhat lower than the values reported by Golet et al. (2002), but were significantly higher
244 than those for the other antimicrobial groups. As mentioned above, these enhanced values
245 were not in correlation with their $\log K_{ow}$ values, indicating the predominant role of polar
246 and/or ionic interactions for the adsorption process. Some authors suggested that sorption of
247 these compounds onto the activated sludge may be accomplished by the electrostatic
248 interactions with the cell membranes of the microorganisms forming activated sludge (Xu et
249 al., 2007).

250 Similarly, it seems that pH-dependent speciation of sulfonamides and trimethoprim in
251 wastewater played a role in their partitioning. At the pH, typically found in WWTP of the city
252 of Zagreb (7.11–7.94), SMX ($pK_{a,2}=5.6$) is mainly present in the anionic form, while for
253 trimethoprim ($pK_{a,2}=6.8$) and especially sulfapyridine ($pK_{a,2}=8.4$) non-ionic species are much
254 more abundant. The results suggested a more efficient adsorption of non-ionic species.
255 However the differences between K_d values for individual sulfonamides and TMP cannot be
256 explained by their lipophilicity expressed as $\log K_{ow}$ values, indicating importance of polar
257 interactions.

258 Unlike other antimicrobials in this study, macrolides are present in wastewater mainly
259 as positively charged species ($pK_{a,1}=8.7-8.9$). Sorption coefficients for AZI ($K_d=486$ and
260 2156) and CLA ($K_d=386$ and 636) were found to be higher than sorption coefficients for
261 sulfonamides, but lower than for fluoroquinolones. The K_d values determined in this study are
262 in a good agreement with the literature data (Göbel et al., 2005). Higher sorption coefficient
263 for AZI as compared with CLA is supported by its higher $\log K_{ow}$ value (4.0). Moreover, it
264 should be taken into account that AZI has two basic sites (nitrogen atoms), while CLA has
265 only one basic site. Such properties could promote the ionic interactions with the negatively
266 charged suspended matter present in wastewater. The distribution coefficient for ERY-H₂O
267 was significantly lower than the sorption coefficients for the other two macrolides, so that the
268 concentration of this compound in the particulate fraction was usually close to or below the
269 quantification limit. At this stage, we cannot find a plausible explanation for such behaviour.

270 It is generally expected that the percentage antimicrobials in the particulate fraction
271 should be correlated with the concentration of suspended solids. In our samples this
272 relationship was proven to be statistically significant for sulfonamides, trimethoprim and
273 macrolides, but not for fluoroquinolones (Fig S2, Supplementary information). This suggests

274 that for fluoroquinolones the variable character of particulate organic matter in RW must have
275 played a significant role.

276 Taking into account a considerable variability of K_d s for all of the investigated
277 antimicrobials among individual samples, it is interesting to note that the average K_d values
278 for RW suspended solids and activated sludge were reasonably similar with no obvious trend
279 regarding relationship between the two solid phases. However, it should be noted that the
280 sorption coefficients for macrolides in mixed liquor were significantly lower than those in
281 RW. Nevertheless, as a consequence of high concentration of suspended solids in the aeration
282 tank (3.5 g/L), the percentage of particle-bound antimicrobials in mixed liquor is extremely
283 high (Fig 1B). For example, more than 90% of fluoroquinolones in the aeration tank were
284 associated with mixed-liquor suspended solids (MLSS). Even for the compounds with
285 relatively low K_d values, such as SMX and TMP, the percentage of antimicrobials in the
286 particulate fraction was 45 and 70%, respectively. However these percentages represent only
287 a snapshot of the partitioning process in the aeration tank and should not be interpreted as the
288 removal efficiencies because only the excess sludge contributes to the removal as it will be
289 discussed below.

290

291 *3.4. Occurrence in secondary effluents and assessment of removal mechanisms*

292 The average concentrations of antimicrobials in secondary effluents of the Zagreb
293 WWTP are presented in Table 4 along with the RW concentrations. It should be noted that
294 only dissolved antimicrobials were determined in secondary effluents, since their estimated
295 concentration in the particulate phase was less than 2% for all target analytes, including
296 highly adsorbable fluoroquinolones, due to the very low concentration of suspended solids in
297 SE (typically below 5 mg/L; Table S2; Supplementary information). As can be seen, all major
298 antimicrobials found in RW were detected in all analysed SE samples indicating their

299 incomplete removal. The most prominent compounds in SE were AZI (350 ± 206 ng/L), SMX
300 (323 ± 135 ng/L) and *N*-Ac-SMX (214 ± 177 ng/L).

301 The average removal efficiencies, presented in Fig. 2, were calculated using two
302 different approaches. In the first approach, which is common in the literature (Watkinson et
303 al., 2007; Vieno et al., 2007; Kasprzyk-Hordern et al., 2009; Gros et al., 2010), the removal
304 efficiencies of individual antimicrobials were calculated by comparing secondary effluent
305 concentrations with the input concentrations in the dissolved phase (Table 4). However, this
306 approach can lead to significant errors for the compounds having high tendency for the
307 sorption onto the solid particles (Deo and Halden, 2010). Along these lines, the overall
308 removal efficiencies were alternatively calculated using a more realistic approach taking the
309 total concentration in RW as a basis. The overall elimination varied significantly among
310 different classes of antimicrobials. As expected, removal efficiencies using the second
311 approach were significantly higher for strongly adsorbable compounds.

312 The highest removal rates (more than 80% using the second approach) were achieved
313 for fluoroquinolones and they are in good agreement with most of the previously reported
314 values for this group of antimicrobials (Golet et al., 2002; Lindberg et al., 2005; Castiglioni et
315 al., 2006; Vieno et al. 2007; Watkinson et al., 2007; Xu et al., 2007).

316 It should be pointed out that literature data on the removal efficiencies for
317 sulfonamides and macrolides are not as consensual as for fluoroquinolones. For example, the
318 average removal rates for sulfonamides SMX and SPY in our study were 34%. This is
319 comparable with the removal efficiency for SMX in Swedish WWTPs (Lindberg et al., 2005)
320 and in WWTPs in the South China (Xu et al., 2007). However, Göbel et al. (2007) reported
321 low and highly variable removal rates for both SMX and SPY during the secondary treatment
322 in two Swiss WWTPs. In fact, concentrations of SMX and SPY in their study were up to two
323 times higher in SE than in primary effluent, leading to the negative removal rates. Authors

324 suggested that such results could be explained with the re-transformation of sulfonamide
325 metabolites during the wastewater treatment. Our study confirmed the importance of *N*-Ac-
326 SMX for the total load of SMX derived compounds, and the calculated removal efficiency for
327 *N*-Ac-SMX (77%) was comparable with the removal rates reported by Göbel et al. (2007).
328 Consequently, it is possible that re-transformation of this metabolite affected the overall
329 removal rate for SMX, but not to such extent that would have lead to a negative elimination.
330 It is reasonable to assume that similar mechanism could apply for SPY as well, however our
331 attempt to identify *N*-acetylsulfapyridine in our wastewater extracts using UPLC-QToF
332 technique, indicated no significant presence of this metabolite.

333 Regarding macrolides, the highest removal efficiency was observed for the macrolide
334 antibiotic CLA (69%), which is in a very good agreement with the value reported by Zuccato
335 et al. (2010). The removal efficiencies for the other two investigated macrolides, AZI and
336 ERY-H₂O were significantly lower. Low removal efficiencies for macrolides, including
337 observations on the negative elimination, have already been reported in the literature (Göbel
338 et al., 2007; Xu et al., 2007, Gulkowska et al., 2008; Gros et al., 2010). Since the presence of
339 de-conjugable metabolites of these compounds (Baselt, 2008) was never reported in the
340 literature, some authors suggested that low removal rates could be explained with the fact that
341 significant portion of these compounds are excreted from the body via bile and feces, so it is
342 possible that overall content of these compounds that enters wastewater treatment plant is
343 underestimated (Göbel et al., 2007). This assumption, however, was never experimentally
344 verified. The results from this study showed that 30% of AZI in RW was adsorbed onto the
345 solid particles. It should also be pointed out that the overall removal efficiency for AZI, based
346 on the total concentration in RW was 51%, while the removal efficiency of ERY-H₂O was
347 even lower (24%).

348 Virtually, no removal was observed for TMP. Other authors also reported low removal
349 efficiencies for this compound, including some negative values as well (Lindberg et al., 2005;
350 Göbel et al., 2007; Gulkowska et al., 2008; Gracia-Lor et al., 2012). It should be noted that
351 97% of TMP is excreted from the human body in the urine, mostly (80% – 90%) unchanged
352 (Baselt, 2008), so that it is very difficult to speculate about possible reasons for the negative
353 removal.

354 The overall removal is the result of combined biological and physicochemical
355 elimination. In order to decouple this two mechanisms a simple mass balance was made by
356 determining mass flows of major antimicrobials in the dissolved and particulate phase. The
357 removal particle-bound fraction consisted of adsorption onto the primary sludge (150 g m^{-3})
358 and the adsorption onto the excess sludge (158 g m^{-3}). The result for the most abundant
359 representatives of antimicrobials SMX, NOR and AZI are presented in Fig. 3, while for the
360 other antimicrobials the results are shown in the supplementary information (Fig. S3;
361 Supplementary information). The physico-chemical elimination was assessed from the mass
362 flows in the primary sludge and excess activated sludge, assuming that the contribution of
363 other physico-chemical processes, such as volatilization, was negligible (Senta et al., 2011).
364 The physico-chemical removal varied from 2% for SMX to 78% for NOR and reflects the
365 partitioning behaviour described above. The results are generally in agreement with the
366 literature reports, which suggested that for compounds with sorption coefficients below 300
367 L/kg, sorption onto the secondary sludge particles is not relevant for their elimination (Joss et
368 al., 2005).

369 The contribution of biological elimination was calculated as the difference between
370 the overall elimination and physico-chemical elimination. The lowest biological elimination
371 was obtained for TMP (-7%), suggesting its possible formation from yet unknown precursors.
372 Moreover, low biological elimination of fluoroquinolones (8 – 22%) clearly confirmed earlier

373 literature reports on the strong predominance of physico-chemical partitioning in their
374 removal during the wastewater treatment (Lindberg et al., 2005). Compared to
375 fluoroquinolones, the biological removal of sulfonamides and macrolides was more efficient
376 however even for the most biodegradable compound CLA it reached only 55%.

377

378 **Conclusion**

379

380 This study showed a widespread occurrence of antimicrobials in raw municipal
381 wastewater in Croatia, with strong predominance of human use antibiotics. Separate analysis
382 of the dissolved and particulate fractions allowed an improved assessment of the biological
383 and physico-chemical removal mechanisms of investigated compounds during the
384 conventional activated sludge treatment. The results pointed out the importance of the
385 sorption onto solid particles, especially for fluoroquinolones and macrolides. It was shown
386 that the partitioning behaviour of antimicrobials cannot be predicted on the basis of simple
387 parameters such as $\log K_{ow}$, which warrants careful study of their mass flows in the
388 wastewater treatment for each individual compound. The biological elimination in the
389 conventional activated sludge treatment was rather modest for most of the antimicrobials,
390 indicating that for a more efficient removal advanced treatment technologies have to be
391 considered.

392

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394

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401

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Figure captions

Fig. 1 Percentage of antimicrobials associated with suspended solids in the WWTP of the city of Zagreb: (A) raw wastewater and (B) mixed liquor from aeration tank. SPY = sulfapyridine; SMX = sulfamethoxazole; N-Ac-SMX = *N*-acetylsulfamethoxazole; TMP = trimethoprim; NOR = norfloxacin; CIP = ciprofloxacin; AZI = azithromycin; ERY-H₂O = dehydroerythromycin; CLA = clarithromycin.

Fig. 2 Overall removal efficiencies of antimicrobials in the WWTP of the city of Zagreb determined a) based on dissolved fraction only (1st approach); b) based on the total concentration in raw wastewater (2nd approach). SPY = sulfapyridine; SMX = sulfamethoxazole; N-Ac-SMX = *N*-acetylsulfamethoxazole; TMP = trimethoprim; NOR = norfloxacin; CIP = ciprofloxacin; AZI = azithromycin; ERY-H₂O = dehydroerythromycin; CLA = clarithromycin.

Fig. 3 Mass flows of dissolved and particulate antimicrobials in WWTP of the city of Zagreb and assessment of the biological elimination; d – dissolved fraction; p – particulate fraction; PC – primary clarifier; SC – secondary clarifier; RW – raw wastewater; SE – secondary effluent; PS – primary sludge; ES – excess sludge; SMX - sulfamethoxazole; NOR – norfloxacin; AZI – azithromycin.

Table 1[Click here to download Table: Table 1_f.docx](#)**Table 1.** List of antimicrobials included in the present study along with their physico-chemical properties¹.

Antimicrobial class	Analyte	Acronym	Usage	log K_{ow}	p <i>K</i> _{a,1}	p <i>K</i> _{a,2}
Sulfonamides	Sulfadiazine	SDZ	Human + veterinary	-0.1	2.1	6.4
	Sulfathiazole	STZ	Veterinary	0	2.0	7.1
	Sulfapyridine	SPY	Human	0.4	2.6	8.4
	Sulfamethazine	SMZ	Veterinary	0.3	2.1	7.5
	Sulfamethoxazole	SMX	Human	0.9	1.9	5.6
	<i>N</i> -acetylsulfamethoxazole	<i>N</i> -Ac-SMX	Metabolite of SMX	1.5	5.0	-
	Trimethoprim ²	TMP	Human + veterinary	0.9	3.2	6.8
Fluoroquinolones	Norfloxacin	NOR	Human	-1.0	3.1	8.6
	Ciprofloxacin	CIP	Human	0.3	3.0	8.7
	Enrofloxacin	ENR	Veterinary	0.9	3.9	7.6
Macrolides	Azithromycin	AZI	Human	4.0	8.7	9.5
	Dehydroerythromycin	ERY-H ₂ O	Human + veterinary ³	3.0 ⁴	8.9 ⁴	-
	Clarithromycin	CLA	Human	1.8	8.9	-
	Roxithromycin	ROX	Human	nd	9.2	-

¹Data on physico-chemical properties are from references: Vazquez et al., 2001; Göbel et al., 2004; Qiang and Adams, 2004; Batt and Aga, 2005; Stoob et al., 2005; Diaz-Cruz et al., 2006; Vieno et al, 2006; Choi et al., 2007; Gros et al., 2007; Sibley and Pedersen, 2008.

² used in combination with sulfonamides; ³ metabolite of erythromycin (ERY); ⁴ data for ERY

Table 2[Click here to download Table: Table 2_f.docx](#)**Table 2.** Occurrence of antimicrobials in Croatian raw municipal wastewaters in April and May 2005 – dissolved fraction (concentrations in ng/L).

Location	SDZ	STZ	SPY	SMZ	SMX	TMP	NOR	CIP	ENR	AZI	ERY-H ₂ O	ROX
Belisce	12	nd	97	175	613	35	nd ²	nd	nd	107	147	nd
Bjelovar	92	nd	385	nd	826	365	1231	196	16	122	66	nd
Cakovec	5	nd	370	nd	293	481	341	149	nd	281	58	nd
Karlovac	23	nd	364	nd	735	758	1843	169	nd	992	407	50
Novi Zagreb	29	4	809	7	11555	2551	2885	777	nd	799	122	nd
Osijek	16	nd	99	nd	1184	1817	1711	1079	nd	778	420	nd
Rijeka	15	nd	732	nd	1094	1045	1282	161	nd	352	171	nd
Sisak	96	nd	89	15	858	347	72	5	8	77	99	nd
Slavonski Brod	2	nd	292	nd	387	588	94	nd	nd	173	24	nd
Split	7	nd	567	nd	675	776	501	114	nd	482	213	nd
Split - sewer center	12	nd	507	nd	829	1068	1300	532	nd	1066	149	nd
Varazdin	132	1	376	2	640	3442	2009	496	18	1025	187	nd
Velika Gorica	28	nd	560	15	1944	2706	894	489	nd	501	275	nd
Vinkovci	12	nd	923	nd	943	659	1186	217	nd	317	48	nd
Zadar	8	nd	931	2	2033	2318	2937	2610	nd	1139	127	13
Zagreb	36	nd	313	17	720	840	581	227	nd	801	61	nd

nd – not detected

Table 3[Click here to download Table: Table 3_f.docx](#)**Table 3.** Occurrence of antimicrobials in Croatian raw municipal wastewaters – particulate fraction (concentrations in ng/L).

Location	SMX	TMP	NOR	CIP	AZI	ERY-H₂O
Bjelovar	1.7	3.4	400	48	73	1.3
Cakovec	nd	1.4	147	21	66	0.7
Karlovac	3.6	2.9	278	40	41	0.9
Novi Zagreb	58	15	680	166	93	0.9
Osijek	6.2	5.0	202	111	37	5.2
Rijeka	9.3	23	949	139	29	2.2
Sisak	3.4	6.1	113	9.3	1.3	1.7
Slavonski Brod	3.8	22	852	67	31	1.5
Split	18	38	492	116	56	nd
Split - sewer center	6.6	12	547	152	22	1.2
Varazdin	1.1	11	22	4.6	76	nd
Velika Gorica	6.5	12	196	56	63	3.5
Vinkovci	16	52	4380	491	36	20
Zadar	21	47	1294	1046	47	1.2
Zagreb	23	16	159	28	67	1.7

nd – not detected

Table 4[Click here to download Table: Table 4_f.docx](#)**Table 4.** Occurrence of antimicrobials in the wastewater treatment plant of the city of Zagreb – dissolved fraction.

Analyte	Raw wastewater				Secondary effluent			
	n ¹	FD ² (%)	Concentration (ng/L)		n ¹	FD ² (%)	Concentration (ng/L)	
			Range	Average ± SD			Range	Average ± SD
SPY	36	100	80 – 442	244 ± 111	34	100	48 – 288	161 ± 73
SMX	36	100	210 – 999	484 ± 189	34	100	119 – 544	323 ± 135
<i>N</i> -Ac-SMX	18	100	427 – 805	656 ± 87	17	100	14 – 486	214 ± 177
TMP	36	100	87 – 219	148 ± 34	34	100	75 – 245	155 ± 49
NOR	36	100	60 – 634	339 ± 181	34	100	24 – 175	97 ± 53
CIP	36	100	29 – 650	333 ± 197	34	100	11 – 168	87 ± 55
AZI	36	100	122 – 1634	502 ± 315	34	100	38 – 784	350 ± 206
ERY-H ₂ O	36	100	25 – 73	44 ± 12	34	100	15 – 59	36 ± 12
CLA	36	100	112 – 300	197 ± 61	34	100	25 – 133	71 ± 34

¹ Number of analyzed samples; ² Frequency of detection

Table 5[Click here to download Table: Table 5_f.docx](#)**Table 5.** Occurrence of antimicrobials in the municipal raw wastewater of the city of Zagreb – particulate fraction.

Analyte	n ¹	FD ² (%)	Concentration (ng/L)		Concentration (ng/g) ³	
			Range	Average ± SD	Range	Average ± SD
SPY	31	100	0.7 – 15.8	6.0 ± 3.5	2.1 – 49.4	23.1 ± 8.9
SMX	31	100	2.2 – 15.9	5.5 ± 2.9	8.1 – 41.6	21.1 ± 7.7
<i>N</i> -Ac-SMX	31	100	0.6 – 10.1	3.2 ± 2.1	1.9 – 26.5	12.5 ± 6.9
TMP	31	100	0.3 – 19.6	6.0 ± 4.4	0.8 – 52.8	22.2 ± 11.4
NOR	31	100	35 – 1924	324 ± 340	101 – 3026	1169 ± 630
CIP	30	100	22 – 887	180 ± 158	64 – 1395	670 ± 302
AZI	31	100	35 – 700	206 ± 149	103 – 1978	819 ± 409
ERY-H ₂ O	31	97	nd – 12.4	3.1 ± 2.7	nd – 38.6	12.0 ± 9.2
CLA	31	100	0.4 – 137	33.6 ± 31.9	1.2 – 322	125 ± 87.8

¹ Number of analyzed samples; ² Frequency of detection; ³ Expressed as mass fraction in suspended solids;

Table 6. Distribution coefficients (K_d) of the selected antimicrobials in raw wastewater and aeration tank mixed liquor.

Compound	Raw wastewater	Mixed liquor (aeration tank)
	K_d (L/kg) ¹	K_d (L/kg) ²
Sulfapyridine	132	454
Sulfamethoxazole	53	111
<i>N</i> -Acetylsulfamethoxazole	16	nd
Trimethoprim	163	316
Norfloxacin	6959	5974
Ciprofloxacin	4875	3934
Azithromycin	2156	486
Dehydrated erythromycin	325	114
Clarithromycin	636	386

K_d values were calculated as the ratio between the concentration of sorbed (in ng/kg) and dissolved (in ng/L) antimicrobials. ¹ average of 31 independent determinations in 24-hr composite samples of raw wastewater; ² measured in grab samples from the aeration tank – average of 2 independent determinations.

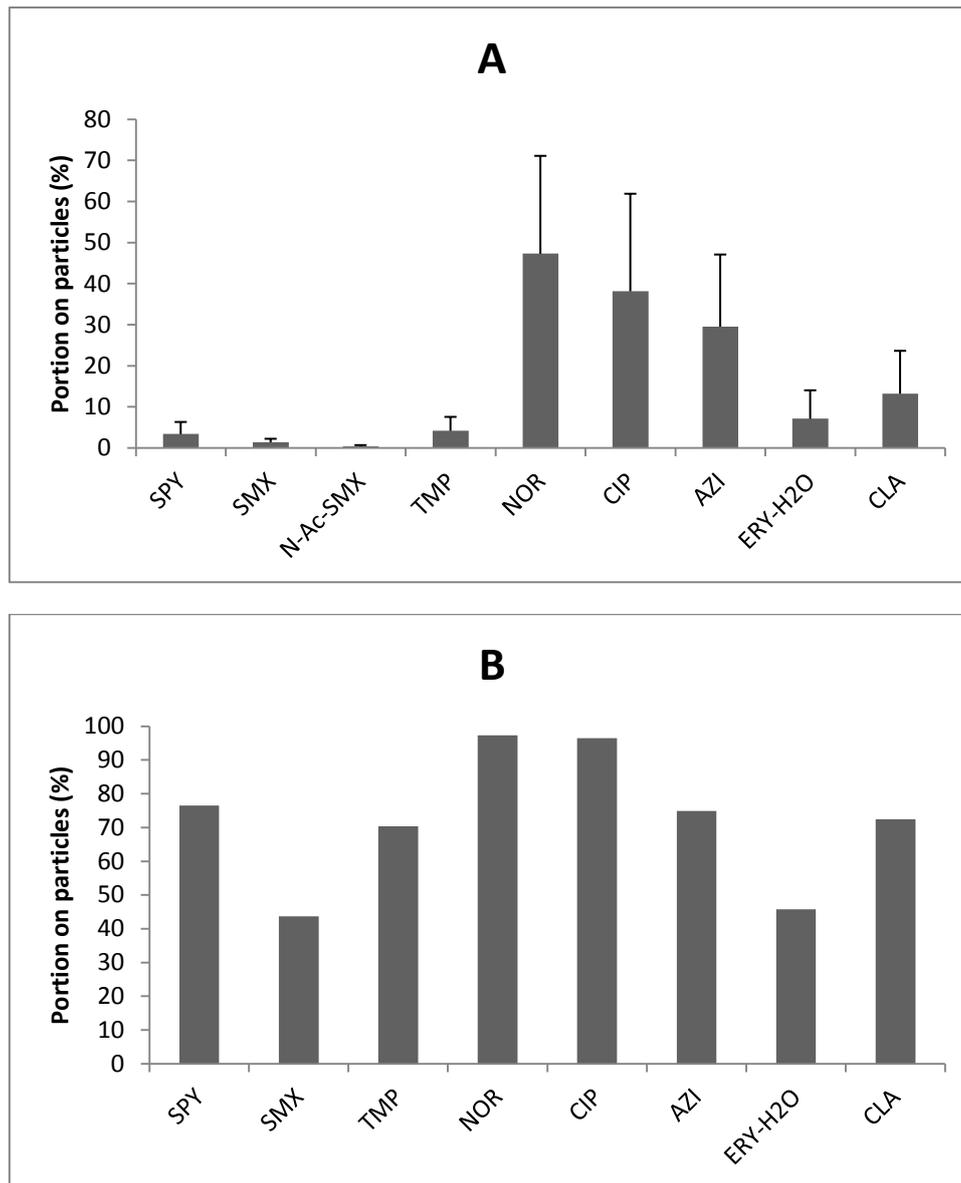


Fig. 1 Percentage of antimicrobials associated with suspended solids in the WWTP of the city of Zagreb: (A) raw wastewater and (B) mixed liquor from aeration tank. SPY = sulfapyridine; SMX = sulfamethoxazole; N-Ac-SMX = *N*-acetylsulfamethoxazole; TMP = trimethoprim; NOR = norfloxacin; CIP = ciprofloxacin; AZI = azithromycin; ERY-H2O = dehydroerythromycin; CLA = clarithromycin.

Figure 2

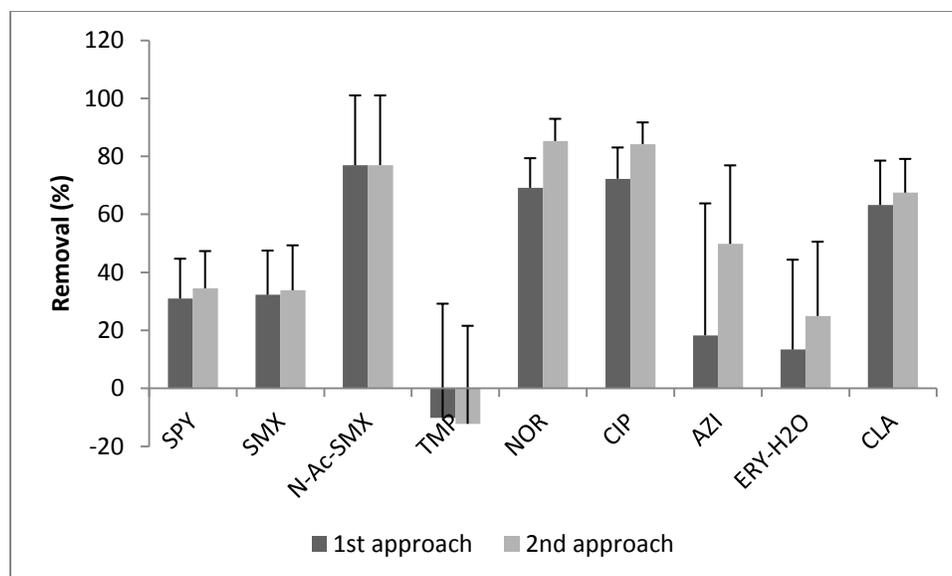
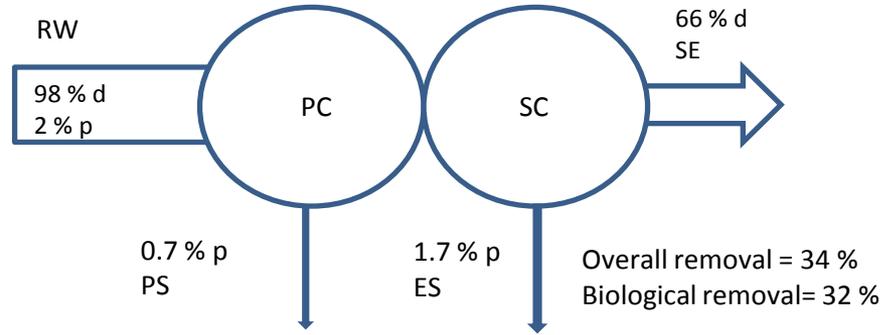


Fig. 2 Overall removal efficiencies of antimicrobials in the WWTP of the city of Zagreb determined a) based on dissolved fraction only (1st approach); b) based on the total concentration in raw wastewater (2nd approach). SPY = sulfapyridine; SMX = sulfamethoxazole; N-Ac-SMX = N-acetylsulfamethoxazole; TMP = trimethoprim; NOR = norfloxacin; CIP = ciprofloxacin; AZI = azithromycin; ERY-H2O = dehydroerythromycin; CLA = clarithromycin.

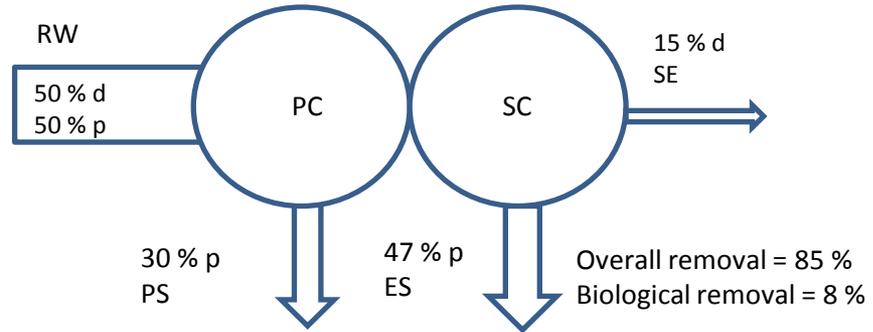
Figure 3

Fig. 3

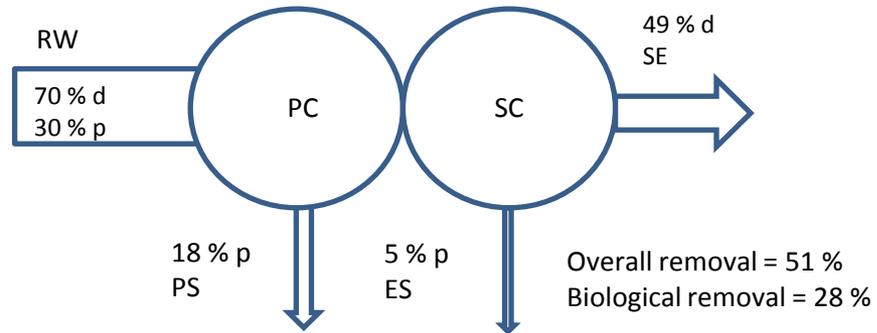
SMX



NOR



AZI



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