

**Optimized extraction, quantification and characterization of seven environmentally relevant metallic nanoparticles in sewage sludge: Occurrence, composition and environmental risk assessment**

**Mavro Lučić<sup>1,3\*</sup>, Radmila Milačić Ščančar<sup>1,2</sup>, Janez Ščančar<sup>1,2</sup>, Željka Fiket<sup>3</sup>, Bor Arah<sup>4</sup>, Janja Vidmar<sup>1,2\*</sup>**

<sup>1</sup>Department of Environmental Sciences, Jožef Stefan Institute, Jamova 39, Ljubljana, Slovenia

<sup>2</sup>Jožef Stefan International Postgraduate School, Jamova 39, 1000 Ljubljana, Slovenia

<sup>3</sup>Division for Marine and Environmental Research, Ruder Bošković Institute, Bijenička cesta 54, Zagreb, Croatia

<sup>4</sup>Center for Electron Microscopy and Microanalysis, Jožef Stefan Institute, Jamova 39, Ljubljana, Slovenia

\*Corresponding authors:

[mlucic@irb.hr](mailto:mlucic@irb.hr)

[janja.vidmar@ijs.si](mailto:janja.vidmar@ijs.si)

**Abstract**

The quantitative analysis of metallic nanoparticles (MNPs) in environmental matrices is crucial for understanding their occurrence, fate, and ecological risks. This study presents an optimized protocol for extracting and quantifying seven of the most commonly used MNPs (Ag, Au, CeO<sub>2</sub>, CuO, PbO, TiO<sub>2</sub> and ZnO) from sewage sludge using single-particle inductively coupled plasma mass

spectrometry (sp-ICP-MS). Several extraction parameters and solutions, including milli-Q water, tetrasodium pyrophosphate (TSPP), and tetramethylammonium hydroxide (TMAH), were systematically evaluated. TSPP at 2.5 mM in a 1:100 sludge-to-reagent ratio was identified as the optimal extraction solution, providing high particle number recovery (>75%) and mass recovery (>84.5%) while minimizing particle transformation. Application to sludge samples from wastewater treatment plants in Croatia resulted in high particle number concentrations of  $1 \times 10^7$  to  $3 \times 10^{10}$  particles/g. The scanning electron microscopy coupled with energy dispersive X-ray spectroscopy (SEM-EDS) providing insights into the origin of the particles and their transformation pathways. The risk assessment concluded that there are negligible to low ecological risks under current agricultural application scenarios, although potential negative impacts should be assessed in long-term projections. The methodology presented in this work addresses critical gaps in environmental monitoring capabilities for engineered NPs in their primary environmental sink.

**Keywords:** Metallic nanoparticle; Single-particle ICP-MS; Sewage sludge; Extraction optimization; Environmental risk assessment

## Environmental Implication

Metal- and metal oxide-based engineered nanoparticles (NPs) are a growing environmental concern due to their increasing use and inevitable release into the environment. We present an optimized analytical approach that enabled a comprehensive analysis of seven metallic NPs, including the first-time quantification of PbO NPs in solid environmentally matrices. The developed analytical method opens up avenues for the routine analysis of metallic NPs, addressing a critical gap in environmental monitoring capabilities and providing essential data for more accurate risk assessments and evidence-based regulatory decisions.

## 1. Introduction

Nowadays, the human population is exposed to an increasing number of chemical pollutants. Metal- and metal oxide-based engineered nanoparticles (NPs) represent an important group of so-called

"*emerging contaminants*" due to their small size (1–100 nm) and high surface reactivity, which enable them to penetrate biological cell walls and membranes (El-Kalliny et al. 2023; Keller et al. 2017). Given their wide array of applications in technology development and manufacturing, metallic NPs (MNPs) emissions into the environment are inevitably on the rise, posing increasing risks to living organisms and human health (Abbas et al. 2020; Besha et al. 2020; Yang et al. 2017). While exposure models are considered essential tools for predicting environmental concentrations of MNPs in regulatory risk assessment processes, these models are inherently prone to generalization. Reliable risk assessments require temporally and spatially resolved analytical data on the concentrations and composition of MNPs in the environment (Keller et al. 2024; Nowack, 2017).

Current research indicates that approximately 95% of MNPs ultimately reach wastewater treatment systems after use (Westerhoff et al. 2011). Throughout the wastewater treatment process, these NPs predominantly partition into sewage sludge rather than remaining in the effluent water. This concentration mechanism creates a critical environmental reservoir for MNPs. The subsequent application of this sludge to agricultural soils—a practice that accounts for approximately 31% of sewage sludge management in EU Member States (Eurostat, 2022) and USA (U.S. Environmental Protection Agency, 2023)—establishes a direct pathway for MNPs to enter food production systems and potentially impact human health. While Council Directive 86/278/EEC established limits for conventional heavy metals in sludge-amended agricultural soils, this framework predates our understanding of NP behavior and does not consider the unique properties of MNPs. Unlike their bulk counterparts, nanoparticulate forms exhibit enhanced bioavailability, greater mobility in soil systems, and potentially higher toxicity due to their increased surface area and reactivity (Keller et al. 2017; Ng et al. 2018; Rippner et al. 2021; Wu et al. 2019), properties that make their environmental fate particularly challenging to predict.

The regulatory and environmental concerns are further complicated by significant methodological barriers that hamper accurate quantification and characterization of NPs in complex environmental matrices. These challenges include: (i) extremely low environmental concentrations (i.e., ng L<sup>-1</sup>) against a complex background of natural particles, dissolved species, and organic matter; (II) transformation processes that alter the physicochemical properties of MNPs during their environmental lifecycle, requiring analytical methods capable of capturing these changes; (III) and lack of standardized protocols for extraction and analysis, resulting in data incomparability across

studies and hindering systematic risk assessment (Abbas et al. 2024; Bundschuh et al. 2018). While electron and atomic microscopy techniques (TEM, SEM, AFM) provide valuable morphological information, their laborious sample preparation requirements limit applicability for routine environmental monitoring. Single-particle inductively coupled plasma mass spectrometry (sp-ICP-MS) has emerged as a promising analytical approach due to its ability to perform sensitive and selective high-throughput measurements with relatively simple sample preparation while simultaneously providing information on number-based particle size distributions and particle mass/number concentrations (Goodman et al. 2025; Laborda et al. 2014; Vidmar et al. 2017).

The validation of analytical methods for NPs in complex matrices using sp-ICP-MS presents unique challenges not encountered with conventional dissolved metal analysis. In contrast to traditional analytes, NPs can undergo multiple transformations during extraction procedures, including dissolution, agglomeration, and even new particle formation (Borowska and Jankowski, 2023; Laycock et al. 2022). These transformations significantly impact the measured particle size distributions, number concentrations, and mass concentrations, often leading to quantitative results that do not accurately reflect the actual NP content in environmental samples. For the quantitative determination of MNPs in environmental samples, several key methodological considerations emerge as critical. No single metric (particle number, mass concentration, or size distribution) provides a complete picture of NPs properties, and thus, truly robust methods must evaluate all relevant metrics, as these parameters have no predictive power for each other. For example, particle aggregation decreases their number concentration and increases the mean particle size, while the mass concentration should remain unchanged. Conversely, dissolution reduces all three measurands, but in different proportions (Laycock et al. 2022). Unlike conventional analytes, where spike recovery of  $100 \pm 10\%$  is typically acceptable, NP recoveries are more challenging to interpret, necessitating a wider acceptable recovery range of  $\pm 25\%$ . This broader range is necessary due to cumulative errors in recovery metrics (Baur et al. 2020; Gray et al. 2013; Laycock et al. 2022).

Several extraction solutions have been employed to extract metallic NPs from complex environmental matrices, including tetrasodium pyrophosphate (TSPP), sodium chloride (NaCl), tetramethylammonium hydroxide (TMAH), various surfactants, and ultrapure water, followed by physical separation techniques such as centrifugation, sedimentation, or filtration (Baur et al. 2020; Moreno-Martín et al. 2022; Schwertfeger et al. 2017; Tou et al. 2021; Vidmar et al. 2022). However,

comprehensive comparisons of extraction methodologies across multiple NP types and extraction solutions have been lacking, with previous studies typically focusing on individual extraction methods or specific NP types.

This study aims to develop and optimize an extraction and analysis method for seven environmentally relevant MNPs (Ag, Au, CeO<sub>2</sub>, CuO, PbO, TiO<sub>2</sub> and ZnO) from sewage sludge. By systematically evaluating extraction efficiency using various solutions (including TSPP and TMAH at different concentrations, and milli-Q water) and optimizing methodological parameters such as settling time and sludge-to-reagent ratios, we aim to establish a standardized protocol for reliable quantification and size characterization of these NPs. This integrated analysis approach overcomes previous knowledge gaps regarding the comparative efficiency of different extraction solutions and provides a more comprehensive assessment of all relevant particle metrics.

The optimized method is then applied to analyze sludge samples from five Croatian wastewater treatment plants (WWTPs), providing valuable data on the occurrence and characteristics of these MNPs in real environmental samples. The quantitative assessment of NPs in sludge samples using the sp-ICP-MS method is complemented by structural and chemical characterization by scanning electron microscopy (SEM) coupled with energy-dispersive X-ray spectroscopy (EDS), which reveals critical insights into the chemical speciation and transformation of MNPs in sewage sludge. Finally, we provide an environmental risk assessment using literature toxicity data for the terrestrial environment, classifying risk levels for each MNP under the current scenario. This comprehensive approach makes it possible to close gaps in current analytical approaches and knowledge of real environmental concentrations of MNPs.

## **2. Material and methods**

### **2.1 Sample collection**

Sewage sludge samples were collected from five municipal WWTPs in Croatia. Sludge from the Zagreb (ZG) WWTP was used for method optimization, after which the optimized protocol was applied to screen for MNPs in samples from WWTPs in Velika Gorica (VG), Čakovec (CAK), Karlovac (KAR), and Sisak (SIS) (Table A1). Following collection, samples were transported to the laboratory, where they were air-dried and sieved through a 2-mm mesh.

## 2.2. Chemicals

Milli-Q water ( $18.2 \text{ M}\Omega\cdot\text{cm}$ ) from Millipore (Bedford, MA, USA) was used to prepare all solutions. Ionic Ag, Au, Ce, Cu, Pb, Ti and Zn standard solutions for ICP-MS ( $1000 \text{ mg L}^{-1}$ ) purchased from Merck (Darmstadt, Germany) were utilized to establish the calibration curves for ICP-MS. Nitric acid (65%) from Kemika (Zagreb, Croatia), hydrochloric acid (37%) from Rotipuran (Karlsruhe, Germany), hydrofluoric acid (47-51%) and boric acid ( $40 \text{ g L}^{-1}$ ) from Fluka (Charlotte, NC, USA) were used for the digestion of sewage sludge samples. Tetrasodium pyrophosphate decahydrate (TSPP, purity 99.8%) and tetramethylammonium hydroxide (TMAH, 25%) were procured from Sigma Aldrich (St. Louis, MO, USA) and used as the extraction reagents. Commercial sodium citrate-stabilized (2 mM) gold nanoparticles with a particle diameter of  $52 \pm 5 \text{ nm}$  (as determined by TEM) and Au mass concentration of  $0.053 \text{ mg mL}^{-1}$  (as determined by ICP-MS), as well as silver nanoparticles with a particle diameter of  $40 \pm 4 \text{ nm}$  (as determined by TEM) and Ag mass concentration of  $0.020 \text{ mg mL}^{-1}$  (as determined by ICP-MS) were obtained from NanoComposix (San Diego, CA, USA). These nanoparticle suspensions were employed to prepare Au and Ag NP standard working solutions and spiked sewage sludge samples. For the spiking experiments, aqueous suspensions of  $\text{CeO}_2$  NPs (30-50 nm) and  $\text{CuO}$  NPs (25-55 nm), each containing 20 wt% of the respective NPs, were purchased from US Research Nanomaterials (Houston, TX, USA), while  $\text{TiO}_2$  nanoparticle standard reference material (SRM 1898) was obtained as a powder from NIST (Gaithersburg, MD, USA). According to the certificate of analysis provided by NIST, SRM 1898 consists of 76% anatase and 24% rutile, with  $\text{TiO}_2$  particles of a primary size  $< 50 \text{ nm}$  (as observed by TEM) and a volume-weighted mean diameter of 71 nm in a water suspension, as determined by laser diffraction spectrometry (LDS). Lead (II) oxide ( $\text{PbO}$ ) and zinc oxide ( $\text{ZnO}$ ) NPs were synthesized as powders by means of a sol-gel method and by employing gelatine as stabilizer. The synthesized particles exhibited diameters of 250 nm for  $\text{PbO}$  NPs and 30 nm for  $\text{ZnO}$  NPs (Figure A1-A3). Details of the synthesis process of  $\text{PbO}$  and  $\text{ZnO}$  NPs can be found in Appendix A. For  $\text{PbO}$  and  $\text{ZnO}$  synthesis, the lead (II) nitrate and zinc nitrate hydrate, as well as gelatine were obtained from Sigma Aldrich (St. Louis, MO, USA).

## 2.3. Physicochemical characterization and multielement analysis of sludge samples

Particle size analysis of sludge samples was performed using a laser-based particle size analyzer (LS 13320, Beckman Coulter Inc.) The particle size distribution (PSD) was calculated on the basis of the

Mie theory of light scattering (optical parameters: refractive index = 1.53; absorption index = 0.1). Prior to analysis, the samples were treated with 2.5 mM TSPP solution to avoid flocculation. Organic matter (OM) content was estimated using the loss on ignition (LOI) method (Salehi et al. 2011). LOI was determined gravimetrically after dry ashing at 360°C over 2 h.

For multielement analysis, the samples were digested in a microwave oven (Multiwave 3000, Anton Paar, Graz, Austria) in a two-step total digestion procedure (I step: 5 mL HNO<sub>3</sub> (65% p.a.) + 1 mL HCl (36% s.p.) + 1 mL HF (48% s.p.), and II step: 6 mL H<sub>3</sub>BO<sub>3</sub> (40 g L<sup>-1</sup>)). The analysis was performed using an inductively coupled plasma triple quadrupole mass spectrometer (ICP-MS/MS, 8900, Agilent Technologies, USA). Analytical quality control was ensured through simultaneous analysis of procedural blanks and certified reference materials (Soil NCS DC 773902, GBW 7410; China National Analysis Centre for Iron and Steel, Beijing, China). Good recoveries (90–100%) were obtained for all measured elements.

#### **2.4. Optimization of nanoparticle extraction from sewage sludge**

To optimize the nanoparticle extraction method for sludge samples, spike suspensions containing Ag, Au, CeO<sub>2</sub>, CuO, PbO, TiO<sub>2</sub> and ZnO NPs were prepared. Most spike solutions of NPs supplied as aqueous suspensions were prepared by bath sonication in milli-Q water for 10 min, with the exception of Ag NPs, which achieved optimal stability in 2 mM trisodium citrate. TiO<sub>2</sub> and ZnO powders were dispersed directly in milli-Q water, while PbO NP powder was dispersed in Moringa Oleifera (MO) extract to obtain 1 mg mL<sup>-1</sup> NP suspensions. The MO extract was prepared by shaking 5 mg of MO powder in 20 mL of milli-Q water for 3 hours and then filtered through a 0.45 µm filter. Suspensions of NPs were then subjected to probe sonication using a Qsonica Sonicator Q125 (L.L.C, Newton, CT, USA), equipped with a 6 mm probe (model CL-18), and operated at 80 % amplitude in pulse mode. Sonication times were optimized for each material: 8 minutes 48 seconds for TiO<sub>2</sub>, 18 minutes for PbO and 12 minutes for ZnO NPs. Since the probe can be a source of TiO<sub>2</sub> NPs, a blank sample (milli-Q water) was sonicated under the same conditions as the TiO<sub>2</sub> NPs suspension. The concentration of TiO<sub>2</sub> NPs detected in the blank was then subtracted from the concentrations measured in the actual samples. During the entire sonication, the vials were placed in an ice bath to prevent the suspension from overheating. Since PbO NPs showed dissolution behavior at high

dilutions, the stability of the spike solution was maintained by adding stock solution to 50 mg MO powder and shaking throughout the extraction process, which ensured excellent stability of the suspension. The optimization of the extractions was performed at least in duplicate, while the final validation was conducted in quintuplicate. Sludge samples (0.20 g each) were weighed into polypropylene tubes and combined with 20 mL of extraction solution. Nanoparticle spike suspensions were added to achieve specific concentrations as follows: 40  $\mu\text{g g}^{-1}$  for Ag, Au, and Ce; 100  $\mu\text{g g}^{-1}$  for Cu; and 1000  $\mu\text{g g}^{-1}$  for Ti, Pb, and Zn. The volumes and concentrations of the added NP suspensions were: Au (151  $\mu\text{L}$  of 0.053  $\text{mg mL}^{-1}$ ), Ag (400  $\mu\text{L}$  of 0.020  $\text{mg mL}^{-1}$ ),  $\text{CeO}_2$  (40  $\mu\text{L}$  of 0.20  $\text{mg mL}^{-1}$ ), CuO (100  $\mu\text{L}$  of 0.20  $\text{mg mL}^{-1}$ ), PbO (200  $\mu\text{L}$  of 1  $\text{mg mL}^{-1}$ ),  $\text{TiO}_2$  (200  $\mu\text{L}$  of 0.10  $\text{mg mL}^{-1}$ ), and ZnO (200  $\mu\text{L}$  of 1  $\text{mg mL}^{-1}$ ). These concentrations were selected to approximately match the levels of respective elements found in sludge from the Zagreb WWTP that was used for method optimization. After the 24 h incubation, the samples were shaken for 24 h at 300 rpm on an orbital shaker (Edmund Bühler GmbH, Germany) and subsequently sonicated for 20 min in a sonication water bath (Iskra PIO, Šentjernej, Slovenia). In this experiment, the probe sonication was also tested for the nanoparticle extraction from sludge samples and resulted in a significantly higher number of recovered NPs (results not shown), which is consistent with previous studies (Bland and Lowry, 2020; Schwertfeger et al. 2017; Tou et al. 2021). However, since the probe can be a source of  $\text{TiO}_2$  NPs, its further use was omitted in the development of the method.

The second step in optimizing the extraction efficiency was to calculate the settling time for NP separation from the liquid phase, which mainly depends on the viscosity of the extractant and the particle density. A particle size threshold of 450 nm was established for nanoparticles spiked to sludge, assuming a particle density of 15  $\text{g cm}^{-3}$ . The particle density of 15  $\text{g cm}^{-3}$  was selected after the testing, as the optimal compromise for simultaneous multielement analysis across a range of particle densities, from  $\text{TiO}_2$  NPs (3.92  $\text{g cm}^{-3}$ ; considering mostly anatase composition) to Au NPs (19.3  $\text{g cm}^{-3}$ ). Based on Stokes' law, the estimated time was calculated using different solution viscosities and then an aliquot was taken from the upper liquid layer (0.1–0.5 cm deep) of the extracted sample for analysis. The effect of different sludge-to-reagent ratios, namely 1:100 (200 mg sludge and 20 mL solution) and 1:1000 (20 mg sludge and 20 mL solution), on the recovery of the particle was also tested.



In the third step of optimization, the different extraction media were examined for their capacity to extract nanoparticles and, at the same time, limit the dissolution and agglomeration of the particles. The solutions tested in these experiments include: 1) milli-Q water, which is the mildest extraction solution and can extract the particles readily mobilized in the sludge; 2) TMAH at various concentrations of 5%, 10%, 15%, 20%, and 25%; and 3) TSPP at various concentrations of 2.5 mM, 10 mM, and 20 mM. Details on the extraction procedures performed when using different extraction solutions can be found in Table A2.

## **2.5. Sp-ICP-MS measurements**

Sp-ICP-MS measurements were performed on a model 7900 ICP-MS instrument from Agilent Technologies (Santa Clara, CA, USA) equipped with a quadrupole mass analyser. The operating parameters (summarized in Table A3) were optimized for plasma robustness and sufficient sensitivity. Silver, Au, Ce, Cu and Pb in aqueous standard solutions were analysed for their most abundant isotopes  $^{107}\text{Ag}$  (51.84%),  $^{197}\text{Au}$  (100%),  $^{140}\text{Ce}$  (88.4%),  $^{63}\text{Cu}$  (69.2%) and  $^{208}\text{Pb}$  (52.4%), while Ti and Zn were measured at  $^{47}\text{Ti}$  (7.3%) and  $^{66}\text{Zn}$  (27.7%) to minimize isobaric interferences from  $^{48}\text{Ca}$  to  $^{48}\text{Ti}$  and  $^{64}\text{Ni}$  to  $^{64}\text{Zn}$ , but also polyatomic interferences from sulphur and nitrogen. Silver, Au, Ce, Pb and Ti were analysed in no gas mode, while Cu and Zn were analysed in helium collision mode (4.3 mL  $\text{min}^{-1}$  He gas flow).

The measurements were run in the time-resolved-analysis (TRA) mode in order to collect the intensity for a single particle as it is vaporized and ionized in the plasma. A short integration dwell time of 100  $\mu\text{s}$  per reading was set. An accurate sample flow rate (at peristaltic pump speed of 0.1 rotations/s) was determined daily by pumping a known weight of milli-Q water at room temperature ( $N=3$ ). The transport efficiency was determined according to the “particle size” method (Pace et al. 2011) from a suspension of 53 nm gold nanoparticles (nanoComposix, San Diego, CA, USA) of known concentration (0.05  $\text{mg mL}^{-1}$ ). The suspension was appropriately diluted with milli-Q water to reach an Au NP concentration of 132  $\text{ng L}^{-1}$ . For calibration, a series of dissolved Au standards (0, 0.1, 0.5, 1, 5 and 10  $\mu\text{g L}^{-1}$ ) were prepared from a stock standard solution of ionic Au (1000  $\text{mg L}^{-1}$ ) in 2% (v/v) HCl/0.5% w/v Thiourea solution (Inorganic ventures, Christiansburg, VA, USA) in milli-Q water. For all samples, the dilution factor was adjusted in such a way that the number of detected particles during 1 min acquisition was  $>100$  to reduce the random error governed by Poisson statistics to 10% and lower

than 3800 to limit relative bias due to multiple particle events to below 3%, assuming a transport efficiency of 0.11 (11%), sample flow rate of 0.33 mL/min, and a dwell time of 100  $\mu$ s.

## **2.6. Data treatment**

The raw signal intensity data from ICP-MS were processed using spCal software (Lockwood et al. 2021). SpCal is an open-source Python-based sp-ICP-MS data processing platform with an interactive graphical user interface, enabling the large data sets to be analysed using transparent algorithms. For our purpose, the Poisson-based data filtering ( $\alpha=0.01$ ) was used to differentiate particle signals from background noise.

To calculate particle size, the particle mass was converted to diameter assuming spherical geometry with specific mass fractions and densities (Table A4). While PbO NPs possess a plate-like structure, the spherical approximation was employed to simplify calculations without compromising the accuracy of recovery determinations. Nanoparticle recovery was evaluated using multiple metrics for comprehensive characterization (see Appendix B for details). The mean particle diameter recovery ( $R_{MD}$ ) was calculated as follows:

$$R_{MD} (\%) = \frac{C_{SS}}{C_S} \times 100 \quad (1)$$

Where  $C_{SS}$  is the mean diameter of particles detected in the spiked sewage sludge samples, and  $C_S$  is the mean diameter of particle in the spike solution prepared in mili-Q water (or in MO extract in the case of PbO). The three additional recovery metrics, including particle number concentration recovery and two mass recovery metrics (particle mass and total mass recoveries), were calculated using the following equation:

$$R_M (\%) = \frac{(C_{SS} - C_C)}{C_S} \times 100 \quad (2)$$

Where  $C_{SS}$  is the concentration of the relevant metric (number of particles, particle mass, and total mass of element) in the spiked sewage sludge samples,  $C_C$  is the concentration of the same metric in the (non-spiked) control sludge sample, and  $C_S$  is the concentration of the same metric in the spike solution. Total mass recovery is based on the mean signal value of the time-resolved analysis data

over the specified range and represents the sum of particulate and ionic mass recovery. The graphs were created in the R platform (R Core Team 2023).

## **2.7. Electron microscope analysis**

The supernatants were obtained from the sludge extracts, which were shaken with milli-Q water for 24 h. One or two drops of the supernatants were then analyzed for structural (morphological) and chemical composition using scanning electron microscopy (SEM, Verios G4 HP) and energy dispersive X-ray spectroscopy (EDXS, Ultim Max SDD 65 mm<sup>2</sup>). The powder of synthesized PbO and ZnO NPs was also analyzed to determine size and morphology.

## **3. Results and discussion**

### **3.1. Physico-chemical characterization of sewage sludge samples**

The physico-chemical characterization of sewage sludge samples from five WWTPs revealed distinctive compositional profiles (Table A5 and A6). The pH values of sludge samples extracted from the Zagreb WWTP exhibited a concentration-dependent response during extraction optimization, with TSPP generating a wide range of pH conditions (pH 6.68-9.55) while all TMAH concentrations produced strongly alkaline environments (pH >13). Considering other WWTPs, the pH values showed slight variation within the mildly acidic to neutral range, from 6.33 (CAK) to 6.88 (SIS). Particle size analysis demonstrated that silt was the predominant fraction across all samples, ranging from 57.94% to 86.34%, with SIS containing the highest proportion and CAK the lowest. Sand content exhibited substantial variation, with CAK having the highest percentage (35.11%) and SIS the lowest (3.58%). Clay content ranged from 4.73% in KAR to 15.21% in VG. When analyzing the particle number, the smallest particles dominate, with over 80% of the total particle count falling within the 40-160 nm range. The primary mode across all samples is 80 nm (Fig. A4). The organic matter (OM) content varied considerably between 470 mg g<sup>-1</sup> (SIS) and 811 mg g<sup>-1</sup> (VG), indicating significant organic enrichment in all sludges. Total metal concentrations revealed notable variations across samples (Table A7), with SIS containing the highest concentrations of Al (25.1 mg g<sup>-1</sup>), Ba (389 µg g<sup>-1</sup>), Mg (9.99 mg g<sup>-1</sup>), Na (2.61 mg g<sup>-1</sup>), Ti (2.10 mg g<sup>-1</sup>), and Ce (28.1 µg g<sup>-1</sup>). ZG exhibited the highest levels of Ca (43.8 mg g<sup>-1</sup>), Ag (3.21 µg g<sup>-1</sup>), Cu (323 µg g<sup>-1</sup>), and Pb (1.09 mg g<sup>-1</sup>), while CAK contained the highest concentrations of Fe (27.4 mg g<sup>-1</sup>), K (8.97 mg g<sup>-1</sup>), P (29.2 mg g<sup>-1</sup>), and Zn (1.17 mg g<sup>-1</sup>).

Comparison with regulatory standards revealed that all metal concentrations were below the Croatian permissible levels for processed sludges used in agriculture (Cu: 600  $\mu\text{g g}^{-1}$ , Pb: 500  $\mu\text{g g}^{-1}$ , Zn: 2,000  $\mu\text{g g}^{-1}$ ), except for Pb in ZG (1,090  $\mu\text{g g}^{-1}$ ), which exceeded the Croatian standard, while remaining within the EU Directive 86/278/EEC upper range (750–1,200  $\mu\text{g g}^{-1}$ ).

### **3.2. Optimization and evaluation of extraction efficiency for nanoparticles in sewage sludge**

By systematically evaluating the extraction efficiency of nanoparticles from sludge samples—taking into account their size, particle number, and mass concentration—several key parameters were optimized, including the sludge-to-reagent ratio, type and concentration of the extraction media, and the settling time needed to separate the extracted nanoparticles from the residual matrix.

First, the optimal sludge-to-reagent ratios (1:100 and 1:1000) were systematically evaluated for NP recovery using TMAH (5% and 25%) as an extraction solution. In these preliminary experiments, NPs with varying densities ( $\text{TiO}_2$ ,  $\text{CeO}_2$ , and Au NPs) were tested. The results showed minimal differences in NP recoveries between the two tested ratios, with the 1:100 ratio providing slightly better extraction efficiency while preventing sample swamping and enhancing spike incorporation into the sludge matrix (Table A8). Based on these findings and subsequent comparative experiments with TSPP, we found that the 1:100 ratio ensures optimal performance for the targeted NPs in method development experiments.

Following extraction and before sp-ICP-MS analysis, nanoparticles must be separated from the residual matrix to avoid clogging the nebulizer with larger sludge particles. Several papers have highlighted the importance of settling particles to maximize the recovery of spiked NPs from solid matrices, as centrifugation can lead to the loss of heavier particles when working with nanoparticles of different densities (Liu et al. 2021; Moreno-Martín et al. 2022; Tou et al. 2021). Based on Stokes' law, the settling time was calculated for particles with a diameter of 450 nm in solutions of varying viscosities (Table A2). For TSPP and milli-Q water, the settling time was set at 2 h and 30 min due to their similar viscosities. For TMAH solutions, the settling times increased with concentration: 3 h and 20 min for 5% TMAH, 4 h and 10 min for 10% TMAH, 5 h and 10 min for 15% TMAH, 6 h and 10 min for 20% TMAH, and 7 h and 20 min for 25% TMAH.

Despite the growing concern about ENPs in the environment, a comprehensive comparison of extraction methodologies across multiple nanoparticle types and extraction solutions has been lacking. Previous studies have typically focused on individual extraction methods or specific metallic NPs, leaving significant knowledge gaps regarding the comparative efficiency of different extraction solutions and the simultaneous assessment of all relevant particle metrics (number recovery, mass recovery, total recovery and diameter changes). The following section of this paper will discuss the recoveries and behavior of seven MNPs across different extraction solutions (milli-Q water, TSPP and TMAH), providing a more integrated analysis than previously available. For partially soluble NPs such as Ag NPs, CuO NPs, and ZnO NPs, extraction methods were evaluated both with and without of 24 h incubation period. These conditions allowed us to assess the impact of temporal factors on nanoparticle stability and recovery efficiency in complex sludge matrix.

*Gold nanoparticles.* For Au NPs, TSPP at different concentrations emerged as the optimal extraction solution, with the 20 mM concentration showing the best performance across multiple recovery parameters (Fig. 1; Table A9). The TSPP maintained the integrity of the Au NPs with minimal change in their size (mean diameter  $100 \pm 1.4\%$ ) while achieving high NP number recovery (for 20 mM TSPP:  $89.5 \pm 8.5\%$ ) and good mass recovery metrics (NP mass recovery:  $93.5 \pm 7.4\%$ ; total mass recovery:  $92.8 \pm 8.2\%$ ). In contrast, TMAH performed poorly, especially at lower concentrations (5%), with a remarkably low NP number recovery ( $1.3 \pm 70.5\%$ ) accompanied by significant size alterations (mean diameter recovery  $137.8 \pm 3.6\%$ ). Increasing TMAH concentration to 25% improved recovery but still resulted in substantial particle size changes. A number of studies have raised concerns about particle agglomeration in the presence of TMAH in biological samples, which could potentially lead to particle loss (Laycock et al. 2022). Since TMAH causes strongly alkaline conditions with pH values above 13.0 (Table A5), it can cause solubilization of various organic compounds in the sludge matrix, including humic substances, proteins and polysaccharides, by deprotonation of organic functional groups (Li et al. 2016). This solubilization process can destabilize the electric double layer surrounding Au NPs, which affects the surface charge of the nanoparticles and the aggregation behavior in the aquatic environment. Diegoli et al. (2008) specifically found that high ionic strength conditions can cause charge screening, creating "bridges" between Au NPs and thereby facilitating their aggregation. The observed concentration-dependent recovery of Au NPs with TMAH (mass recovery; 8.0 % at 5 % TMAH vs. 74.6 % at 25 % TMAH)

should be further investigated beyond the scope of this study. Future research should specifically examine the mechanisms behind this counterintuitive behavior, possibly exploring the interplay between the partial or complete solubilization of organic matter and the changing ionic strength environment at different TMAH concentrations. Milli-Q water demonstrated moderate extraction ability (NP number recovery:  $57.0 \pm 5.9\%$ ; NP mass recovery:  $83.8 \pm 12.6\%$ ); however, TSPP significantly outperformed it in NP number recovery, achieving almost 30% higher efficiency.

*Silver nanoparticles.* Ag NPs exhibited distinct behavior in extraction studies (Fig. 1; Table A10).

Notably, 24 h incubation resulted in significant dissolution and diminished recoveries across all extractants (e.g., 2.5 mM TSPP yielded only  $67.5 \pm 4.3\%$  size recovery,  $35.2 \pm 16.6\%$  number recovery, and merely  $13.4 \pm 8.6\%$  mass recovery). This pronounced degradation is likely due to complex sludge interactions in which humic acids and ammonia form soluble Ag  $(\text{NH}_3)_2^+$  complexes, while chloride ions produce AgCl<sub>2</sub> and AgCl<sub>(aq)</sub> species (Jiang et al. 2022), with these reactions occurring despite pH conditions that would normally reduce dissolution. To circumvent these dissolution challenges, we modified our protocol to eliminate the incubation period and prevented the time-dependent dissolution processes that would otherwise occur. Under these optimized conditions, TSPP at varying concentrations (2.5, 10, and 20 mM) demonstrated excellent performance metrics: number recoveries of  $96.5 \pm 6.8\%$ ,  $100.7 \pm 3.0\%$ , and  $108.2 \pm 3.6\%$ , respectively, with corresponding mass recoveries of  $107.4 \pm 6.7\%$ ,  $109.3 \pm 2.5\%$ , and  $116.4 \pm 3.0\%$ . Moreover, milli-Q water also performs well under these specific conditions, providing high recovery rates ( $102.8 \pm 0.2\%$  size recovery,  $96.3 \pm 1.7\%$  number recovery,  $108.8 \pm 0.3\%$  mass recovery). Interestingly, the sludge matrix with extraction solution provided additional stabilization of Ag NPs compared to spike solutions in 2 mM trisodium citrate, explaining the slightly increased but acceptable mass recoveries. TMAH extractant exhibited clear concentration dependence: lower TMAH concentrations (5-10%) maintained satisfactory number recovery ( $117.3 \pm 0.5\%$  and  $105.8 \pm 1.6\%$ ) despite elevated mass recovery ( $130.5 \pm 3.3\%$  and  $155.0 \pm 3.1\%$ ), while higher TMAH concentrations (15-25%) dramatically reduced recovery efficiency (declining to  $45.6 \pm 25.5\%$  at 15%) due to pronounced Ag NP aggregation. These findings emphasize the critical importance of controlling both sample aging and extraction conditions when analyzing Ag NPs in high-organic environmental matrices, as the time-dependent transformation processes significantly impact nanoparticle recovery and characterization results.

*Cerium nanoparticles.* TSPP performed well with consistent and favorable performance for CeO<sub>2</sub> NPs (Fig. 1; Table A11). At 2.5 mM concentration, NP number recovery reached  $90.6 \pm 9.1\%$ , with moderate mass recovery metrics (NP mass recovery:  $75.7 \pm 4.8\%$ ; total mass recovery:  $100.0 \pm 10.2\%$ ). TSPP extraction resulted in a slight reduction in mean particle diameter ( $94.8 \pm 2.2\%$ ), suggesting minimal transformation of analyzed particles. Higher TSPP concentrations (10 and 20 mM) showed similar effectiveness in number recovery ( $91.4 \pm 6.3\%$  and  $92.4 \pm 5.7\%$  respectively) with comparable size changes, indicating that CeO<sub>2</sub> NPs respond consistently to TSPP extraction regardless of concentration within the tested range. TMAH extraction presented an interesting pattern across its concentration gradient. All TMAH concentrations (5-25%) demonstrated high NP number recovery, ranging from  $105.5 \pm 0.5\%$  at 5% TMAH to  $125.3 \pm 5.5\%$  at 20% TMAH, while mass recovery remained satisfactory ( $78.8 \pm 6.0\%$  to  $90.0 \pm 11.2\%$ ) and increased proportionally with TMAH concentration. Milli-Q water showed moderate effectiveness for CeO<sub>2</sub> NP extraction, with number recovery at  $52.4 \pm 2.2\%$  and mass recovery at  $74.0 \pm 8.0\%$ . However, milli-Q water caused an increase in mean particle diameter ( $113.5 \pm 0.6\%$ ), suggesting potential agglomeration into the larger particles. The substantial size alteration coupled with moderate recovery indicates that water alone is insufficient for optimal CeO<sub>2</sub> NP recovery from sludge due to its limited ability to disperse organic matter and disrupt metal-organic bridges.

*Copper nanoparticles.* 2.5 mM TSPP demonstrated excellent nanoparticle number recovery both with incubation ( $93.3 \pm 8.0\%$ ) and without incubation ( $93.7 \pm 3.9\%$ ), making it the most effective extraction solution for CuO NPs (Fig. 1; Table A12). However, this high recovery was accompanied by a reduction in mean particle diameter ( $85.9 \pm 3.6\%$  with incubation;  $84.5 \pm 0.5\%$  without incubation) and variable mass recovery ( $72.9 \pm 25.0\%$  with incubation;  $82.2 \pm 1.2\%$  without incubation), with immediate extraction yielding more consistent mass recovery. To investigate the potential aggregation of the CuO NPs in the spike solution, which can lead to smaller diameters in the sludge samples (Cervantes-Avilés et al. 2021), we also tested the spike solution prepared in an isolated dissolved sludge matrix. The results remained consistent with  $87.1 \pm 4.7\%$  of mean particle recovery, excluding aggregation as a contributing factor. Higher TSPP concentrations (10 and 20 mM) showed concentration-dependent effects that varied with incubation time. Without incubation, 20 mM TSPP yielded higher performance (NP number:  $91.3 \pm 4.2\%$ ; mass recovery:  $76.1 \pm 0.4\%$ ) compared to the 24 h incubation results (NP number:  $65.9 \pm 5.3\%$ ; mass recovery:  $65.3 \pm 9.7\%$ ), suggesting that

prolonged exposure in sludge media may promote dissolution. In contrast, TMAH performed poorly for NP number recovery, with effectiveness declining sharply as concentration increased (from  $37.3 \pm 0.5\%$  at 5% TMAH to  $23.8 \pm 15.2\%$  at 25% TMAH). Interestingly, TMAH showed poor recovery of particle number but satisfactory recovery of mass (from  $100.3 \pm 0.5\%$  at 5 % TMAH to  $70.3 \pm 12.2\%$  at 25 % TMAH) and higher recovery of mean diameter (from  $102.4 \pm 12.4\%$  at 10 % TMAH to  $111.8 \pm 1.6\%$  at 15 % TMAH), suggesting aggregation of CuO NPs under alkaline conditions. Milli-Q water demonstrated limited extraction capability with incubation (NP number recovery:  $39.8 \pm 10.3\%$ ; mass recovery:  $61.4 \pm 20.7\%$ ) but showed slightly better performance for mass recovery without incubation ( $71.9 \pm 21.1\%$ ).

*Lead nanoparticles.* For PbO NPs, extraction with milli-Q water alone yielded moderate NP number ( $71.9 \pm 16.2\%$ ) and mass ( $73.1 \pm 22.1\%$ ) recoveries (Fig. 1; Table A14), with relatively low total mass recovery ( $44.1 \pm 14.7\%$ ) and a constant particle size ( $103.5 \pm 0.7\%$ ). In comparison, 2.5 mM TSPP exhibited the most balanced performance among all tested solutions. It achieved good recovery of NP mass ( $84.1 \pm 14.3\%$ ) alongside moderate recovery of number ( $76.5 \pm 8.7\%$ ), while maintaining acceptable recovery of particle size ( $105.6 \pm 0.6\%$ ) and total mass ( $80.6 \pm 8.9\%$ ). Higher TSPP concentrations showed diminished performance for nanoparticle-specific metrics; 10 mM TSPP resulted in moderate NP mass recovery ( $52.2 \pm 0.2\%$ ) and lower number recovery ( $49.8 \pm 7.9\%$ ), while 20 mM TSPP yielded poor results for both NP number ( $37.8 \pm 13.9\%$ ) and mass ( $38.2 \pm 13.7\%$ ). Notably, total mass recovery increased substantially with higher TSPP concentrations, reaching  $220.9 \pm 14.5\%$  at 20 mM, suggesting increasing interaction with the sludge matrix at higher pH values of 9.55 (Table A5). TMAH exhibited highly anomalous behavior across all tested concentrations (5%, 10%, and 25%), with recovery values frequently exceeding 100%, particularly for mass recovery metrics (e.g., 5% TMAH: NP mass  $277.1 \pm 13.8\%$ , total mass  $793.9 \pm 5.0\%$ ; 25% TMAH: NP mass  $448.3 \pm 2.7\%$ , total mass  $1858.2 \pm 1.7\%$ ). The extraordinarily high total mass recovery, especially with 25% TMAH, strongly suggests solubilization of background lead from the sludge matrix itself and the formation of soluble hydroxide complexes, where the alkaline conditions not only mobilize the added Pb spikes but also dissolve naturally occurring Pb within the sludge, causing concentrations to increase disproportionately in spiked samples compared to control sludge samples (Zeng et al. 2017; Pranudta et al. 2021).



*Titanium nanoparticles.* Titanium dioxide NPs exhibited specific recovery patterns for the different extractants with different behaviors in mean size, number and mass of recovery metrics (Fig. 1; Table A13). The mean size recovery was consistently above 100 % for most extraction methods, indicating possible agglomeration effects, with TSPP extractions yielding  $107.5 \pm 2.0$  % at 2.5 mM and peaking at  $112.3 \pm 1.2$  % at 10 mM. TMAH demonstrated a concentration-dependent trend in size recovery, starting at  $97.2 \pm 2.0$ % (5% TMAH) and increasing to  $111.4 \pm 9.0$ % at 25% concentration. For particle number recovery, TSPP consistently provided the best results, with optimal recovery at 10 mM ( $99.6 \pm 3.4$ %), while TMAH showed significant variability, performing best at 15% concentration ( $81.1 \pm 7.1$ %) but declining sharply at both 10% ( $61.1 \pm 11.9$ %) and 25% ( $59.8 \pm 32.6$ %). Mass recovery patterns revealed complex dynamics, with milli-Q water producing unusually high recovery ( $118.3 \pm 2.6$ %), TSPP showing a concentration-dependent increase from  $90.4 \pm 11.6$ % to  $108.3 \pm 1.1$ %, and TMAH demonstrating the most variable results with a pronounced minimum at 10% concentration ( $48.3 \pm 0.6$ %). Overall, these results indicate that the extraction efficiency of TiO<sub>2</sub> NPs is highly dependent on both extractant type and concentration, with an inverse relationship observed in certain cases between particle number and size/mass recovery. This phenomenon can be attributed to the strong affinity of TiO<sub>2</sub> NPs for organic matter in complex environmental matrices such as sludge (Wang et al. 2021). Stronger extractants (particularly evident when comparing milli-Q water and TSPP) more effectively disrupt organic-nanoparticle interactions, yielding more accurate number concentration, albeit potentially at the expense of mass recovery efficiency.

*Zinc nanoparticles.* ZnO NPs presented significant analytical challenges, with limited recovery success across tested extraction methods, thus only results of milder solutions are reported (Table A15). Both milli-Q water and 2.5 mM TSPP extractions yielded poor recoveries, with no improvement observed in non-incubated samples (results not shown). These extraction methods resulted in substantial particle size reduction (mean diameter  $64.4 \pm 5.7$ % for milli-Q water and  $63.4 \pm 0.7$ % for 2.5 mM TSPP), indicating immediate dissolution and transformation during the extraction process. Nanoparticle number recovery was markedly deficient for both methods, with milli-Q water achieving only  $34.4 \pm 3.0$ % recovery and 2.5 mM TSPP yielding  $42.0 \pm 25.8$ % recovery with high variability. Even worse were the extremely low mass recovery metrics, with NP mass recoveries of merely  $5.0 \pm 15.0$ % and  $6.9 \pm 28.7$ %, while total mass recoveries reached only  $21.2 \pm 36.8$ % and  $0.4 \pm 17.8$ % for milli-Q water and TSPP, respectively. The dramatic size reduction coupled with poor recovery metrics

strongly suggests high reactivity and/or very poor stability, and substantial dissolution of ZnO NPs in the sludge environment, likely due to complexation with abundant organic ligands in the sludge matrix, particularly extracellular polymeric substances and natural organic matter that can bind dissolved  $\text{Zn}^{2+}$  ions and accelerate nanoparticle transformation (Gomez-Gonzalez et al. 2022). The enhanced dissolution is further facilitated by the increased surface area-to-volume ratio of smaller nanoparticles (approximately 30 nm in diameter), which exhibit greater dissolution compared to larger particles (Mudunkotuwa et al. 2011). Consequently, quantitative results for ZnO NPs should be interpreted with considerable caution, as existing extraction protocols appear inadequate for accurately preserving and recovering these highly reactive NPs from environmental matrices.

#### **Figure 1.**

*Further recommendations.* Based on the comprehensive evaluation of the extraction efficiency of seven NP types (Au, Ag,  $\text{CeO}_2$ , CuO,  $\text{TiO}_2$ , PbO and ZnO), 2.5 mM TSPP was finally validated as the optimal extraction solution (Fig. 2). The optimized extraction protocol for stable nanoparticles (i.e., Au,  $\text{CeO}_2$ , PbO,  $\text{TiO}_2$ ) comprised the treatment of 200 mg of sludge sample with 20 mL of 2.5 mM TSPP solution, corresponding to a sludge-to-reagent ratio of 1:100. The mixture was incubated for 24 h, followed by sonication for 20 minutes to enhance nanoparticle dispersion. Subsequent gravitational settling for 2 h and 30 min allowed separation of the solid matrix and the resulting supernatant was appropriately diluted prior to the analysis of MNPs using sp-ICP-MS. However, for partially soluble nanoparticles (i.e., Ag NPs and CuO NPs), the incubation step was eliminated to preserve nanoparticle integrity. Thus, immediate extraction with 2.5 mM TSPP solution without incubation yielded optimal recovery results, maintaining nanoparticle stability while effectively separating them from the sludge matrix. This procedure consistently demonstrated high performance in particle number (over 75 %) and mass recovery (over 84.5 %), while preserving reasonable particle size consistency throughout (Fig. 2). Based on the consistent high recovery rates across multiple metrics, we recommend 2.5 mM TSPP as a reliable extraction solution for routine analytical procedures involving MNPs in complex environmental samples, particularly when preservation of NP properties is essential for further characterization and quantification.

#### **Figure 2.**

### 3.3. Occurrence and origin of (nano)particles in sewage sludge

The optimized extraction protocol was applied to characterize MNPs in sewage sludge samples collected from five WWTPs in Croatia. This characterization was combined with a geochemical analysis strategy that normalizes observed element concentrations in bulk sludge samples to background levels. The strategy calculates local enrichment factors (EFs; see Appendix A) to assess contamination levels and determine their origin. The EFs revealed distinctive metal accumulation patterns across the five studied WWTPs (Fig. A5). Silver showed notably high enrichment (EFs ~40-250) across all facilities, indicating mainly anthropogenic origin. This was mirrored in NP concentrations (Fig. 3), with Ag NPs being particularly abundant in VG ( $2.24 \times 10^8$  particles/g; mean diameter of 22.3 nm, Fig. 4) and CAK ( $1.48 \times 10^8$  particles/g; mean diameter of 20.2 nm). Strong enrichment was observed for Pb in the Zagreb WWTP (EFs ~70-90), corroborated by exceptionally high PbO NP concentrations measured specifically in ZG sludge ( $2.68 \times 10^{10}$  particles/g; mean diameter of 25.4 nm), far exceeding other sites. Similarly, Zn and Cu showed significant enrichment (EFs ~9.0-31.5 for Zn; 12.3-35.1 for Cu), with the highest EFs in VG. This correlates well with the NP concentration data, where VG exhibited the highest ZnO concentration ( $3.25 \times 10^9$  particles/g; mean diameter of 78.0 nm), while CuO NP concentrations were notably elevated across ZG, VG, and CAK ( $1.22\text{--}1.33 \times 10^{10}$  particles/g; mean diameter of 40.4–42.8 nm). These findings suggest substantial anthropogenic inputs from the respective catchment areas. In contrast, Ce displayed no enrichment (EFs ~1.0), revealing predominantly natural sources, with CeO<sub>2</sub> NPs being most concentrated in SIS and CAK sludge ( $2.3 \times 10^9$  and  $2.59 \times 10^9$  particles/g; mean diameter of 19.5 and 20.1 nm). Titanium showed slight enrichment (EFs >1.5 for ZG and VG), indicating a partially anthropogenic origin in the bulk sludge (Westerhoff et al. 2015), whereas TiO<sub>2</sub> NP concentrations were highest in VG ( $8.56 \times 10^9$  particles/g; mean diameter of 111 nm) and SIS ( $7.58 \times 10^9$  particles/g; mean diameter of 99.5 nm).

#### Figure 3.

The mass concentration of MNPs is consistent with the patterns of particle number concentration and confirms the spatial distribution trends observed across the five Croatian WWTPs (Fig. A6). VG consistently exhibited the highest concentrations for most elements on average (TiO<sub>2</sub>: 70.73  $\mu\text{g g}^{-1}$ , ZnO: 8.40  $\mu\text{g g}^{-1}$ , CuO: 5.41  $\mu\text{g g}^{-1}$ , Ag: 0.019  $\mu\text{g g}^{-1}$ , Au: 0.003  $\mu\text{g g}^{-1}$ ), indicating a significant anthropogenic input from industrial activities and urban runoff in its catchment area. An exception to

this pattern was ZG, which had elevated PbO concentrations (on average  $5.00 \mu\text{g g}^{-1}$ ) compared to other sites ( $0.18\text{-}0.77 \mu\text{g g}^{-1}$ ), and SIS, which recorded the highest CeO<sub>2</sub> concentrations (on average  $0.120 \mu\text{g g}^{-1}$ ), consistent with the natural geological sources indicated by the enrichment factor analysis. When compared to total metal content obtained through acid digestion, extraction efficiency varied across sludge samples (Table A16), with most MNPs representing less than 1% of total metal content for Ag, Ce, Pb, and Zn, while Cu and Ti showed higher percentages (even 4.4% and 10.3%, respectively, at VG). These recovery percentages reflect the limited extractability of MNPs from the sludge matrix, attributed to strong particle-matrix interactions and organic matter complexation (Wang et al. 2021). Additionally, nanoparticle transformation processes during wastewater treatment may further reduce extraction efficiency (Kim et al. 2014).

#### **Figure 4.**

*SEM-EDS analysis.* The SEM-EDS provided crucial insights into the composition and transformation pathways of MNPs within the complex matrix of WWTP sludge. While sp-ICP-MS analysis indicated the presence of indigenous Au NPs in some sludge samples (e.g., VG, CAK, and ZG), SEM-EDS analysis confirmed only the presence of spiked Au NPs that were introduced for extraction optimization. Following their introduction, Au NPs primarily interacted with sludge components possibly through surface associations, particularly with sulphur-containing organic matter and Fe-compounds (Fig. A7). Silver particles in the native samples were predominantly present as AgS, as demonstrated by the near-stoichiometric atomic ratios of Ag:S, along with secondary associations of Ca and P in addition to various minor elements (Fig. A8). This indicates that the transformed Ag NPs were likely at least partially amorphous, forming AgS species whose properties (such as higher solubility) differed from the crystalline Ag<sub>2</sub>S commonly observed after passage through wastewater systems and treatment plants (Kraas et al. 2017). Cerium-containing particles identified in sludge samples presented a distinctive rare earth element (REE) signature (Fig. A9). SEM-EDS analysis detected a remarkable assemblage of multiple REEs (La, Ce, Pr, Nd, Sm, Gd) alongside thorium and phosphorus - a composition strongly indicative of monazite or processed REE-phosphate materials (Dar et al. 2025). These REE-phosphate particles showed minimal transformation during anaerobic digestion, maintaining their fundamental structure due to the exceptional stability of rare earth phosphates under reducing conditions. TiO<sub>2</sub> NPs revealed very

similar resistance to weathering and a relatively simple composition dominated by Ti and O with minor Ca association (Fig. A10), suggesting manufactured TiO<sub>2</sub> nanomaterials rather than naturally occurring titanium minerals (Kiser et al. 2009). The compositional purity and lack of elements typically associated with natural titanium minerals (Fe, Al, Si) provide compelling evidence for an anthropogenic origin (Karkee and Gundlach-Graham, 2024), likely from consumer products such as cosmetics, sunscreens, paints, and food additives. This conclusion aligns with research by Kiser et al. (2009) and Westerhoff et al. (2011), who traced similar TiO<sub>2</sub> signatures in wastewater systems to consumer products. The SEM-EDS analysis of CuO NPs indicated compositional profiles that suggest multiple sources. The CAK sample contained numerous particles with a specific composition dominated by copper-chromium oxides (Fig. A11). This distinctive elemental signature strongly suggests industrial origins, most likely from metal finishing, electroplating, or pigment manufacturing processes (Feng et al. 2023). In contrast, naturally occurring Cu particles in sewage sludge exhibited a fundamentally different profile, with Cu present only as a minor component within the broader organo-mineral matrix (Fig. A12). Pb particles in the studied sludge samples can originate from various sources. While the corrosion of Pb pipes or components within the water distribution system is one potential contributor (Peng et al. 2022), the particles observed are probably characteristic of typical products resulting from anaerobic digestion and phosphate-stabilized waste systems (Fig. A13 and A14). The observed elemental associations, such as Pb with S and Cl (potentially forming PbS, PbSO<sub>4</sub> or chloropyromorphite, Pb<sub>5</sub>(PO<sub>4</sub>)<sub>3</sub>Cl, respectively), align with the chemical transformations known to occur under the reducing conditions of anaerobic digestion and within phosphate-rich stabilization environments (Chen et al. 2021; Zeng et al. 2017). Analysis of Zn particles also revealed their predominant occurrence as zinc sulphide (ZnS), frequently incorporating or associating with various cations including Na, K, and Ca (Fig. A15). This composition represents the typical end-product of Zn transformation during anaerobic digestion, aligning with findings by Kim et al. (2014), who documented the prevalence of zinc sulphide forms in such environments. The transformation from diverse inputs—originating from sources like galvanized materials, personal care products, and tire wear—to ZnS is driven by the activity of sulphate-reducing bacteria and zinc's high affinity for sulphide, a process detailed by Lombi et al. (2012). However, some particles exhibit more complex compositions, incorporating elements like P, Al, and Fe, which suggests the presence of multiple Zn species or varied surface associations (Fig. A16). Interestingly, variations in Zn particle composition

were noted between different treatment plants; some showed simpler, sulphide-dominated forms, while others displayed these more complex elemental mixtures. These differences likely stem from variations in the initial Zn sources and overall wastewater characteristics specific to each catchment area.

### **3.4. Environmental risk assessment of metallic nanoparticles**

To assess preliminary environmental risks of MNPs introduced to soil via sludge amendment, we compiled relevant toxicity data for the terrestrial environment (Table A17). Risk quotients (RQ), calculated as the ratio between predicted environmental concentrations (PEC) and predicted no-effect concentrations (PNEC), were used to evaluate potential ecological impacts. PEC values were derived using measured mass concentrations of NPs in sludge, assuming a standard application rate of 1.6 tons per hectare and soil characteristics (0.2 m mixing depth, 1500 kg m<sup>-3</sup> density), yielding a soil mass of 3000 tons per hectare. PNEC values were derived from obtained toxicological benchmarks using the lowest values found in the literature for soil ecosystems. However, no toxicological data for PbO NPs were available; therefore, following the EU REACH methodology, values from bulk PbO were used with the highest assessment factor (1000) applied to account for the uncertainty (European Chemicals Agency, 2012).

Risk levels were classified as negligible ( $RQ < 0.01$ ), low ( $0.01 \leq RQ < 0.1$ ), moderate ( $0.1 \leq RQ < 1$ ), or high ( $RQ \geq 1$ ) based on calculated RQs. All MNPs exhibited negligible to low ecological risks across all samples (Fig. 5). Gold, Ce, and Ti showed consistently negligible risk profiles with RQ values below 0.01. While Cu and Ag demonstrated slightly higher values, they remained within the low-risk category, as did Pb and Zn. The latter two metals presented the highest RQ values among all analyzed metals, though still within the low-risk range, with maximum values observed in samples ZG and VG. These findings indicate that current application rates of sewage sludge pose minimal ecological risks from MNPs to soil ecosystems.

#### **Figure 5.**

Beyond the values used in the PEC/PNEC calculations, the ratios between the mass concentration of particulate NPs and the total mass concentration (particulate + dissolved) of indigenous metallic NPs in sludge samples obtained by sp-ICP-MS provide additional context for the interpretation of the

ecological risk assessments. The data reveal considerable differences in the ratios, indicating their specific tendency to partition between particulate and ionic forms. TiO<sub>2</sub> and CeO<sub>2</sub> NPs showed the highest ratios (2.48 and 0.68 respectively), suggesting greater stability and lower tendency to transform into dissolved species. PbO NPs (0.42), Au NPs (0.34) and CuO NPs (0.23) demonstrated intermediate ratios, while ZnO NPs and Ag NPs exhibited the lowest ratios (0.09 and 0.04, respectively). These findings align with previous spiking experiments that documented the dissolution behavior of Ag and ZnO NPs during the incubation stage. Although our assessment indicates that current sewage sludge application rates pose minimal ecological risks from MNPs, long-term projections or higher application rates warrant careful consideration (Wang et al. 2024). Continued applications could potentially elevate risk levels for resistant MNPs, while others like Ag and ZnO NPs may exert environmental impacts through the higher solubility and various leaching scenarios in the soil environment (Abbas et al. 2020; Kraas et al. 2017; Pradas del Real et al. 2016). Further investigation is essential, given the severe limitations in available nano-specific toxicity data (notably for PbO NPs) and the need for comprehensive long-term monitoring to validate these projections and ensure continued environmental safety.

#### **4. Conclusions**

This study presents a comprehensive analytical approach for the extraction, quantification and characterization of seven environmentally relevant MNPs (Ag, Au, CeO<sub>2</sub>, CuO, PbO, TiO<sub>2</sub> and ZnO) in sewage sludge. Through a systematic evaluation of extraction methodologies, we identified 2.5 mM TSPP and a sludge-to-reagent ratio of 1:100 as the optimal extraction solution. This method provides consistently high performance in particle number recovery (>75%) and mass recovery (>84.5%) while maintaining particle size integrity for most NPs, as identified by sp-ICP-MS. This approach proved to be particularly valuable in investigating the occurrence and concentrations of MNPs in complex environmental matrices, as shown by the analysis of sludge samples from Croatian wastewater treatment plants.

Complementary SEM-EDS analysis revealed insights into nanoparticle composition and transformation pathways, detecting metal-specific associations with matrix components. Some metals underwent sulfidation processes—Ag, Pb and Zn formed AgS, PbS and ZnS, respectively—while others remained largely untransformed, such as Ce in REE-phosphate minerals and Ti in TiO<sub>2</sub>.

The most MNPs currently present negligible to low ecological risks under present agricultural application scenarios. The differential tendencies of metallic NPs to remain in particulate form (TiO<sub>2</sub>>CeO<sub>2</sub>>PbO>Au>CuO>ZnO>Ag) provided additional context for interpreting ecological risks, suggesting that Ag and ZnO NPs may pose environmental risks primarily through their dissolution tendency in high-organic-rich environments. This integrated approach, combining multiple analytical techniques and evaluating complementary particle metrics, represents a significant advancement in environmental monitoring capabilities for MNPs as emerging contaminants.

## **CRedit authorship contribution statement**

**Mavro Lučić:** Writing – original draft, Validation, Methodology, Investigation, Funding acquisition, Data curation, Conceptualization. **Radmila Milačić Ščančar:** Writing – review & editing, **Janez Ščančar:** Writing – review & editing, Funding acquisition. **Željka Fiket:** Writing – review & editing. **Bor Arah:** Data curation, Formal analysis. **Janja Vidmar:** Writing – review & editing, Conceptualization.

## **Declaration of Competing Interest**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

## **Acknowledgements**

This research was supported by the Slovenian Research and Innovation Agency – ARIS Program P1-0143. Mavro Lučić's research fellowship at the Jožef Stefan Institute was funded by the European Union – NextGenerationEU program. We gratefully acknowledge the staff of the wastewater treatment plants for their assistance in providing sludge samples. We extend special thanks to Dr. Marin Ganjto from the Zagreb WWTP for providing substantial sample quantities that enabled our method development, and to Dr. Hana Fajković for supplying sludge samples from the Velika Gorica WWTP.

## **Data Availability**

The sp-ICP-MS data of the indigenous MNPs in the sewage sludge samples will be made publicly available on Zenodo at [DOI], licensed under the Creative Commons Attribution 4.0 International (CC BY 4.0) license.

## **References**



682 Abbas, Q., Yousaf, B., Amina, Ali, M.U., Munir, M.A.M., El-Naggar, A., Rinklebe, J., Naushad, M.,  
 683 2020. Transformation pathways and fate of engineered nanoparticles (ENPs) in distinct interactive  
 684 environmental compartments: a review. *Environ. Int.* 138, 105646.  
 685 <https://doi.org/10.1016/j.envint.2020.105646>

686 Abbas, Q., Zia-ur-Rehman, M., Ullah, H., Munir, M.A.M., Ali, M.U., Ali, M., Pikon, K., Yousaf, B., 2024.  
 687 Recent advances in the detection and quantification of manufactured nanoparticles (MNPs) in  
 688 complex environmental and biological matrices. *J. Clean. Prod.* 471, 143454.  
 689 <https://doi.org/10.1016/j.jclepro.2024.143454>

690 Baur, S., Reemtsma, T., Stärk, H.J., Wagner, S., 2020, Surfactant assisted extraction of incidental  
 691 nanoparticles from road runoff sediment and their characterization by single particle-ICP-MS.  
 692 *Chemosphere.* 246, 125765. <https://doi.org/10.1016/j.chemosphere.2019.125765>

693 Besha, A.T., Liu, Y., Bekele, D.N., Dong, Z., Naidu, R., Gebremariam, G.N., 2020. Sustainability and  
 694 environmental ethics for the application of engineered nanoparticles. *Environ. Sci. Policy.* 103, 85–  
 695 98. <https://doi.org/10.1016/j.envsci.2019.10.013>

696 Bland, G.D., Lowry, G.V., 2020. Multistep Method to Extract Moderately Soluble Copper Oxide  
 697 Nanoparticles from Soil for Quantification and Characterization. *Anal. Chem.* 92, 9620–9628.  
 698 <https://doi.org/10.1021/acs.analchem.0c00824>

699 Borowska, M., Jankowski, K., 2023, Basic and advanced spectrometric methods for complete  
 700 nanoparticles characterization in bio/eco systems: current status and future prospects. *Anal. Bioanal.*  
 701 *Chem.* 415, 4023–4038. <https://doi.org/10.1007/s00216-023-04641-7>

702 Bundschuh, M., Filser, J., Lüderwald, S., McKee, M.S., Metreveli, G., Schaumann, G.E., Schulz, R.,  
 703 Wagner, S., 2018. Nanoparticles in the environment: where do we come from, where do we go to?  
 704 *Environ. Sci. Eur.* 30, 1–17. <https://doi.org/10.1186/s12302-018-0132-6>

705 Cervantes-Avilés, P., Huang, X., Keller, A.A., 2021. Dissolution and Aggregation of Metal Oxide  
 706 Nanoparticles in Root Exudates and Soil Leachate: Implications for Nanoagrochemical Application.  
 707 *Environ. Sci. Technol.* 55, 20, 13443–13451. <https://doi.org/10.1021/acs.est.1c00767>

708 Chen, Z., Xing, R., Yang, Y., Zhao, Z., Liao, H., Zhou, S., 2021. Enhanced in situ Pb (II) passivation  
 709 by biotransformation into chloropyromorphite during sludge composting. J. Hazard. Mater. 408,  
 710 124973. <https://doi.org/10.1016/j.jhazmat.2020.124973>

711 Council of the European Union. COUNCIL DIRECTIVE of 12 June 1986 on the protection of the  
 712 environment, and in particular of the soil, when sewage sludge is used in agriculture, Official Journal  
 713 of the European Communities § (1986)

714 Dar, S.A., Balaram, V., Roy, P., Mir, A.R., Javed, M., Teja, M.S., 2025. Phosphorite deposits: A  
 715 promising unconventional resource for rare earth elements (REEs). Geosci. Front. 102044.  
 716 <https://doi.org/10.1016/j.gsf.2025.102044>

717 Diegoli, S., Manciualea, A.L., Begum, S., Jones, I.P., Lead, J.R., Preece, J.A., 2008. Interaction  
 718 between manufactured gold nanoparticles and naturally occurring organic macromolecules. Sci. Total.  
 719 Environ. 402, 1, 51–61. <https://doi.org/10.1016/j.scitotenv.2008.04.023>

720 El-Kalliny, A.S., Abdel-Wahed, M.S., El-Zahhar, A.A., Hamza, I.A., Gad-Allah, T.A., 2023.  
 721 Nanomaterials: a review of emerging contaminants with potential health or environmental impact.  
 722 Discov. Nano. 18, 68. <https://doi.org/10.1186/s11671-023-03787-8>

723 European Chemicals Agency, 2012. Guidance on Information Requirements and Chemical Safety  
 724 Assessment. [https://echa.europa.eu/guidance-documents/guidance-on-information-requirements-](https://echa.europa.eu/guidance-documents/guidance-on-information-requirements-andchemical-safety-assessment)  
 725 [andchemical-safety-assessment](https://echa.europa.eu/guidance-documents/guidance-on-information-requirements-andchemical-safety-assessment) (accessed 2025-05-13).

726 Eurostat, 2022. Sewage sludge production and disposal [*env\_ww\_spd*]. [Online].

727 Keller, A.A., Zheng, Y., Praetorius, A., Quik, J.T.K., Nowack, B., 2024. Predicting environmental  
 728 concentrations of nanomaterials for exposure assessment - a review. NanoImpact. 33, 100496.  
 729 <https://doi.org/10.1016/j.impact.2024.100496>

730 Feng, J., Burke, I.T., Chen, X., Stewart, D.I., 2023. Assessing metal contamination and speciation in  
 731 sewage sludge: implications for soil application and environmental risk. Rev. Environ. Sci. Biotechnol.  
 732 22, 1037–1058. <https://doi.org/10.1007/s11157-023-09675-y>

733 Filipović, F., Grčić, I., Bermanec, V., Kniewald, G., 2013. Monitoring of total metal concentration in  
 734 sludge samples: Case study for the mechanical–biological wastewater treatment plant in Velika  
 735 Gorica, Croatia. *Sci. Total. Environ.* 447, 17-24. <https://doi.org/10.1016/j.scitotenv.2012.12.078>  
 736 Goodman, A.J., Benner, B.F., Montañó, M.D., 2025. Out of the lab and into the environment: the  
 737 evolution of single particle ICP-MS over the past decade. *Environ Sci: Nano.* 3, 12, 1789-1800.  
 738 <https://doi.org/10.1039/D4EN00804A>  
 739 Gomez-Gonzalez, M.A., Rehkämper, M., Han, Z., Ryan, M.P., Laycock, A., Porter, A.E., 2022. ZnO  
 740 Nanomaterials and Ionic Zn Partition within Wastewater Sludge Investigated by Isotopic Labeling.  
 741 *Glob. Chall.* 6, 2100091. <https://doi.org/10.1002/gch2.202100091>  
 742 Gray, E.P., Coleman, J.G., Bednar, A.J., Kennedy, A.J., Ranville J.F., Higgins, C.P., 2013, Extraction  
 743 and analysis of silver and gold nanoparticles from biological tissues using single particle inductively  
 744 coupled plasma mass spectrometry. *Environ. Sci. Technol.* 47, 14315–14323.  
 745 <https://doi.org/10.1021/es403558c>  
 746 Jiang, J., Wang, X., Zhang, Y., Zhang, J., Gu, X., He, S., Duan, S., Ma, J., Wang, L., Luo, P., 2022.  
 747 The Aggregation and Dissolution of Citrate–Coated AgNPs in High Ammonia Nitrogen Wastewater  
 748 and Sludge from UASB–Anammox Reactor. *Int. J. Environ. Res. Public Health.* 19, 15, 9502.  
 749 <https://doi.org/10.3390/ijerph19159502>  
 750 Karkee, H., Gundlach-Graham, A., 2024, Two-stage hierarchical clustering for analysis and  
 751 classification of mineral sunscreen and naturally occurring nanoparticles in river water using single-  
 752 particle ICP-TOFMS. *Environ. Sci.: Nano.* 11, 4162–4173. <https://doi.org/10.1039/D4EN00288A>  
 753 Keller, A.A., Adeleye, A.S., Conway, J.R., Garner, K.L., Zhao, L., Cherr, G.N., Hong, J., Gardea-  
 754 Torresdey, J.L., Godwin, H.A. and Hanna, S. et al., 2017. Comparative environmental fate and toxicity  
 755 of copper nanomaterials. *NanoImpact.* 7, 28–40. <https://doi.org/10.1016/j.impact.2017.05.003>  
 756 Kim, B., Levard, C., Murayama, M., Brown, G.E., Jr. Hochella, M.F.Jr., 2014. Integrated Approaches  
 757 of X-Ray Absorption Spectroscopic and Electron Microscopic Techniques on Zinc Speciation and  
 758 Characterization in a Final Sewage Sludge Product. *J. Environ. Qual.* 43, 908-916.  
 759 <https://doi.org/10.2134/jeq2013.10.0418>

760 Kiser, M.A., Westerhoff, P., Benn, T., Wang, Y., Perez-Rivera, J., Hristovski, K., 2009, Titanium  
 761 nanomaterial removal and release from wastewater treatment plants. *Environ. Sci. Technol.* 43, 17,  
 762 6757–6763. <https://doi.org/10.1021/es901102n>

763 Kraas, M., Schlich, K., Knopf, B., Wege, F., Kägi, R., Terytze, K. and Hund-Rinke, K., 2017, Long-  
 764 term effects of sulfidized silver nanoparticles in sewage sludge on soil microflora. *Environ. Toxicol.*  
 765 *Chem.* 36, 3305-3313. <https://doi.org/10.1002/etc.3904>

766 Laborda, F., Bolea, E., Jiménez-Lamana, J., 2014, Single Particle Inductively Coupled Plasma Mass  
 767 Spectrometry: A Powerful Tool for Nanoanalysis. *Anal. Chem.* 86, 5, 2270–2278.  
 768 <https://doi.org/10.1021/ac402980q>

769 Laycock, A., Clark, N.J., Clough, R., Smith, R., Handy, R.D., 2022, Determination of metallic  
 770 nanoparticles in biological samples by single particle ICP-MS: A systematic review from sample  
 771 collection to analysis. *Environ. Sci.: Nano.* 2, 9, 420–453. <https://doi.org/10.1039/D1EN00680K>

772 Li, D., Zhou, Y., Tan, Y., Pathak, S., bin Abdul Majid, M., Ng, W.J., 2016. Alkali-solubilized organic  
 773 matter from sludge and its degradability in the anaerobic process. *Bioresour. Technol.* 200, 579–586.  
 774 <https://doi.org/10.1016/j.biortech.2015.10.083>

775 Lockwood, T.E., Gonzalez De Vega, R., Clases, D., 2021. An interactive Python-based data  
 776 processing platform for single particle and single cell ICP-MS. *J. Anal. At. Spectrom.* 36, 2536–2544.  
 777 <https://doi.org/10.1039/D1JA00297J>

778 Lombi, E., Donner, E., Tavakkoli, E., Turney, T.W., Naidu, R., Miller, B.W., Scheckel, K.G., 2012. Fate  
 779 of Zinc Oxide Nanoparticles during Anaerobic Digestion of Wastewater and Post-Treatment  
 780 Processing of Sewage Sludge. *Environ. Sci. Technol.* 46, 16, 9089-9096.  
 781 <https://doi.org/10.1021/es301487s>

782 Moreno-Martín, G., Gómez-Gómez, B., León-González, M.E., Madrid, Y., 2022. Characterization of  
 783 AgNPs and AuNPs in sewage sludge by single particle inductively coupled plasma-mass  
 784 spectrometry. *Talanta.* 238, 123033. <https://doi.org/10.1016/j.talanta.2021.123033>

785 Mudunkotuwa, I.A., Rupasinghe, T., Wu, C.-M., Grassian, V.H., 2011. Dissolution of ZnO  
 786 Nanoparticles at Circumneutral pH: A Study of Size Effects in the Presence and Absence of Citric  
 787 Acid. *Langmuir*. 28, 1, 396–403. <https://doi.org/10.1021/la203542x>

788 Ng, D.Q., Chu, Y., Tan, S.W., Wang, S.L., Lin, Y.P., Chu, C.H. Soo, Y.-L., Song, Y.-F., Chen, P.-J.,  
 789 2018. In vivo evidence of intestinal lead dissolution from lead dioxide (PbO<sub>2</sub>) nanoparticles and  
 790 resulting bioaccumulation and toxicity in medaka fish. *Environ. Sci.: Nano*. 6, 2, 580–591.  
 791 <https://doi.org/10.1039/C8EN00893K>

792 Nowack, B., 2017. Evaluation of environmental exposure models for engineered nanomaterials in a  
 793 regulatory context. *NanoImpact*. 8, 38-47. <https://doi.org/10.1016/j.impact.2017.06.005>

794 Pace, H.E., Rogers, N.J., Jarolimek, C., Coleman, V.A., Higgins, C.P., Ranville J.F., 2011.  
 795 Determining transport efficiency for the purpose of counting and sizing nanoparticles via single  
 796 particle inductively coupled plasma mass spectrometry. *Anal. Chem.* 83, 9361–9369.  
 797 <https://doi.org/10.1021/ac201952t>

798 Peng, Y.-C., Lu, Y.-F., Lin, Y.-P., 2022. Release of Particulate Lead from Four Lead Corrosion  
 799 Products in Drinking Water: A Laboratory Study Coupled with Microscopic Observations and  
 800 Computational Fluid Dynamics. *Environ. Sci. Technol.* 56, 17, 12218-12227.  
 801 <https://doi.org/10.1021/acs.est.2c02461>

802 Pradas del Real, A.E., Castillo-Michel, H., Kaegi, R., Sinnet, B., Magnin, V., Findling, N., Villanova, J.,  
 803 Carrière, M., Santaella, C., Fernández-Martínez, A., Levard, C., Sarret, G., 2016. Fate of Ag-NPs in  
 804 Sewage Sludge after Application on Agricultural Soils. *Environ. Sci. Technol.* 50, 4, 1759–1768.  
 805 <https://doi.org/10.1021/acs.est.5b04550>

806 Pranudta, A., Chanthapon, N., Kidkhunthodc. P., El-Moselhy, M.M., Nguyen, T.T., Padungthon, S.,  
 807 2021. Selective removal of Pb from lead-acid battery wastewater using hybrid gel cation exchanger  
 808 loaded with hydrated iron oxide nanoparticles: Fabrication, characterization, and pilot-scale validation.  
 809 *J. Environ. Chem. Eng.* 9, 5, 106282. <https://doi.org/10.1016/j.jece.2021.106282>

810 R core team, 2023. R: a language and environment for statistical computing. R Foundation for  
 811 statistical computing, Vienna, Austria. <http://www.R-project.org>

812 Rippner, D.A., Margenot, A.J., Fakra, S.C., Aguilera, L.A., Sohng, J., Dynarski K.A., Waterhouse, H.,  
813 McElroy, N., Wade, J., Hind, S.R., Green, P.G., Peak, D., McElrone, A.J., Chen, N., Feng, R., Scow,  
814 K.M., Parikh, S.J., 2021. Microbial response to copper oxide nanoparticles in soils is controlled by land  
815 use rather than copper fate. *Environ. Sci.: Nano.* 8, 3560-3576. <https://doi.org/10.1039/D1EN00656H>

816 Salehi, M., Beni, O.H., Harchegani, H.B., Borujeni, I.E., Motaghian, H., 2011. Refining soil organic  
817 matter determination by loss-on-ignition. *Pedosphere.* 21, 473–482. [https://doi.org/10.1016/S1002-](https://doi.org/10.1016/S1002-0160(11)60149-5)  
818 [0160\(11\)60149-5](https://doi.org/10.1016/S1002-0160(11)60149-5)

819 Schwertfeger, D.M., Velicogna, J.R., Jesmer, A.H., Saatcioglu, S., McShane, H., Scroggins, R.P.,  
820 Princz, J.I., 2017. Extracting metallic nanoparticles from soils for quantitative analysis: method  
821 development using engineered silver nanoparticles and SP-ICP-MS. *Anal. Chem.* 89, 4, 2505–2513.  
822 <https://doi.org/10.1021/acs.analchem.6b04668>

823 Tou, F., Niu, Z., Fu, J., Wu, J., Liu, M., Yang, Y., 2021. Simple Method for the Extraction and  
824 Determination of Ti-, Zn-, Ag-, and Au-Containing Nanoparticles in Sediments Using Single-Particle  
825 Inductively Coupled Plasma Mass Spectrometry. *Environ. Sci. Technol.* 55, 15, 10354–10364.  
826 <https://doi.org/10.1021/acs.est.1c00983>

827 Tou, F., Yang, Y., Feng, J., Niu, Z., Pan, H., Qin, Y., Guo, X., Meng, X., Liu, M., Hochella, M. F., 2017.  
828 Environmental Risk Implications of Metals in Sludges from Waste Water Treatment Plants: The  
829 Discovery of Vast Stores of Metal Containing Nanoparticles. *Environ. Sci. Technol.* 51, 9, 4831–4840.  
830 <https://doi.org/10.1021/acs.est.1c00983>

831 U. S. Environmental Protection Agency. Basic Information about Biosolids. [https://www.epa.gov/](https://www.epa.gov/biosolids/basic-information-about-biosolids#uses)  
832 [biosolids/basic-information-about-biosolids#uses](https://www.epa.gov/biosolids/basic-information-about-biosolids#uses) accessed 12 March 2025.

833 Vidmar, V., Milačič Ščančar, R., Ščančar, J., 2017. Sizing and simultaneous quantification of  
834 nanoscale titanium dioxide and a dissolved titanium form by single particle inductively coupled plasma  
835 mass spectrometry. *Microchem. J.* 132, 391–400. <https://doi.org/10.1016/j.microc.2017.02.030>

836 Vidmar, J., Zuliani, T., Milačič, R., Ščančar, J., 2022. Following the Occurrence and Origin of Titanium  
837 Dioxide Nanoparticles in the Sava River by Single Particle ICP-MS. *Water.* 14, 6, 959.  
838 <https://doi.org/10.3390/w14060959>

839 Wang, J., Zhao, X., Wu, F., Tang, Z., Zhao, T., Niu, L., Fang, M., Wang, H., Wang, F., 2021. Impact of  
 840 montmorillonite clay on the homo- and heteroaggregation of titanium dioxide nanoparticles (nTiO<sub>2</sub>) in  
 841 synthetic and natural waters. *Sci. Total. Environ.* 784, 147019.  
 842 <https://doi.org/10.1016/j.scitotenv.2021.147019>

843 Westerhoff, P., Lee, S., Yang, Y., Gordon, G.W., Hristovski, K., Halden, R.U., Herckes, P., 2015.  
 844 Characterization, Recovery Opportunities, and Valuation of Metals in Municipal Sludges from U.S.  
 845 Wastewater Treatment Plants Nationwide. *Environ. Sci. Technol.* 49, 16, 9479–88.  
 846 <https://doi.org/10.1021/es505329g>

847 Westerhoff, P.K., Song, G., Hristovski, K., and Kiser, A., 2011. Occurrence and removal of titanium at  
 848 full scale wastewater treatment plants: Implications for TiO<sub>2</sub> nanomaterials. *J. Environ. Monit.* 13,  
 849 1195–1203. <https://doi.org/10.1039/C1EM10017C>

850 Wu, F., Harper, B.J., Harper, S.J., 2019. Comparative dissolution, uptake, and toxicity of zinc oxide  
 851 particles in individual aquatic species and mixed populations. *Environ. Toxicol. Chem.* 38, 3, 591–602.  
 852 <https://doi.org/10.1002/etc.4349>

853 Yang, Y., Chen, B., Hower, J., Schindler, M., Winkler, C., Brandt, J., Di Giulio, R., Ge, J., Liu, M., Fu,  
 854 Y., Zhang, L., Chen, Y., Priya, S., Hochella, M.F. Jr., 2017. Discovery and ramifications of incidental  
 855 Magnéli phase generation and release from industrial coal-burning. *Nat. Commun.* 8, 194.  
 856 <https://doi.org/10.1038/s41467-017-00276-2>

857 Zeng, G., Wan, J., Huang, D., Hu, L., Huang, C., Cheng, M., Xue, W., Gong, X., Wang, R., Jiang, D.,  
 858 2017. Precipitation, adsorption and rhizosphere effect: the mechanisms for phosphate-induced Pb  
 859 immobilization in soils—a review. *J. Hazard. Mater.* 339, 354–367.  
 860 <https://doi.org/10.1016/j.jhazmat.2017.05.038>

861 Figure 1. Bar graphs show the percentage recovery of six MNPs (Au, Ag, CeO<sub>2</sub>, CuO, PbO and TiO<sub>2</sub>)  
862 in different extraction solutions: Mili-Q water; 2.5; 10; and 20 mM TSPP. The error bars represent the  
863 standard deviation of at least two measurements.

864 Figure 2. Bar graphs show the percentage recovery of six metallic NPs (Au, Ag, CeO<sub>2</sub>, CuO, PbO and  
865 TiO<sub>2</sub>) using 2.5 mM TSPP as the optimal extraction solution. The error bars represent the standard  
866 deviation of five measurements.

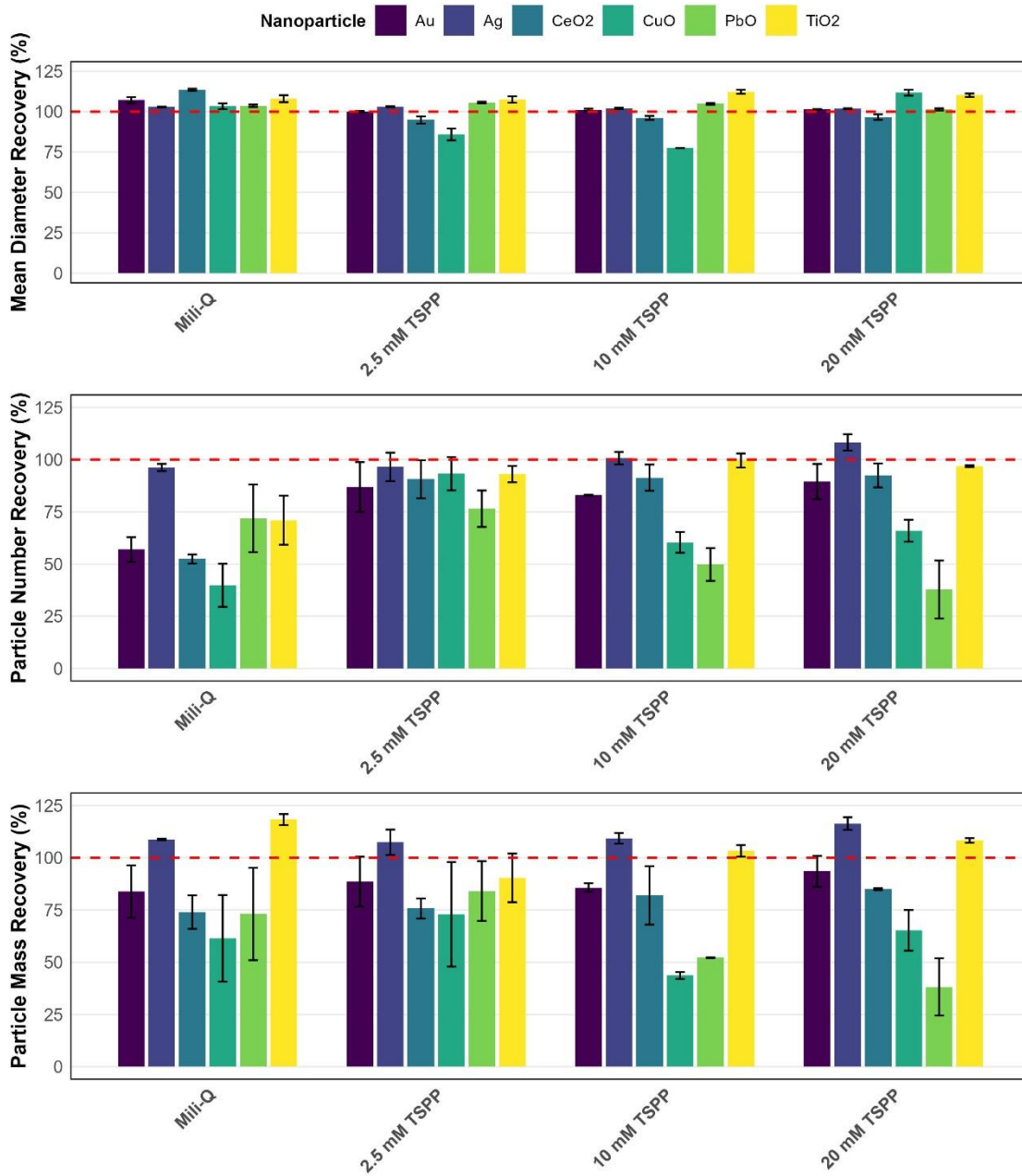
867 Figure 3. Particle number concentrations of Ag-, Au-, CeO<sub>2</sub>-, CuO-, PbO-, TiO<sub>2</sub>-, and ZnO-based NPs  
868 in sewage sludge samples in Croatia.

869 Figure 4. The size distribution of Ag-, Au-, CeO<sub>2</sub>-, CuO-, PbO-, TiO<sub>2</sub>-, and ZnO-based NPs in the  
870 samples containing their highest particle concentrations. The highest concentrations were found for  
871 Au, Ag, TiO<sub>2</sub>, and ZnO in VG sample; CeO<sub>2</sub> in SIS sample; and CuO and PbO in ZG sample.

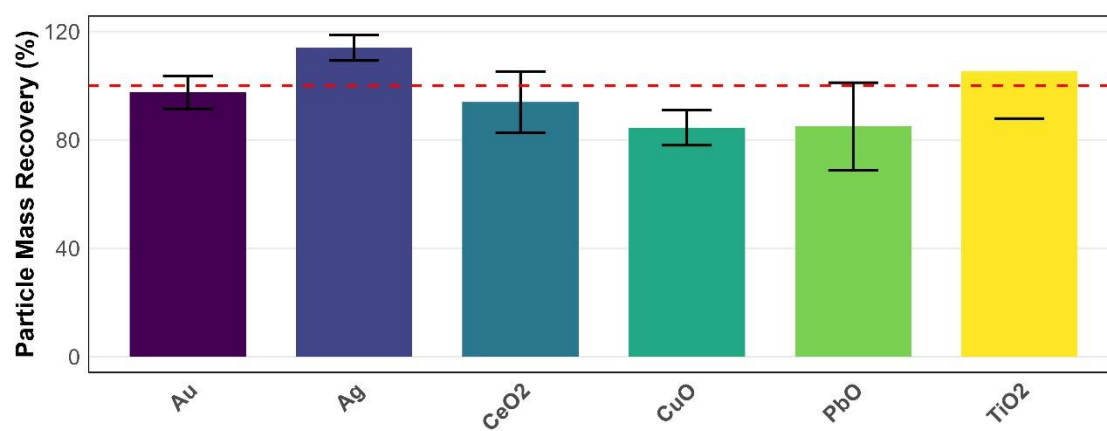
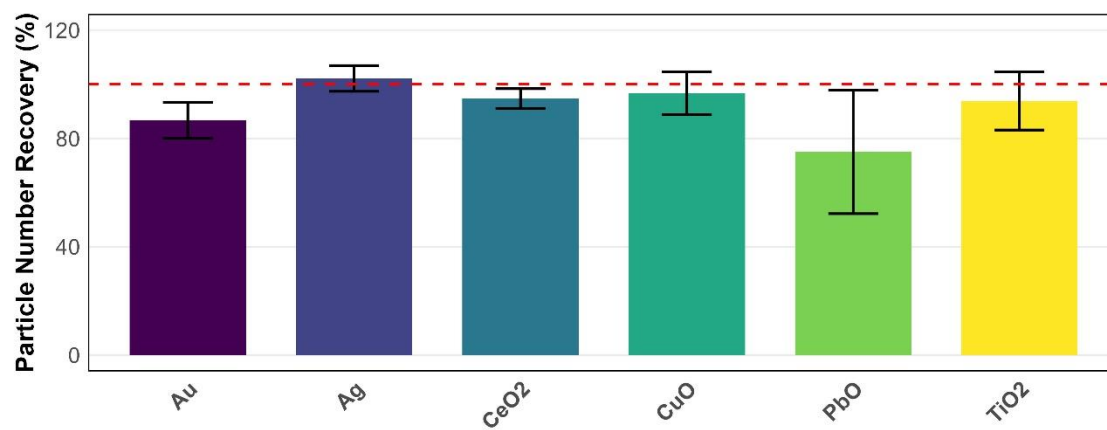
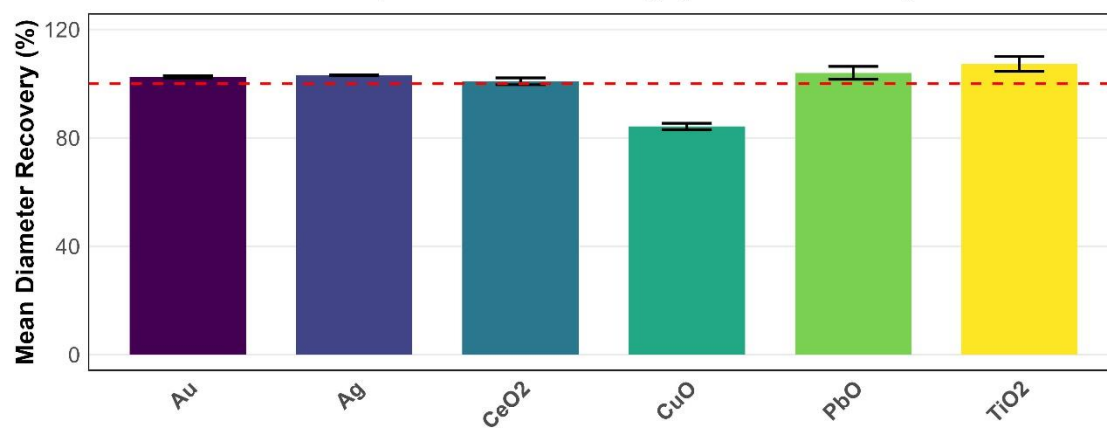
872 Figure 5. Results of risk quotient (RQ) calculation. Results can be interpreted as follow: (I)  $RQ < 0.01$ :  
873 Negligible risk; (II)  $0.01 \leq RQ < 0.1$ : Low risk; (III)  $0.1 \leq RQ < 1$ : Moderate risk; (IV)  $RQ \geq 1$ : High risk.

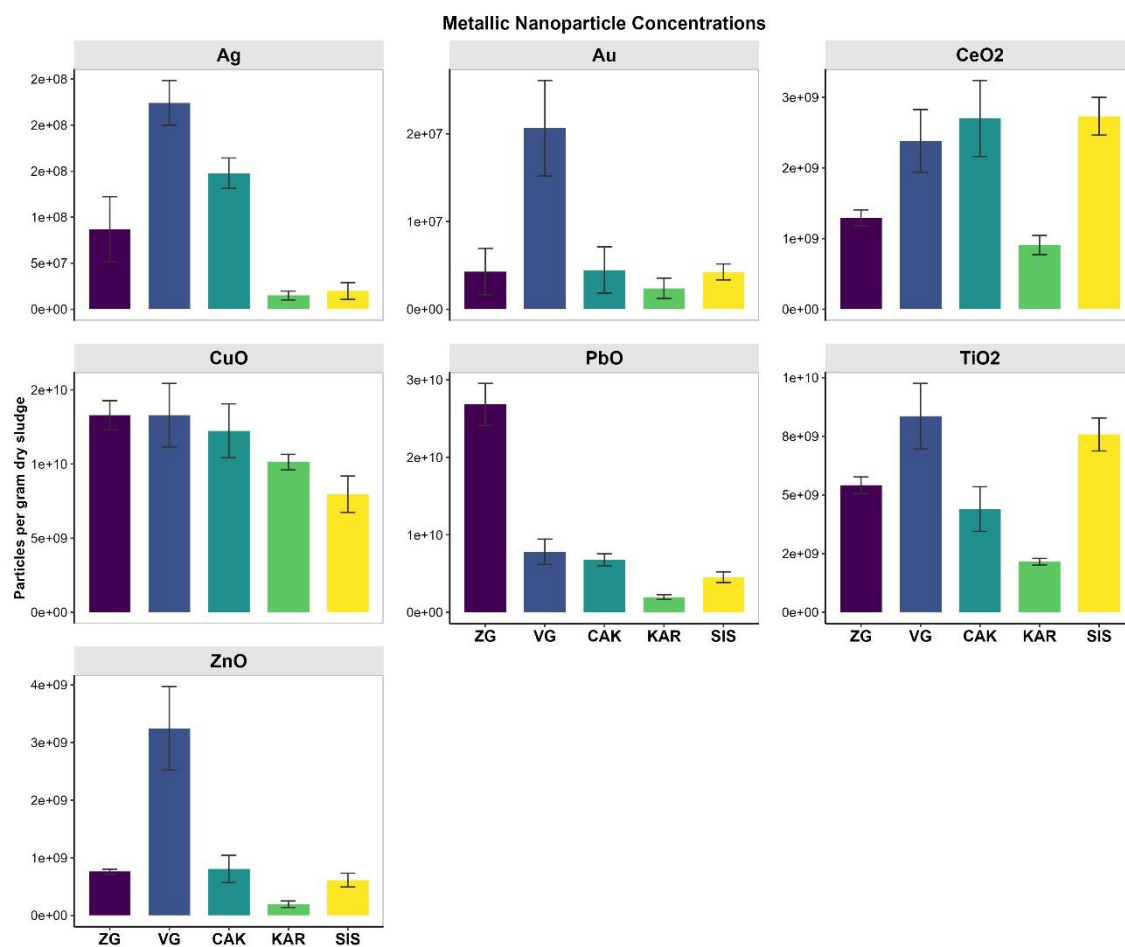


## Nanoparticle Recovery

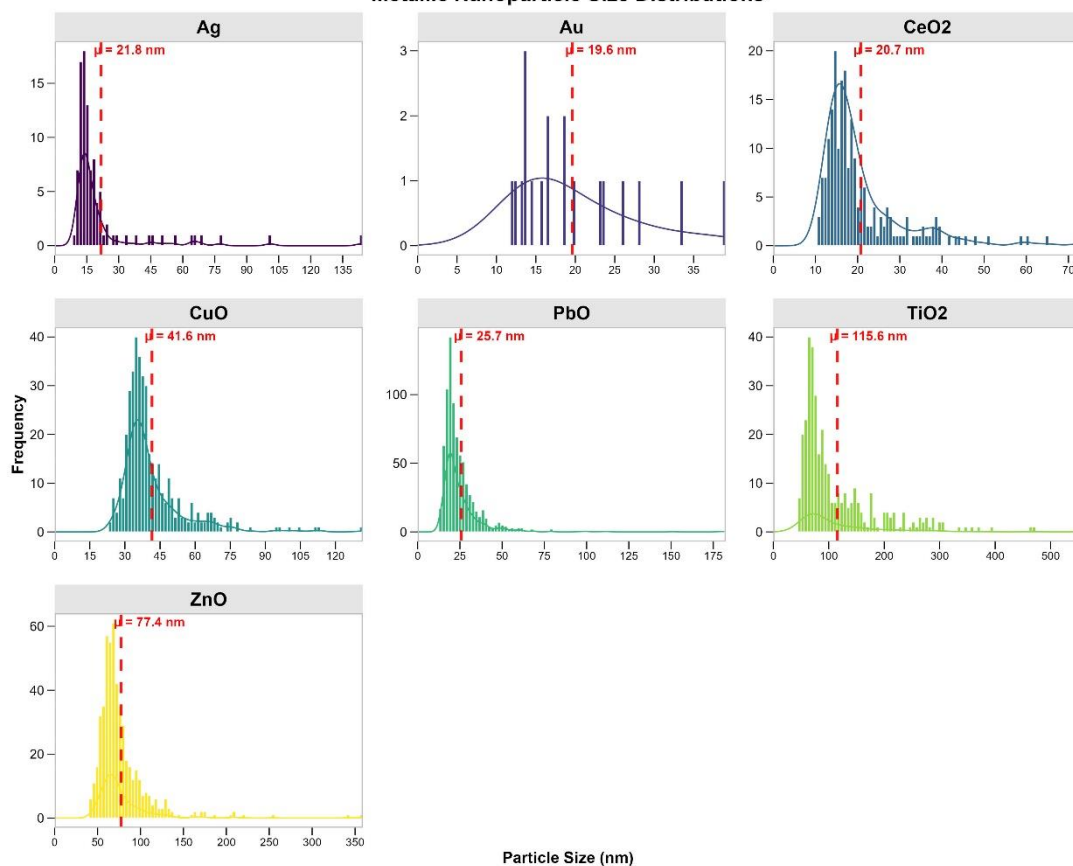


### Nanoparticle Recovery (2.5 mM TSPP)





# Metallic Nanoparticle Size Distributions



4

