

1 **Copper determination by ETAAS in fish tissue cytosols with minimal**  
2 **sample pretreatment**

3

4 Zrinka Dragun\* and Biserka Raspor

5

6 Ruđer Bošković Institute, Division for Marine and Environmental Research, Bijenička

7 cesta 54, P.O.Box 180, 10002 Zagreb, Croatia

8

9 \* corresponding author:

10 Phone: xx385-1-4680216;

11 Fax: xx385-1-4680242;

12 E-mail: [zdragun@irb.hr](mailto:zdragun@irb.hr)

13

14

1 **Abstract**

2 The metal concentrations in the fish tissue cytosols offer an important parameter in  
3 environmental monitoring studies as possible indicators of metal bioavailability in  
4 ambient water. The aim of this study was to examine the possibility of direct Cu  
5 determination in the cytosols of chub gills and intestines by electrothermal atomic  
6 absorption spectrometry (ETAAS) with minimal sample pretreatment and without the  
7 use of modifiers. Tube wall and platform atomizations were tested. When tube wall  
8 atomization was applied, a progressive decline of Cu absorbance was observed with  
9 increasing firing count. On the other hand, stability of Cu absorbance throughout the  
10 graphite tube lifetime (up to 250 firings) was obtained with platform atomization. Signal  
11 stability, together with excellent measurement repeatability ( $\leq 5\%$ ) and acceptable Cu  
12 recovery from spiked samples (77-89%), make ETAAS with platform atomization  
13 applicable to the determination of low Cu concentrations ( $< 27 \mu\text{g L}^{-1}$ ) in environmental  
14 samples with complex organic matrices, such as undigested fish tissue cytosols, even  
15 without the use of chemical modifiers.

16

17

1 **INTRODUCTION**

2

3 During the fall and spring seasons of 2005 and 2006, a metal pollution study was  
4 carried out in the Sava River (Croatia) in the course of the European Union Sixth  
5 Framework Programme project (SARIB project, INCO-CT-2004-509160). Among  
6 others, metal bioavailability in the river water was assessed using the bioindicator  
7 organisms (1,2). Fish are often used as bioindicators of water pollution with metals (3),  
8 and since the European chub (*Squalius cephalus* L.) is the fish species widespread in  
9 European freshwater, it was selected for this study. Two representative tissues, gills and  
10 intestine, were chosen for the metal determinations. As organs in direct contact with the  
11 ambient water and ingested food, they are expected to respond quickly to changes in  
12 metal exposure (4).

13

14 For environmental analyses, determination of total metal concentrations in digested  
15 tissues is the usual procedure. From an analytical point of view, the problem of matrix  
16 complexity is thereby eliminated, but valuable information on subcellular metal  
17 distribution and trophically available metal levels is lost. In our study, metals were  
18 therefore determined in tissue cytosol fractions, which contain heat-sensitive proteins  
19 (such as enzymes) and heat-stable proteins (such as metallothioneins) (5). The presence  
20 of increased metal concentrations in the fraction containing heat-sensitive proteins  
21 (potentially metal sensitive fraction) is expected to be associated with some biological  
22 impairment (6). Metals sequestered by metallothioneins, on the other hand, are  
23 considered detoxified, and thus not available to more sensitive cellular fractions (4). The  
24 cytosolic metal concentrations, as potential indicators of metal bioavailability in

1 ambient water, represent the important and very interesting parameter in environmental  
2 monitoring studies.  
3  
4 To avoid the analytical interferences of a complex organic matrix, tissue cytosols should  
5 be digested prior to Cu determination. However, microwave digestion requires the use  
6 of additional chemicals (mineral acid, hydrogen peroxide) and expensive equipment. It  
7 is, therefore, a costly and time-consuming procedure, as well as a potential source of  
8 sample contamination. It should also be pointed out that the concentrations of metals in  
9 the cytosols of fish tissues are rather low (1, 2), and any additional sample dilution  
10 could result in concentrations below the instrument detection limit. Considering these  
11 facts, we measured the concentrations of several metals (Fe, Zn, Mn, Cu, and Cd)  
12 directly in undigested gill and intestine cytosols to avoid possible sample contamination  
13 and over-dilution, and to make the procedure as simple and low-cost as possible. Iron  
14 and Zn were measured by flame atomic absorption spectrometry (AAS), while Mn, Cu,  
15 and Cd were measured by electrothermal atomic absorption spectrometry (ETAAS) due  
16 to their low concentrations in diluted cytosols (Cu and Mn  $<30 \mu\text{g L}^{-1}$ ; Cd  $<2 \mu\text{g L}^{-1}$ ) (1,  
17 2). Metal determination by ETAAS in fish tissue cytosols without pretreatment has, to  
18 our knowledge, never been reported in the literature. However, the concentrations of Al,  
19 Cd, Co, Cr, Cu, Fe, Mn, Mo, Ni, Pb, and Se were measured in different types of  
20 undigested samples (milk, urine, serum, blood, wheat flour slurry, coconut water) but  
21 usually with the use of various chemicals as dilution or modifier solutions. The oxidant  
22 mixture ( $\text{HNO}_3 + \text{H}_2\text{O}_2$ ) was most frequently applied, but the use of ethanol, tertiary  
23 amines, Triton® X-100 and Pd was also reported (7-15).  
24

1 The main objective of our study was to solve the problems encountered as a  
2 consequence of the complex organic matrix in the direct Cu determination in undigested  
3 fish tissue cytosols. The comparative measurements on tube wall and platform  
4 atomization were performed by ETAAS without use of chemical modifiers. Due to  
5 decreased background, improved long-term stability and minimal chemical  
6 interferences, the use of platforms is recommended for the measurement of metals in  
7 samples with complex matrices (16). However, we first attempted to measure Cu by  
8 applying tube wall atomization, due to lower cost of simple graphite tubes compared to  
9 tubes with integrated platforms. The results obtained with tube wall and platform  
10 atomization were compared, and more favourable approach for Cu measurement in the  
11 undigested fish tissue cytosols was additionally analytically characterized. The detection  
12 and quantification limits, measuring ranges, repeatability, and Cu recovery from spiked  
13 samples are listed, providing new information about the direct metal determination in  
14 cytosolic types of samples by ETAAS (17).

15

## 16 **EXPERIMENTAL**

17

### 18 **Isolation of Gill and Intestine Cytosols**

19 The gills, ranging from 0.5 to 1.0 g, and the intestines, ranging from 1.0 to 2.0 g, were  
20 obtained from specimens of European chub (*Squalius cephalus* L.) caught in the Sava  
21 River (Croatia), cut up, and diluted in five volumes of cooled homogenizing buffer (1,  
22 2). The applied gill homogenization buffer (GHB) was 100 mM Tris-HCl/Base buffer  
23 (Sigma Chemical Co., USA) with a pH equal to 8.1 at 4°C, and supplemented with the  
24 reducing agent, 1 mM dithiotreitol (Sigma Chemical Co., USA). The intestine  
25 homogenization buffer (IHB) additionally contained inhibitors of proteolytic activity,

1 0.5 mM phenylmethanesulphonyl fluoride (PMSF) and 0.006 mM leupeptine (LEU) (18).  
2 High purity water (Milli-Q™ Water System, Millipore, Madrid, Spain) was used for the  
3 preparation of the homogenization buffers. Homogenization was performed in ice-  
4 cooled tubes using a Glas-Col homogenizer (USA), set at 6,000 rpm. The homogenates  
5 were then centrifuged in the Sorval RC28S centrifuge (Kendro, USA) at 50,000×g for 2  
6 h at 4°C. After the centrifugation, supernatant (S50), which is a water-soluble cytosolic  
7 tissue fraction, was separated from the pellet, then five times diluted with Milli-Q water  
8 (1:5), and deep frozen until subsequent metal determination. Diluted cytosols contained  
9 proteins [determined according to Lowry et al. (19)] in a concentration of approximately  
10 2-4 mg mL<sup>-1</sup> (1, 2).

11

## 12 **Standard and Reference Solutions**

13 Regarding the protein content, it was impossible to match the composition of the  
14 standard solutions and the cytosol samples. Thus, Cu calibration standards were  
15 prepared by the appropriate dilutions of a certified Cu standard solution (1,000 mg L<sup>-1</sup>,  
16 ICP standard in 2-3% HNO<sub>3</sub>, Merck, Darmstadt, Germany) with five times diluted GHB  
17 and IHB (as described above). The reference solution SW-HM-47, distributed by Vituki  
18 (Budapest, Hungary) as a part of the Qualco Danube intercomparison study, was used to  
19 examine the possible influence of the calibration standard composition on Cu  
20 determination. This reference solution is an acidified sample of river water with an  
21 assigned Cu concentration of 14.1 µg L<sup>-1</sup>.

22

## 23 **Instrumentation**

24 A Varian SpectrAA 220 (Australia) deuterium-corrected atomic absorption  
25 spectrometer, equipped with a Varian GTA 100 graphite tube atomizer, an autosampler,

1 and multielement hollow cathode lamp (Cu/Zn), was employed in this study. Argon  
2 (99.99% purity) was used as a purge gas. The instrumental parameters are presented in  
3 Table I. The Cu absorbance signal was calculated by using the peak area mode to  
4 minimize potential differences in the atomization rate and signal shape between the  
5 calibration standards and the samples. It was shown experimentally that differences in  
6 matrix composition can alter peak width and height; therefore, absorbance integration  
7 used in the peak area mode produces smaller variations in the signals than peak height  
8 measurement (20). The cytosolic Cu concentrations were measured by tube wall  
9 atomization, using partitioned pyrolytically coated graphite tubes, and by platform  
10 atomization, using the universal tubes with integrated platforms. The heating programs  
11 for tube wall (21) and platform atomization are presented in Table II. There is time  
12 delay between the temperature of the tube and the temperature of the platform, and  
13 therefore the heating program used for graphite tubes with platforms involved higher  
14 drying, ashing, and atomization temperatures compared to the heating program used for  
15 tube wall atomization. A cool-down step is also included at the end of the platform  
16 heating program to ensure that the temperature of the platform is the same for each  
17 sample injection (22).

18

## 19 **RESULTS AND DISCUSSION**

20

### 21 **Effect of Tube Ageing**

22 At first, Cu determination was performed directly in undigested cytosols by a well-  
23 established ETAAS method with tube wall atomization, which is commonly used for Cu  
24 measurement in acidified solutions (21). The method was applied to more complex  
25 sample matrices without any adjustments and without use of modifiers. The obtained

1 results initially seemed to give reliable information on Cu concentrations in fish tissue  
2 cytosols. Excellent repeatability was achieved after Cu determination in duplicates, with  
3 relative standard deviations (RSD) mostly below 3%, while low background absorbance  
4 pointed to complete removal of the complex matrix during sample ashing at 800°C  
5 (Figure 1a). However, with an increasing number of tube firings, the decrease of Cu  
6 absorbance signal became obvious (Figure 1b). It was, therefore, impossible to compare  
7 the results obtained with the increasing number of firings with those obtained with the  
8 new graphite tube.

9  
10 Tube ageing was previously reported as one of the main factors that can affect metal  
11 determination by ETAAS (23). This problem had to be solved to obtain more reliable  
12 Cu measurements, and thus the stability of the Cu absorbance signal was compared for  
13 tube wall and platform atomization. Cu concentrations were recorded periodically in the  
14 selected samples of gill and intestine cytosols during the tube utilization, approximately  
15 up to 200 tube firings. The study was performed separately for gill and intestine  
16 cytosols to establish if their different matrices damage the tube at different rates. For  
17 intestine cytosols the same sample (IN-1, Table III) was used for both tube wall and  
18 platform atomization measurements, so the starting absorbance was the same in both  
19 cases (Figure 2). For gill cytosols, due to the small sample volume, two samples of  
20 different concentrations were used, G-1 for tube wall atomization and G-2 for platform  
21 atomization (Table III). When tube wall atomization was used, a progressive decline of  
22 Cu absorbance signal was observed (Figure 2). After ~200 firings, the signal decreased  
23 approximately 30%, and the rates of tube impairment caused by gill and intestine  
24 cytosol matrices were comparable. When tube wall was replaced with platform  
25 atomization, the Cu absorbance signal recorded after ~200 firings was comparable to



1 the signal obtained with the new tube. It varied only slightly (signal RSD 3-4%)  
2 throughout tube usage (Figure 2). A possible explanation for signal decline is that the  
3 determination of trace metals in biological samples with complex organic matrices gives  
4 rise to the formation of carbonaceous residues inside the tube, which could seriously  
5 alter the effectiveness of the drying and ashing temperatures (24).

6  
7 According to Welz et al. (25), the analytical useful lifetime of graphite tubes is defined  
8 by the number of firings which can be made until the analytical signal drops to about  
9 80% of its initial value, and/or the RSD begins to deteriorate significantly. The severe  
10 effect of tube ageing observed when Cu is measured in chub tissue cytosols using tube  
11 wall atomization significantly reduced the tube lifetime. Based on our analyses, it is  
12 limited to approximately 130-150 firings (Figure 2). However, even within this limited  
13 number of firings, the signal was not stable, but continually decreasing (Figure 2).  
14 Therefore, unlike tube wall atomization, the use of platform atomization can be  
15 recommended for Cu measurements in undigested fish tissue cytosols even without the  
16 use of modifiers, since it provides longer stability of the absorbance signal. Based on  
17 our practical experience, graphite tubes with integrated platforms can be used up to 250  
18 firings without signal deterioration.

19

## 20 **Method Characterization**

21 Since the main purpose of each analytical measurement is to obtain reliable information  
22 about the analyte content in the analyzed sample (26), it is important to characterize the  
23 method performance whenever applied to new types of samples. For the  
24 characterization of the Cu determination in undigested chub tissue cytosols by ETAAS  
25 with platform atomization, the limit of detection, limit of quantification, linearity of

1 measuring range, repeatability, and recovery from spiked samples were assessed mainly  
2 following the International Recommendation R100 (27). Five times diluted cytosols of  
3 six gill and five intestine samples were used, with the Cu concentrations ranging from  
4 7.25-16.66  $\mu\text{g L}^{-1}$  and 18.51-26.45  $\mu\text{g L}^{-1}$ , respectively (Table III). Due to the  
5 differences in matrix composition of gill and intestine cytosols, method characterization  
6 was performed separately for each type of cytosol.

7

### 8 *Calibration*

9 The preliminary results indicated that Cu concentrations in the five times diluted  
10 cytosols of chub tissues mainly fall into the concentration range of 10-20  $\mu\text{g L}^{-1}$ . Since  
11 it is optimal that the calibration range encompasses values up to 150% of the expected  
12 measurements (28), the linearity was tested up to the Cu concentration of 27  $\mu\text{g L}^{-1}$ ,  
13 which was also roughly indicated as the upper calibration limit for Cu determination by  
14 ETAAS (21). According to some recommendations (29), it is sufficient to calculate the  
15 regression coefficient as a measure of linearity; and when it reaches at least 0.999, it can  
16 be assumed that the analytical signal is proportional to the measured analyte  
17 concentration in the defined calibration range. The regression coefficients for Cu  
18 calibration straight lines were equal to 0.997 and 0.998 for gills and intestine  
19 calibrations, respectively (Figure 3 a-b), and thus nearly complied with the requirements  
20 for linearity. However, better fit was achieved when the quadratic calibration curve was  
21 applied, with higher regression coefficients ( $r=0.9999$  in both cases) and lower  
22 intercepts ( $a=0.003-0.004$ ). The slopes of the quadratic calibration curves ( $b=0.013-$   
23  $0.014$ ) were comparable for both calibrations (Figure 3 a-b). The differences in the  
24 composition of calibration standards obviously did not affect the size of Cu absorbance  
25 signal.

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25

### *Method Detection Limits and Measuring Ranges*

For the determination of the detection and quantification limits (LOD and LOQ, respectively), Cu concentrations were measured consecutively 10 times in the blank solutions, and their compositions are provided in Table I. The LOD was calculated on the basis of three standard deviations of the blank (27), and the LOQ on the basis of 10 standard deviations of the blank (30).

For gill calibration (standards prepared with GHB; see section above), the Cu LOD amounted to  $0.27 \mu\text{g L}^{-1}$  (Figure 3a), while for intestine calibration (standards prepared with IHB; see section above) the LOD amounted to  $0.82 \mu\text{g L}^{-1}$  (Figure 3b). The presence of two additional components in the intestine calibration standards, PMSF and leupeptine, evidently increased the noise (standard deviation of the blank). Therefore, the measuring range, which is defined as the range of Cu concentrations from the LOQ value to the upper limit of the calibration range (26), is broader for the gill calibrations with the lower limit set at  $1 \mu\text{g L}^{-1}$ . The lower limit of the measuring range for the intestine calibrations was set at  $3 \mu\text{g L}^{-1}$ .

### *Repeatability*

The repeatability was determined for the gill and intestine calibration standards (see above) with four concentrations selected within the calibration range. It was also established for six gill and five intestine cytosol samples. The indicator of the measurement repeatability was RSD (%) of 10 consecutively measured Cu concentrations in each calibration standard solution and in each cytosol sample (27).

1 The obtained RSDs (%) of Cu measurements in standard solutions were <5%, whether  
2 the calibration standards were prepared with GHB or IHB (Table IV). The repeatability  
3 observed for the Cu determination in undigested cytosols was also excellent, and the  
4 RSDs were mainly  $\leq 5\%$  for both gill and intestine cytosol samples (Table IV).

5

#### 6 *Influence of Calibration Standard and Sample Composition on Cu Recovery*

7 The influence of calibration standard composition on Cu recovery was tested using the  
8 reference acidified river water sample (SW-HM-47), with the assigned Cu concentration  
9 of  $14.1 \mu\text{g L}^{-1}$  and the range of acceptance from  $11.3\text{-}16.9 \mu\text{g L}^{-1}$ . Copper concentration  
10 in SW-HM-47 (obtained when the calibration curve was prepared with the acidified  
11 standards) was  $14.7 \mu\text{g L}^{-1}$ . Similar Cu concentrations were obtained from two  
12 calibration curves, prepared with gill and intestine homogenization buffers (described  
13 above), and amounted to  $14.8$  and  $15.2 \mu\text{g L}^{-1}$ , respectively. Based on comparable  
14 results, whether the calibration curve was created with acidified solutions (104%) or  
15 homogenization buffers (105-108%), it can be concluded that the complex composition  
16 of the calibration standards used in this study did not affect the accuracy of the Cu  
17 measurement.

18

19 Since the certified reference material with matrix corresponding to gill or intestine  
20 cytosol was not commercially available, the reliability of Cu determination in the  
21 cytosols using platform atomization was tested by establishing Cu recovery from the  
22 spiked samples. The volumes of  $300 \mu\text{L}$  of five-time diluted gill or intestine cytosols  
23 were mixed with the appropriate volumes of Cu calibration standards. Although the  
24 cytosols were additionally diluted with calibration standards, the Cu recovery still  
25 should indicate if the complex cytosol matrix affects the measurement reliability.

1 Copper recovery from spiked gill cytosols (Table V) was  $77\pm 9\%$ , while it was higher  
2 for spiked intestine cytosols ( $89\pm 2\%$ ). Unlike the components of the homogenizing  
3 buffers, the organic matrix clearly caused the decrease of Cu recovery. Still, the  
4 recoveries of added Cu from both the gill and the intestine cytosols can be regarded as  
5 acceptable results for environmental monitoring studies, in which the comparability of  
6 the results is more important than the absolute accuracy.

7

## 8 **CONCLUSION**

9

10 The severe deterioration of Cu absorbance signal as a consequence of tube ageing was  
11 observed during the direct Cu measurements in undigested fish gill and intestine  
12 cytosols by ETAAS with tube wall atomization. Due to the longer signal stability (up to  
13 250 tube firings), excellent repeatability, and acceptable recovery from spiked samples,  
14 ETAAS with platform atomization can be recommended for the determination of low  
15 Cu concentrations in fish tissue cytosols, requiring neither previous sample digestion  
16 nor use of additional modifiers.

17

1 **ACKNOWLEDGMENTS**

2 The financial support by the Ministry of Science, Education and Sport of the Republic  
3 of Croatia (project No. 098-0982934-2721), is acknowledged. This study was carried  
4 out as a part of the European FP6 project **SAva RIver Basin: Sustainable Use,**  
5 **Management and Protection of Resources** (INCO-CT-2004-509160). The authors are  
6 especially grateful to M.Sc. Vlatka Filipović Marijić and M.Sc. Marijana Podrug for  
7 isolation of the intestine and gill cytosols, and determination of the protein  
8 concentrations.

9

10

## 1 REFERENCES

- 2 1. Z. Dragun, B. Raspor, and M. Podrug, *Chemosphere* 69, 911 (2007).
- 3 2. V. Filipović Marijić, and B. Raspor, *Toxicol. Lett.* 172S; S159 (2007).
- 4 3. A. Chovanec, R. Hofer, and F. Schiemer, *Fish as bioindicators*. In: B.A. Markert,  
5 A.M. Breure, and H.G. Zechmeister (eds) *Bioindicators and Biomonitors:*  
6 *Principles, Concepts and Applications*. Elsevier Science Ltd., Amsterdam, pp.  
7 639-676 (2003).
- 8 4. L.D. Kraemer, P.G.C. Campbell, and L. Hare, *Environ. Pollut.* 138, 324 (2005).
- 9 5. W.G. Wallace, and S.N. Luoma, *Mar. Ecol.-Prog. Ser.* 257, 125 (2003).
- 10 6. W.G. Wallace, T.M.H. Brouwer, and G.R. Lopez, *Environ. Toxicol. Chem.* 19,  
11 962 (2000).
- 12 7. P. Viñas, N. Campillo, I. López-García, and M. Hernández-Córdoba, *Anal. Chim.*  
13 *Acta* 356, 267 (1997).
- 14 8. N. Campillo, P. Viñas, I. López-García, and M. Hernández-Córdoba, *Anal. Chim.*  
15 *Acta* 390, 207 (1999).
- 16 9. N. Campillo, P. Viñas, I. López-García, and M. Hernández-Córdoba, *Talanta* 48,  
17 905 (1999).
- 18 10. M. González, M. Gallego, and M. Velcárcel, *Talanta* 48, 1051 (1999).
- 19 11. N. Campillo, P. Viñas, I. López-García, and M. Hernández-Córdoba, *Anal.*  
20 *Biochem.* 280, 195 (2000).
- 21 12. A.A. Almeida, and J.L.F.C. Lima, *Atom. Spectrosc.* 22, 324 (2001).
- 22 13. P.R.M. Correia, E. de Oliveira, and P.V. Oliveira, *Talanta* 57, 527 (2002).
- 23 14. P.C. Aleixo, and J.A. Nóbrega, *Food Chem.* 83, 457 (2003).
- 24 15. J. Naozuka, and P.V. Oliveira, *J. Braz. Chem. Soc.* 17, 521 (2006).
- 25 16. L.M. Voth-Beach, *Varian Instruments At Work*, No. AA-54 (1985).

- 1 17. Z. Dragun, and B. Raspor, *J. Anal. Atom. Spectrom.* 20, 1121 (2005).
- 2 18. V. Filipović Marijić, and B. Raspor, *Mar. Pollut. Bull.* 54, 935 (2007).
- 3 19. O.H. Lowry, N.J. Rosebrough, A.L. Farr, and R.J. Randall, *J. Biol. Chem.* 193,  
4 265 (1951).
- 5 20. W. Slavin, D.C. Manning, and G.R. Carnrick, *Atom. Spectrosc.* 2, 137 (1981).
- 6 21. E. Rothery, *Analytical methods for graphite tube atomizers*. Varian Australia Pty  
7 Ltd., Mulgrave, publication No. 85-100848-00 (1988).
- 8 22. L.M. Voth, *Varian Instruments At Work*, No. AA-45 (1985).
- 9 23. M.J. Cal-Prieto, A. Carlosena, J.M. Andrade, P. López-Mahía, S. Muniategui, and  
10 D. Prada, *J. Anal. Atom. Spectrom.* 18, 29 (2003).
- 11 24. R. Sabé, R. Rubio, and L. García-Beltrán, *Anal. Chim. Acta* 419, 121 (2000).
- 12 25. B. Welz, G. Schlemmer, H.M. Ortner, and W. Wegscheider, *Prog. Anal.*  
13 *Spectrosc.* 12, 111 (1989).
- 14 26. P. Konieczka, *Crit. Rev. Anal. Chem.* 37, 173 (2007).
- 15 27. International organization of legal metrology, *International Recommendation*  
16 *R100: Atomic Absorption Spectrometer systems for measuring metal pollutants in*  
17 *water*. OIML Secretariat, USA, TC16/SC2 (2002).
- 18 28. M. Thompson, S.L.R. Ellison, and R. Wood, *Pure Appl. Chem.* 74, 835 (2002).
- 19 29. M. Dobecki, *Zapewnienie jakości analiz chemicznych (in Polish)*. Instytut  
20 *Medycyny Pracy im. Prof. J. Nofera, Łódź* (2004).
- 21 30. J.K. Taylor, *Quality Assurance of Chemical Measurements*. Lewis Publishers Inc.,  
22 *Chelsea* (1987).
- 23



1 **FIGURE CAPTIONS**

2

3 **Fig. 1.** Illustration of Cu peak profile for gill cytosol sample (G1) obtained by tube wall  
4 atomization: **(a)** in the new tube; **(b)** after approximately 150 firings.

5

6 **Fig. 2.** Dependence of individual Cu absorbance signals periodically measured in gill  
7 and intestine cytosols on the number of tube firings for tube wall and platform  
8 atomizations, as indicated in the legend.

9

10 **Fig. 3.** Copper calibration curves obtained by ETAAS with platform atomization; LOD  
11 and LOQ values are denoted on the graph, as well as the regression equations: **(a)** Cu  
12 calibration standards prepared with GHB (Tris/DTT; see text); **(b)** Cu calibration  
13 standards prepared with IHB (Tris/DTT/PMSF/LEU; see text).

**TABLE I**  
**Common Instrumental Parameters for Cu Measurement in Undigested Gill and Intestine Cytosols by ETAAS With Tube Wall and Platform Atomization**

<b>Blank Gills</b>	20 mM Tris HCl/Base, 0.2 mM DTT	<b>Background Correction</b>	Deuterium Lamp
<b>Blank Intestine</b>	20 mM Tris HCl/Base, 0.2 mM DTT, 0.1 mM PMSF, 1.2 µM leupeptine	<b>Copper Standards</b>	9.0 µg L <sup>-1</sup> 18.0 µg L <sup>-1</sup> 27.0 µg L <sup>-1</sup>
<b>Sampling Mode</b>	Automix	<b>Calibration Algorithm</b>	Quadratic
<b>Measurement Mode</b>	Peak Area	<b>Sample Volume</b>	20 µL
<b>Wavelength</b>	324.8 nm	<b>Total Volume</b>	35 µL
<b>Slit Width</b>	1.0 nm	<b><sup>a</sup>Copper Bulk Concentration</b>	18.0 µg L <sup>-1</sup>
<b>Lamp Current</b>	7.0 mA	<b>Gas Type</b>	Argon

<sup>a</sup>Copper bulk concentration refers to the concentration of Cu standard solution which is used for preparation of the other two calibration standards in automix sampling mode.

**TABLE II**  
**Heating Programs for Cu Measurement by ETAAS With Tube Wall and Platform Atomization**

<sup>a</sup> Heating Program for Tube Wall Atomization				<sup>b</sup> Heating Program for Platform Atomization			
Stage	Temperature °C	Time s	Gas Flow L min <sup>-1</sup>	Stage	Temperature °C	Time s	Gas Flow L min <sup>-1</sup>
<b>Hot Inject</b>	60			<b>Hot Inject</b>	100		
<b>Drying</b>	85	5.0	3.0	<b>Drying</b>	100	5.0	3.0
	95	40.0	3.0		210	25.0	3.0
	120	10.0	3.0		300	10.0	3.0
<b>Ashing</b>	800	5.0	3.0	<b>Ashing</b>	1000	5.0	3.0
	800	2.5	3.0		1000	30.0	3.0
	800	0.5	0.0		1000	2.0	0.0
<b>Atomization</b>	2300	1.1	0.0	<b>Atomization</b>	2600	0.8	0.0
	2300	2.0	0.0		2600	2.0	0.0
<b>Clean</b>	2800	2.0	3.0	<b>Clean</b>	2800	2.0	3.0
	2800	5.0	3.0		2800	3.0	3.0
				<b>Cooling</b>	40	21.9	3.0
					40	35.0	3.0

<sup>a</sup>Heating program for tube wall atomization was taken from Varian GTA manual (21).

<sup>b</sup>Heating program for platform atomization was recommended by Varian experts (personal communication).

**TABLE III**  
**Concentrations of Cu in Diluted Gill and Intestine Cytosols Measured by ETAAS With Platform Atomization (<sup>a</sup>n=10)**

Gill Cytosol	Cu (µg L <sup>-1</sup> )	Intestine Cytosol	Cu (µg L <sup>-1</sup> )
G-1	14.03±0.56	IN-1	18.51±0.69
G-2	7.25±0.08	IN-2	21.55±1.18
G-3	8.55±0.11	IN-3	20.65±0.52
G-4	9.60±0.15	IN-4	19.24±0.41
G-5	16.66±0.45	IN-5	26.45±0.20
G-6	14.59±0.40		

<sup>a</sup>n refers to number of replicates

**TABLE IV**  
**Repeatability of Cu Measurements by ETAAS With Platform Atomization in Gill and Intestine Cytosols, as well as in two Types of Calibration Standards (<sup>a</sup>n=10)**

		Gills (% RSD)	Intestine (% RSD)
<sup>b</sup> Standards	4 µg L <sup>-1</sup>	0.8	1.7
	9 µg L <sup>-1</sup>	4.6	3.3
	18 µg L <sup>-1</sup>	4.7	1.1
	24 µg L <sup>-1</sup>	0.5	0.9
<sup>c</sup> Samples	G-1/IN-1	4.0	3.7
	G-2/IN-2	1.1	5.5
	G-3/IN-3	1.3	2.5
	G-4/IN-4	1.6	2.1
	G-5/IN-5	2.7	0.7
	G-6	2.7	

<sup>a</sup>n refers to number of replicates.

<sup>b</sup>Standards for gill calibration prepared with Tris/DTT buffer; standards for intestine calibration prepared with Tris/DTT/PMSF/LEU buffer.

<sup>c</sup>Gill cytosol samples (G), intestine cytosol samples (IN).

**TABLE V**  
**Copper Recovery From Chub Gill and Intestine Cytosols**  
**Spiked With the Standard Solutions**  
**(Volumes of 300  $\mu\text{L}$  of diluted gill or intestine cytosols were**  
**mixed with the appropriate volumes of Cu calibration**  
**standards; the results are referring to the ETAAS**  
**measurements with platform atomization.)**

Sample	Added Cu		<sup>c</sup> Recovery of Added Cu	
	$\mu\text{g L}^{-1}$	$\mu\text{g L}^{-1}$	$\mu\text{g L}^{-1}$	%
<sup>a</sup> G-1	10.91	8.07		<b>74.0</b>
G-2	5.14	4.38		<b>85.2</b>
G-3	8.18	6.80		<b>83.1</b>
G-4	9.60	7.44		<b>77.5</b>
G-5	16.62	13.20		<b>76.5</b>
G-6	9.60	5.81		<b>60.5</b>
	<b>Average Recovery</b>			<b>76.6<math>\pm</math>8.9</b>
<sup>b</sup> IN-1	12.00	10.59		<b>88.2</b>
IN-2	15.00	13.87		<b>92.4</b>
IN-3	15.00	13.29		<b>88.6</b>
IN-4	12.00	10.38		<b>86.5</b>
IN-5	15.00	13.49		<b>89.9</b>
	<b>Average Recovery</b>			<b>89.1<math>\pm</math>2.2</b>

<sup>a</sup> G - gill cytosol.

<sup>b</sup> IN - intestine cytosol.

<sup>c</sup> Average of three consecutive measurements.

Figure 1.

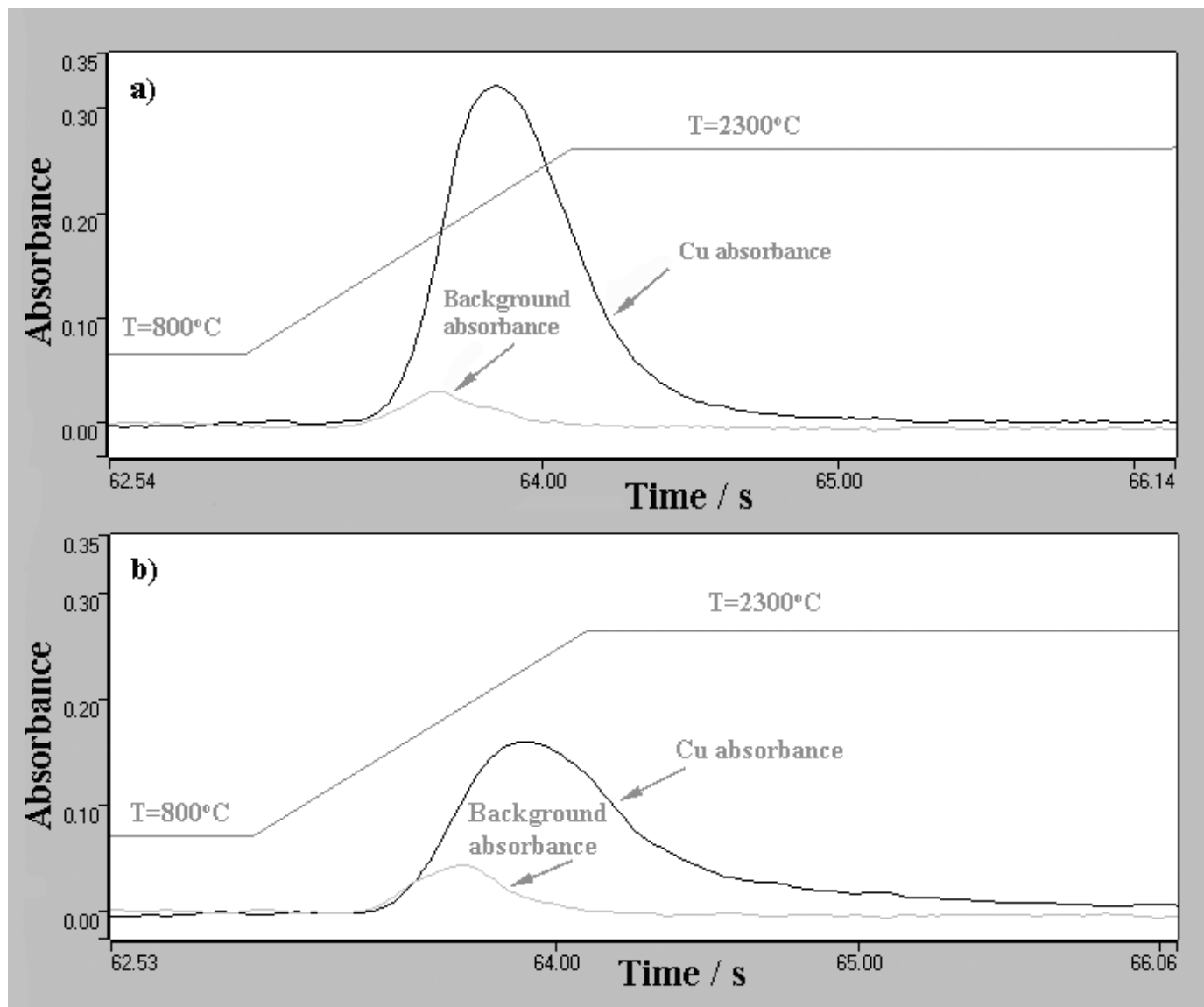


Figure 2.

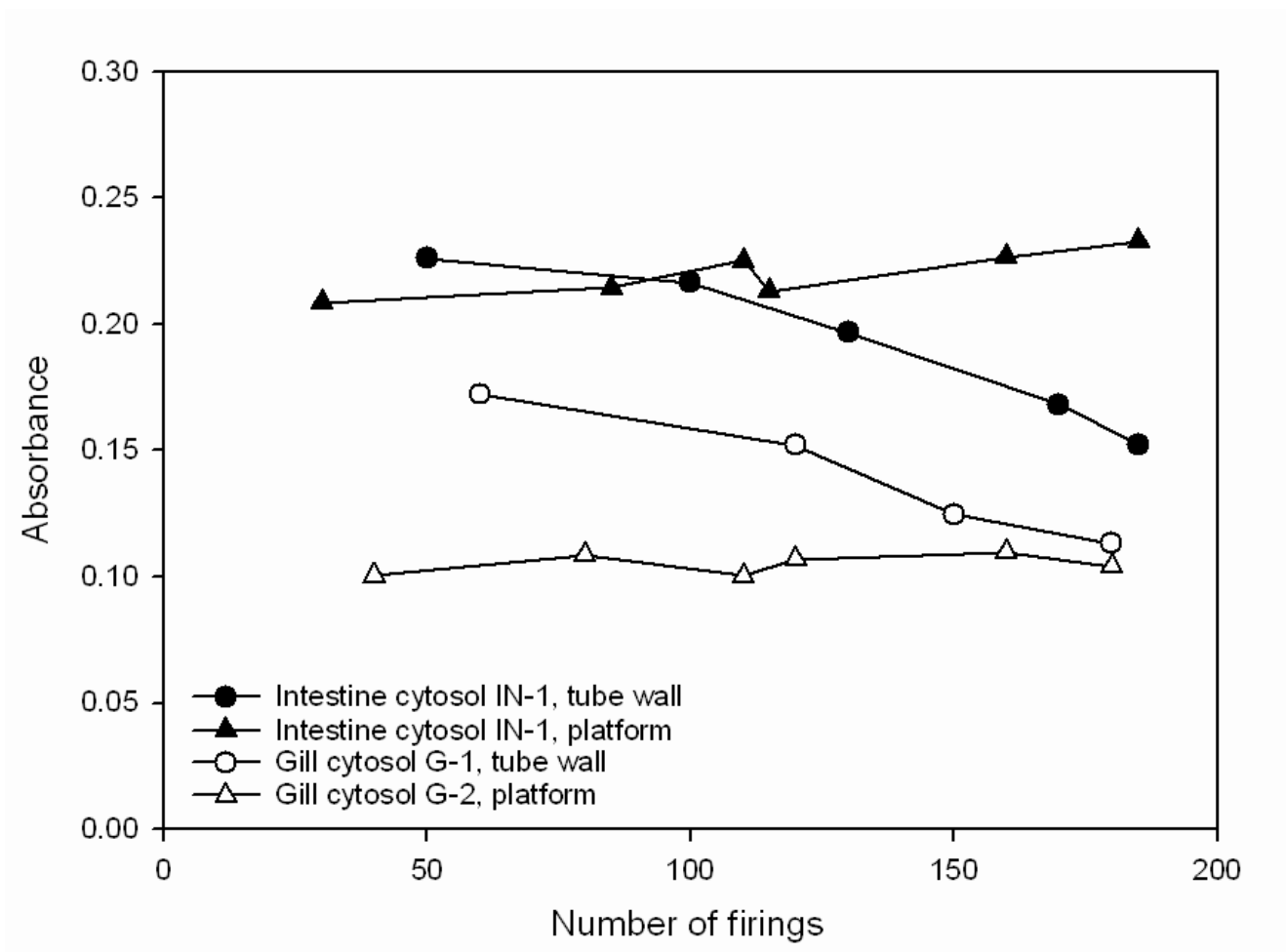


Figure 3.

